# Hartland Landfill <br> Groundwater, Surface Water and Leachate Monitoring Program Annual Report 

(April 2022 to March 2023)

Capital Regional District

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| Draft | September 27, 2023 | For Review | RM | Ryan Mills | Project Manager |
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## Executive Summary

Based on our review of historical data and interpretation of the groundwater, surface water, and leachate quality data collected at Hartland between April 2022 and March 2023, the Annual Monitoring Program provided an effective assessment of landfill performance and compliance with respect to groundwater, surface water, and leachate flow and quality. Based on historical data and review of the 2022/23 data, AECOM made the following interpretations:

## Leachate Flow

Leachate elevation data collected in 2022/23 indicate that leachate mounding persisted in the Phase 1 landfill, and leachate elevations in Phase 1 were generally stable and exhibited minor seasonal variations. The leachate mound in the upper portion of the refuse is interpreted as being 'perched' above the regional bedrock groundwater flow system, with relatively high water levels and strong downward hydraulic gradients.

Like in 2021/22, leachate elevations in the Phase 2 Basin exceeded the elevation of the Lower Leachate Lagoon in November and remained above the elevation of the lagoon for the rest of the monitoring period. Historically, the leachate elevation in Phase 2 was approximately 1 to 2 m lower than the elevation of the Lower Leachate Lagoon. CRD confirmed that this trend was likely due to calibration/ instrument drift, and were not indicative of a change in operations. . The leachate elevation in Phase 2 was well-below the groundwater elevations observed at locations GW-36-1-1 and GW-37-1-1, indicating that the hydraulic trap was preserved throughout the monitoring year.

In 2022/23, leachate discharge volumes were lower than those observed in 2021/22. The total volume of leachate discharged in $2022 / 23$ was $520,740 \mathrm{~m}^{3}$, approximately $7.5 \%$ less than in $2021 / 22$. The contrast in leachate discharge volumes likely reflects a lower volume of precipitation in 2022/23 compared to 2021/22. It is also possible the lower volume of leachate may be due to biofouling of the North Purge Wells and a resultant decrease in the volume of leachate extracted from these wells over the past year.

In 2022/23, a total of $30,580 \mathrm{~m}^{3}$ of leachate was collected from the South Purge Wells, approximately $13.3 \%$ less than in the previous monitoring year. A total of $14,277 \mathrm{~m}^{3}$ of leachate was collected from the North Purge Wells, approximately $21.3 \%$ less than in the previous monitoring year. Groundwater levels measured in and around the purge well systems were consistent with historical values. The leachate discharge volumes and consistent groundwater elevations suggest that both purge well systems functioned effectively in 2022/23.

## Groundwater Flow

In 2022/23, groundwater flow patterns observed at Hartland were consistent with historical interpretations, with some variations in the North Ridge area. Regional groundwater flows from Mount Work northeast to the north-south trending valley that underlies the northern portions of the Phase 1 and Phase 2 landfills. Most of the northward groundwater flow in the bedrock below the landfill is captured by the Toutle Valley Underdrain, Phase 2 Heal basin leachate collection system, springs discharging to the lower lagoon, and the north and south purge well systems. Groundwater monitors east of Phase 1 confirm flow from east to west toward the landfill, preventing off-site migration to the east. Groundwater elevations in the North Ridge area still exhibited strong seasonal fluctuations, and the Highland Fault continued to function as a barrier to groundwater, impeding groundwater flow to the east.

Around the North Ridge and Hartland North Pad, located northwest of Phase 2, groundwater flows radially from a topographic high situated north of Phase 2. Throughout 2022/23, continued blasting operations along the North Ridge resulted in reductions in both the topography and the groundwater potentiometric surface. Although groundwater elevations in the North Ridge area continued to exhibit seasonal fluctuations, the intensity of the fluctuations was less pronounced.

## Groundwater Quality

The groundwater quality results from 2022/23 indicate that leachate impacted groundwater was contained within the landfill property. At the north end of the landfill, leachate impacted groundwater extended just north of the unlined Lower Leachate Lagoon and through the middle of the lined Upper Leachate Lagoon but did not extend off-site. Leachate impacts were observed in well GW-106-1-1, but were limited to an area less than 20 meters northwest of the Phase 2 Basin. South of the landfill, leachate impacted groundwater did not extend off-site. Leachate related exceedances were confined to the landfill footprint on the east side of Phase 1 and are inferred to extend to the western extent of the waste footprint within the Phase 2
landfill. These results indicate that the leachate collection system continued to function as intended, minimizing surface water and groundwater quality impacts.

In 2022/23, Boundary Compliance wells and off-site monitoring wells met CSR AW and DW standards, except for an anomalous copper concentration exceedance at GW-21-1-1 in May 2022. However, this recorded exceedance may not accurately represent true conditions as indicated by the high RPD discrepancy between parent and duplicate samples. Dissolved copper in the parent sample was not detected but highly elevated in the duplicate sample. Similar to previous years, most exceedances were present in groundwater wells near leachate purge wells and known leachate sources. However, nitrate concentrations in several groundwater wells located downgradient of aggregate stockpiles exceeded applicable CSR DW standards on one or more sampling event.

In 2022/23, groundwater in many areas of the landfill exhibited elevated conductivity, nitrate, and sulphate concentrations, reflecting the impacts of aggregate production, transport, stockpiling and use for construction at Hartland. These elevated aggregate runoff parameters were observed around the Northwest Stockpile, North of phases 1 and 2, around the Northeast Stockpile, south of Phase 1, and throughout the surface water system. AECOM is currently updating the Groundwater, Surface Water and Leachate Monitoring Plan to capture the cumulative impact of various activities, including aggregate production, stockpiling, leachate discharge and on-going construction. The following paragraphs summarize groundwater quality observations from different areas of the landfill:

## North of the Landfill

- In 2022/23, annual average conductivity values in the North Purge Wells were generally higher than those in 2021/22, reflecting more concentrated leachate. Changes in leachate quality in the North Purge Wells may reflect more concentrated leachate due to lower precipitation in 2022/23 than 2021/22, or mixing of leachate from Phase 1 and the Lower Leachate Lagoon.
- Operation of the Phase 1 North Purge Well System continued to mitigate leachate impacts north of the landfill, as indicated by long-term stable or decreasing concentrations of leachate indicator parameters at locations GW-40-1-1, GW-20-1-1 and GW-21-1-1. However, nitrate concentrations at location 40 increased considerably from the previous monitoring year.
- Groundwater quality in proximity to the Phase 2 Basin confirms the hydraulic trap leachate collection system is effectively containing leachate north of Phase 2. Groundwater quality 100 m north of Phase 2 continued to show low concentrations of leachate indicator parameters, indicating groundwater quality is not affected by landfill leachate. The increase in nitrate and sulphate concentrations in groundwater is interpreted to be due to runoff from aggregate stockpiles and roads constructed with aggregate.
- Along the northern edge of the Phase 2 Basin, groundwater quality is primarily impacted by runoff from aggregate stockpiles. GW-105-1-1 and GW-106-1-1 showed evidence of leachate impacted groundwater, although this is thought to be limited to an area less than 20 meters northwest of the Phase 2 Basin.
- Groundwater quality at Boundary Compliance Station GW-31-1-1 met all applicable CSR standards in 2022/23. Elevated sulphate and nitrate concentrations observed at this location are attributed to aggregate runoff.


## Hartland North Pad

- Groundwater quality at the Hartland North Pad was slightly deteriorated, with elevated conductivity, nitrate, and sulphate concentrations observed in some wells (e.g., wells GW-44-1-1, GW-62-1-1, GW-77-1-1, GW-78-1-1, GW-87-1-1, GW-88-1-1). The concurrent increase in conductivity, sulphate, and nitrate concentrations suggests widespread impacts of aggregate runoff on shallow groundwater quality.
- Groundwater quality in GW-91-1-1 and GW-92-1-1 was generally consistent with background conditions, except for elevated conductivity and sulphate concentrations. Nitrate concentrations remained low at both sites. The elevated sulphate may be associated with natural sulphide oxidation. Groundwater at location GW-94-1-1 was clearly impacted by aggregate runoff, indicating it cannot be considered a background station.


## South of the Landfill

- Groundwater quality south of the landfill met all applicable CSR standards. Although ammonia concentrations in some wells south of the landfill were slightly elevated, they were within historical ranges and well below the applicable CSR standards.
- Groundwater quality at several locations (e.g., GW-85-1-1, GW-60-1-1, and GW-71-1-1) showed no evidence of landfill leachate impacts. However, elevated nitrate and moderate sulphate concentrations reflect impacts from aggregate stockpiling and use, which may be related to runoff from the paved area and wind-blown or transported aggregate dust. Since 2020, chloride concentrations have been occasionally elevated at some monitoring stations (e.g., GW-85-1-1 and GW-60-1-1), but the elevated chloride concentrations have not correlated with elevated ammonia concentrations. High CI/Na molar ratios (>1) suggest there is an additional source of chloride other than road salt, but the source of chloride is currently unknown.


## East of the Landfill

- Water quality along the east boundary of the Phase 1 landfill was consistent with previous years, and concentrations of all parameters were below applicable CSR standards. However, elevated sulphate and nitrate concentrations were observed at Site 16 reflecting the influence of aggregate runoff from the Northeast Stockpile.


## Domestic Well Water Quality

As part of the CRD's groundwater quality monitoring program, sixteen (16) domestic wells within a 2 km radius of the landfill were sampled in 2022/23. The groundwater quality data was consistent with historic results, meeting all applicable federal and provincial drinking water quality guidelines (CDWQ and SDWQG). This indicates that offsite domestic water wells continue to remain unimpacted by landfill leachate.

## Surface Water Quality

Surface water quality data collected in 2022/23 confirmed that nearby surface water bodies, including Tod Creek, Durrance Lake and Durrance Creek and Killarney Lake continue to be unimpacted by landfill leachate. However, surface water quality monitoring stations at the landfill continued to show signs of water quality degradation, especially in the area northwest of Phase 2.

In 2021/22, dissolved copper concentrations exceeded BCWQG-STA values at 8 stations. In 2022/23, dissolved copper concentrations exceeded the BCWQG-STA once in January at SW-S-12.

## North of the Landfill and Downstream of the North Pad

- Surface water at boundary compliance location Sw-N-05 continued to exhibit elevated nutrient concentrations in 2022/23, resulting in non-compliant conditions. Nitrate concentrations at Sw-N-05 exceeded BCWQG-STA during November 2022 sampling event, suggesting an impact on surface water from quarrying, aggregate production, stockpiling and use. The absence of concurrent elevated ammonia and chloride concentrations indicates surface water was not impacted by leachate. The occasionally elevated ammonia concentrations may be associated with nitrate reduction via denitrification under reducing conditions.
- Surface water quality at boundary compliance station Sw-N-16 met BCWQG-STA, except for one total iron exceedance during the February 2023 sampling event. The iron exceedance observed at $\mathrm{Sw}-\mathrm{N}-16$ was likely due to the disturbance of sediment during sampling, as indicated by the reported higher that historic TSS concentration. Surface water quality at Sw-N-16 was not impacted by leachate, but continued to exhibit minor influence from nearby construction activities involving blasting, aggregate production, transport and placement, and excavation of organic soils. Similar aggregate impacts were observed downstream at stations Sw-N-17 and Sw-N-45.
- Surface water at Sw-N-18 reflected dilute landfill leachate impacts and may also have been impacted by aggregate runoff. In November 2022, the plug and diversion measures at surface water station SW-N-18 were removed, and a diversion pipe with valve was constructed in late 2022/early 2023 with the option to direct surface water either to the NWSP or the upper leachate lagoon dependant on water quality.
- In the Hartland North Pad area, surface water quality at boundary compliance stations (Sw-N-41s1 and Sw-N-42s1) met BCWQG-STA in 2022/232, except for TSS. Leachate indicator parameters remained low in 2022/23, indicating
surface water was not impacted by landfill leachate. The slightly elevated sulphate, nitrate and conductivity concentrations indicate continued minor impacts from aggregate production, stockpiling and use.
- Historically, surface water stations Sw-N-14 and Sw-N-CS2 were used to monitor background conditions north of the landfill, but elevated conductivity and nitrate concentrations were observed in 2022/23, and sulphate concentrations were elevated in Sw-N-14. It indicated that surface water quality at these stations has been impacted by aggregate runoff and they are no longer suitable for use as background monitoring locations.
- Sw-S-52 consistently showed no signs of impacts related to the landfill, confirming that the water quality remains representative of the background conditions south of the landfill.
- Surface water quality downgradient of the North Pad (Sw-N-41s3) exhibited slightly elevated nitrate concentrations ( 1.28 to $1.57 \mathrm{mg} / \mathrm{L}$ ) corresponding to low to moderate sulphate concentrations ( 17.0 to $20.0 \mathrm{mg} / \mathrm{L}$ ). In the absence of elevated sulphate concentrations, it is difficult to interpret whether the elevated nitrate concentrations reflect a background process, , or dilute runoff from aggregate stockpiles originating at Hartland landfill. Historically, nitrate concentrations have generally been below $0.2 \mathrm{mg} / \mathrm{L}$, but they occasionally elevated to a peak level of $10 \mathrm{mg} / \mathrm{L}$ in 2007. Further downstream of $\operatorname{Sw-N-41s3,~water~quality~at~Sw-N-41s4~was~consistent~with~background~conditions,~and~}$ showed no signs of aggregate or leachate impacts.
- Further downstream to the north of the landfill, at the confluence of Durrance Creek and Tod Creek (Sw-N-64 and Sw-$\mathrm{N}-65$ ), surface water quality showed no impacts from landfill leachate or aggregate runoff. The slightly elevated nitrate ( $<1.5 \mathrm{mg} / \mathrm{L}$ ) and sulphate ( $<20 \mathrm{mg} / \mathrm{L}$ ) concentrations at $\mathrm{Sw}-\mathrm{N}-63$ may have originated from the on-site aggregate runoff, or may be associated with nutrients derived from the surrounding agricultural lands.


## South of the Landfill

- Water quality at the boundary compliance location (Sw-S-04) met the BCWQG-STA values for all analytes in all samples collected during 2022/23, except for one total iron concentration observed in May 2022. Surface water quality along the south boundary was not impacted by leachate but may be influenced by dilute aggregate runoff.
- Water quality at Sw-S-52 (not a compliance location) was representative of background water quality. In 2022/23, concentrations of all parameters were below the BCWQG-STA. Concentrations of leachate indicator parameters were consistent with previously reported values.
- Like in 2021/22, surface water quality south of the recycling area (Sw-S-03, Sw-S-12) exhibited several BCWQG-STA exceedances, including TSS, dissolved and total iron, copper, and chloride during one or more sampling dates. Elevated ammonia, nitrate, conductivity and sulphate concentrations at these stations may be related to aggregate dust from the south face of Phase 1 and runoff from paved areas surrounding the bin facility that experiences heavy traffic and several industrial activities.


## Leachate Quality

In 2022/23, the leachate quality observed in the Hartland Valve Chamber followed the requirements of the Waste Discharge Authorization, except for COD exceedances on multiple sampling dates. Based on discussions with the analytical laboratory, CRD confirmed that the noted COD exceedances were due to the use of compromised/expired preservatives that were provided to CRD by the laboratory, and the exceedances do not likely reflect in-situ leachate quality. Overall, average annual leachate concentrations in 2022/23 were comparable with those measured in 2021/22.

## Quality Assurance and Quality Control

Upon review of the quality assurance and quality control data collected in 2022/23, groundwater, surface water and leachate sampling and laboratory analysis have produced reliable results that are acceptable for the purposes of this monitoring report.

## Compliance with Operational Certificate and Waste Discharge Authorization

Groundwater quality, surface water quality, and leachate quality data were used to assess compliance with the Amended Operational Certificate and Waste Discharge Authorization and are discussed individually below.

## Groundwater

A total of 36 groundwater monitoring wells were identified as Boundary Compliance Monitoring Wells. Water quality data collected from these wells were compared to the CSR standards for the protection of freshwater aquatic life and drinking water
to assess compliance with the landfill Operating Certificate and protect both current and future uses of the groundwater resource.

With respect to groundwater quality, Hartland Landfill remained in compliance with the Operational Certificate in 2022/23 except for one (1) copper exceedance at Boundary Compliance Station GW-21-1-1. However, this recorded exceedance may not be representative of in-situ groundwater quality due to high RPD discrepancy. Dissolved copper was not detected in the parent sample but was highly elevated in the duplicate sample. Overall, the copper exceedance is unrelated to landfill activities, as indicated by low concentrations of parameters associated with aggregate runoff and leachate.

## Surface Water

A total of five (5) surface water monitoring stations have been identified as Boundary Compliance stations surrounding Hartland Landfill. These stations are concentrated along the southern and northern property boundaries, downgradient of areas that have the potential to be impacted by leachate or landfill runoff. Water quality data collected from the Boundary Compliance stations were compared to the BCWQG-STA criteria to assess compliance with the Landfill Operational Certificate.

Some water quality impacts observed at the Boundary Compliance stations were caused by sources other than landfill leachate or aggregate runoff, including turbid samples collected under low-flow conditions and ongoing construction activities. In 2022/23, surface water quality was slightly deteriorated, exhibiting widespread impacts related to aggregate production and stockpiling. Throughout the monitoring year, highly elevated conductivity, sulphate, nitrate and/or ammonia concentrations consistent with aggregate runoff were observed at Boundary Compliance stations Sw-N-05, and Sw-N-16. Nitrate concentrations exceeded the BCWQG-STA at Sw-N-05 during the May and November 2022 sampling events. Additionally, moderately elevated conductivity, sulphate, and nitrate concentrations at Sw-N-41s1 and Sw-N-42s1 indicate minor impacts from aggregate production, stockpiling and use. Ultimately, in 2022/23, surface water quality at Sw-N-05 was not compliant with the Landfill Operational Certificate. Table 9-1 summarizes BCWQG-STA exceedances observed at Hartland in 2022/23.

Table ES-1. Surface Water Quality Compliance at Property Boundary Stations

| Station | General <br> Parameters | Nutrients | Metals | Comments |
| :---: | :---: | :---: | :---: | :---: |
| North of the Landfill | Nitrate (1) | None | $\bullet$ <br> Sw-N-05NoneNitrate exceedances are associated with aggregate production and <br> stockpiling at Hartland. The nitrate originates from leaching of ANFO <br> residue left on the aggregate after blasting. |  |
| Sw-N-16 | None | None | Total Iron (1) | Exceedances are anticipated to be related to turbid flow conditions <br> following a prolonged dry period. Continued monitoring to assess these <br> anomalous results. |
| SW-S-04 | None | None | Total Iron (1) | Exceedances are anticipated to be related to turbid flow conditions <br> following a prolonged dry period. Continued monitoring to assess these <br> anomalous results. |

## Leachate

The Hartland Valve Chamber is the Compliance Monitoring Station for the Waste Discharge Authorization under the CRD Sewer Use Bylaw, Bylaw 2922. During the monitoring period, leachate discharges at the Hartland Valve Chamber were in compliance with the Waste Discharge Authorization requirements.

## Recommendations

Based on the findings of this report, our recommendations are summarized in Table ES-2:
Table ES-2. Summary of Recommendations

| Leachate Collection System |  | Status |
| :---: | :---: | :---: |
| 1 | Closely monitor water levels and leachate quality in the north purge wells to verify the effectiveness of the leachate collection system. Water levels in well 52-3-0, adjacent to 52-4-0-P7 have slowly increased since 2021/22 and may indicate diminished drawdown and leachate collection in this area. A step test should be conducted on each north purge well to measure the specific capacity which is an indicator of well performance. The measurements should be compared to historical assessments to determine the need for well rehabilitation. Options for maintaining lower leachate levels in P7 and P9 should be further investigated to continue improving groundwater quality west of the lower leachate lagoon. | New/Ongoing |
| 2 | Closely monitor water levels and leachate quality in the south purge wells to verify the effectiveness of the leachate collection system and identify opportunities for improvements. Several pump failures were reported for south purge wells P3 and P10. Increased water levels above operational targets were observed in P1, reaching a peak of 150.6 m . Pumping elevations in the south purge wells (P2, P3, P4 and P10) should be maintained at elevations below 140 m asl. Pumping elevations in P1 should be maintained near the bottom of the screened interval around 146 m asl. | New/Ongoing |
| 3 | Periodically validate the pumping levels and the extent of the drawdown cones surrounding the north and south purge well systems (next assessment in 2024) to confirm the proper functioning of the purge wells. All procedures should follow the Standard Operating Procedure (SOP) - North Purge Well Drawdown Cone Verification (AECOM 2016), with interpretation of results by a qualified professional. Water levels in purge wells and pump maintenance should be conducted regularly to confirm the efficiency of the purge wells. | Ongoing |
| 4 | Conduct a detailed assessment of the effectiveness of the hydraulic trap and leachate collection systems including the north purge wells and south purge wells based on the design of the Phase $4 / 5 / 6$ quarry and liner system. This is required to confirm the landfill will perform as intended as the landfill extends further north and west, and as additional lifts are constructed. Recent groundwater and surface water characterization between the Phase 2 Basin and the Northwest Sedimentation Pond suggests additional leachate containment or groundwater management measures need to be implemented to mitigate the potential for off-site leachate migration and non-compliant water quality at $\mathrm{Sw}-\mathrm{N}-05$. | Ongoing |
| Runoff and Infiltration Associated with Aggregate Stockpiles |  |  |
| 5 | Update the aggregate impact indicator parameters and thresholds based on recent geochemical testing results for aggregate samples and recommendations of the Aggregate Management Plan that is presently being developed. | New |
| 6 | Minimize the spatial extent and volume of aggregate stockpiles outside of the leachate collection system. Where this is not feasible, stockpiles should be covered with low permeability temporary tarps as soon as practical to minimize sulphate, ammonia, nitrate and TSS impacts on downgradient groundwater and surface water quality. Direct runoff from aggregate stockpiles should be diverted away from natural water courses as it is known to exceed BCWQ guidelines for sulphate and some nitrogenous compounds. This approach proved to be effective for mitigation of historical aggregate impacts at the Hartland North Pad. | Ongoing |
| Groundwater Monitoring Program |  |  |
| 7 | Advance a network of boreholes into the bedrock slope west of the Phase 2 landfill to characterize the geology, hydrogeology and groundwater quality. This will also allow for establishment of a long-term groundwater monitoring network west and upslope of the Phase 2 landfill to support continued evaluations of hydraulic trap performance and monitor groundwater quality. | New/Ongoing |
| 8 | Groundwater monitoring wells in proximity to the Phase 2 Basin and Northwest Sedimentation Pond should be closely monitored to confirm the hydraulic trap leachate collection system is effectively containing leachate north of Phase 2. Continued quarrying may result in greater connection between the groundwater flow system on the west side of the Highland Fault and the Phase 2 hydraulic trap, resulting in lowering of the surrounding groundwater levels and increased leachate generation. | New/Ongoing |
| 9 | Decommission any monitoring wells that will be affected by quarrying, aggregate stockpiling, landfill development in advance of any damage to satisfy the requirements of the British Columbia Groundwater Protection Regulation. Based on near term construction activities, it is anticipated that monitoring wells 27-1-1, $78-1-1$, and $78-2-1$ will be inevitably impacted. Monitoring wells 27-12 and $93-1-1$ should also be decommissioned. | New |
| 10 | Establish a new background groundwater monitoring well further upgradient of the Northwest Stockpile to replace 94-1-1. Water quality in well 94-1-1 is no longer considered representative of | New |


|  | background groundwater quality. It is possible that this monitoring well could be coordinated with investigation of the bedrock hydrogeology west of the Phase 2 landfill. |  |
| :---: | :---: | :---: |
| 11 | The existing Groundwater, Surface Water and Leachate Monitoring Plan should be updated to ensure that it remains effective in monitoring the impacts current and future of landfill operations, including aggregate production, stockpiling, leachate discharge and on-going construction. This work is currently underway. | Ongoing |
| 12 | Groundwater wells should be surveyed once every five years to verify well condition and ensure geodetic well elevations are accurate (i.e., next survey in 2025). | Ongoing |
| 13 | The elevation of the leachate mound in Phase 1 and 2 should be determined at least once every five years (i.e. next assessment in 2025). | Ongoing |
| 14 | Conduct a review of the landfill development plan and filling plan every two years to ensure the existing monitoring network and monitoring program remain sufficient and interpretation of the data benefits from a complete understanding of the landfill design and operations over the next five years. The next review should be conducted in 2024 following completion of the Phase 4/5/6 quarry and liner design. | Ongoing |
| 15 | As required by the Amended Operational Certificate, the results of the annual monitoring program should continue to be reviewed and interpreted by a Qualified Professional experienced in assessing the impacts of landfill leachate at large municipal landfills similar to Hartland. | Ongoing |
|  | Surface Water and Leachate Monitoring Program |  |
| 16 | Add sodium to the surface water analytical packages. Analyzing sodium alongside chloride can help determine if elevated chloride concentrations originate from road salt application. | New |
| 17 | Establish a new background surface water station upgradient of the Phase 2 landfill to replace background water quality monitoring locations $\mathrm{Sw}_{\mathrm{w}} \mathrm{N}-14$ and Sw-N-C52 which are no longer representative of background conditions. | New |
| 18 | Surface water quality at locations Sw-N41s4, Sw-N-63, Sw-N-64, and Sw-N-65 should be sampled on quarterly basis to delineate the impact of aggregate runoff and assess its effect on the receiving environment. | New |
| 19 | Improve surface water flow monitoring upstream of Sw-N-05 in Heal Creek to ensure it provides an accurate measurement of surface water discharge from the landfill. Accurate flow measurements are important for evaluating environmental impacts and ensuring adequate collection and conveyance capacity. | New |
| 20 | Improve surface water management north of the Phase 2 landfill to minimize impacts of aggregate runoff on groundwater and surface water that is not captured by the leachate collection system. This may require lining of the NWSP and installation of an underdrain to allow for management of groundwater separately from surface water in the area. Additional sediment control measures and efforts to reduce the quantity of blasting residuals contained in aggregate stockpiles may help reduce impacts on water quality as quarry development becomes increasingly close to the northern property boundary and the water quality boundary compliance monitoring stations. | New |
| 21 | Characterize the chemistry of residual wastewater solids and stabilized biosolids (solids and leachate) to allow for future evaluation of any impacts to leachate chemistry. This information may be available from pilot studies or operational monitoring programs. | Ongoing |
| 22 | Determine the source of chloride, ammonia, dissolved copper and nitrate observed in surface water south of the Phase 1 landfill. Additional waste has been placed on the western and southern portions of Phase 2 over recent years and occasional leachate seeps and runoff from the truck wash facility have been noted in the past. Changes in activities at the south end of the landfill and management of impacted surface runoff may play a role. A multilevel monitoring well cluster should be established west of the bin facility and well 85-1-1 to resolve whether the source of impacts to surface water are due to runoff or discharge of leachate impacted groundwater. | New/Ongoing |
| 23 | Resume leachate sampling from the Phase 2 Cleanout as soon as the sampling pump is replaced. This information will be important for tracking changes in leachate chemistry as Phase 2 Cell 4/5/6 are developed. | Ongoing |
| 24 | In addition to the Sewer Use Criteria, leachate quality results for trace organic compounds should be compared to CSR standards for the protection of drinking water and aquatic life to allow for screening of data to identify parameters in leachate that exceed CSR standards and guide any refinements to the monitoring program in future years. Additionally, an updated list of emerging contaminants will be integrated into the monitoring program for the 2023/2024. | Ongoing |
|  | Construction Management |  |
| 25 | Blasting and quarrying activities should continue be to be conducted under the direction of a qualified blasting professional to minimize the potential for blast-enhanced fracturing, with possible negative impacts on hydraulic properties. This has been demonstrated to have important implications on groundwater elevations west of the Highland Fault and the volume of seepage reporting to the Phase 2 Basin as the base of the quarry has been lowered In circumstances where blasting might induce substantial topographic alterations or changes to the elevation of the | New |


|  | base of the Phase 2 quarry, consultation with a hydrogeologist is recommended to evaluate <br> potential implications on the performance of the hydraulic trap and the leachate collection system. |  |  |  |
| :---: | :--- | :--- | :---: | :---: |
| $\mathbf{2 6}$ | The placement of aggregate, road salt, dust suppressant and herbicides should be carefully <br> considered and documented to help understand the causes of potential future concentrations of <br> conductivity, ammonia, chloride, nitrate, sulphate and select metals at groundwater and surface <br> water monitoring locations. | Ongoing |  |  |
| Quality Assurance and Quality Control |  |  |  |  |
|  | Quality assurance for laboratory analyses should continue to be evaluated quarterly, and any <br> discrepancies should be resolved with the laboratory and CRD sampling personnel within one <br> month of receiving the laboratory results. The appropriate notation should be added to the data <br> files to explain the reason for the low precision and the steps taken, if any, to improve the sampling <br> or laboratory procedures. | Ongoing |  |  |

## Table of Contents

1. Introduction ..... 17
2. Site Description ..... 19
2.1 Physiography ..... 19
2.2 Geology ..... 19
2.2.1 Surficial Geology. ..... 19
2.2.2 Bedrock Geology ..... 19
2.3 Climate ..... 20
2.4 Significant Site Activities in 2022/23 ..... 20
2.5 Applicable Regulatory Criteria ..... 21
2.5.1 Groundwater ..... 21
2.5.2 Surface Water ..... 22
2.5.3 Leachate ..... 22
3. Methods and Quality Assurance ..... 23
3.1 Field Techniques ..... 23
3.2 Quality Assurance ..... 24
3.2.1 Groundwater and Surface Water ..... 24
3.2.2 Leachate ..... 25
3.3 Statistical Assessment of Temporal Trends ..... 25
3.4 Summary ..... 41
4. Groundwater Flow. ..... 42
4.1 Data ..... 42
4.2 Regional Groundwater Flow in the Bedrock. ..... 42
4.3 Leachate Elevations in Phase 1 and Phase 2. ..... 43
4.3.1 Phase 1 ..... 43
4.3.2 Phase 2 Basin ..... 44
4.4 Groundwater Flow in the Bedrock Aquifer Near the Landfill. ..... 45
4.4.1 East of Phase 1 ..... 45
4.4.2 North of Phase 1 ..... 45
4.4.3 South of Phase 1 ..... 46
4.4.4 North of Phase 2 ..... 47
4.5 Summary ..... 48
4.5.1 Leachate Flow ..... 48
4.5.2 Groundwater Flow ..... 48
5. Groundwater Quality Monitoring Wells ..... 60
5.1 Compliance Groundwater Monitoring Locations ..... 60
5.2 Assessment of Groundwater Quality Impacts ..... 60
5.3 Electrical Conductivity ..... 61
5.4 Overview of Groundwater Quality Exceedances ..... 62
5.5 Monitoring Sites North of the Phase 1 Landfill ..... 62
5.5.1 Monitoring Site 58 ..... 62
5.5.2 Monitoring Sites 52 (P7), 80 (P8) and 81 (P9) ..... 63
5.5.3 Monitoring Site 40 ..... 72
5.5.4 Monitoring Sites 20 and 21 ..... 72
5.5.5 Monitoring Site 31 ..... 72
5.5.6 Monitoring Sites 29 and 30 ..... 73
5.5.7 Monitoring Sites 28 and 39 ..... 73
5.6 Monitors West and North of the Phase 2 Landfill ..... 75
5.6.1 Background Groundwater Quality ..... 75
5.6.2 Wells North of the Phase 2 Landfill ..... 75
5.6.3 Wells near Hartland North Pad (Residual Treatment Facility) ..... 80
5.7 Monitors South of the Phase 1 Landfill ..... 84
5.7.1 South Purge Wells (P1, P2, P3, P4 and P10) ..... 84
5.7.2 Monitoring Site 85 ..... 84
5.7.3 Monitoring Site 60 ..... 84
5.7.4 Monitoring Site 07 ..... 85
5.7.5 Monitoring Sites 71, 72, 73 ..... 85
5.7.6 Monitoring Site 04 ..... 88
5.7.7 Monitoring Site 19 ..... 88
5.8 Monitors East of the Phase 1 Landfill ..... 88
5.8.1 Monitoring Sites 17 and 18 ..... 88
5.8.2 Monitoring Site 16 ..... 89
5.9 Summary ..... 89
6. Groundwater Quality in Domestic Wells ..... 93
6.1 Monitoring Locations ..... 93
6.2 Domestic Well Quality ..... 93
7. Surface Water Quality near the Landfill ..... 95
7.1 Compliance Monitoring Locations ..... 95
7.1.1 Regulatory Comments ..... 95
7.3 Assessment of Surface Water Quality Impacts ..... 100
7.4 Data ..... 100
7.5 Overview of Surface Water Exceedances ..... 101
7.6 Surface Water Quality North of Phase 1 ..... 101
7.6.1 Monitoring Site Sw-N-16 ..... 103
7.6.2 Monitoring Site Sw-N-05 ..... 103
7.6.3 Monitoring Sites Sw-N-14 ..... 103
7.6.4 Monitoring Site Sw-N-19 ..... 104
7.6.5 Monitoring Sites Sw-N-45 and Sw-N-17 ..... 104
7.7 Surface Water Quality North of Phase 2 ..... 106
7.7.1 Surface Water Quality in the Hartland North Pad Area ..... 106
7.7.2 Surface Water Quality in the Phase 2 Area ..... 106
7.8 Surface Water Quality Further North of the Landfill ..... 107
7.9 Surface Water Quality South of the Landfill ..... 108
7.9.1 Upgradient Surface Water Quality ..... 108
7.9.2 Surface Water Quality Near and South of the Recycling Area ..... 108
7.10 Summary ..... 114
8. Leachate ..... 115
8.1 Compliance Monitoring Locations ..... 115
8.2 Data ..... 115
8.3 Leachate Generation and Discharge ..... 115
8.4 Leachate Quality ..... 116
8.4.1 Routine Monthly Leachate Analyses and Sewer Use Bylaw Comparison ..... 116
8.4.2 Quarterly Trace Organic Analysis at Hartland Valve Chamber. ..... 120
8.5 Summary ..... 120
9. Conclusions ..... 121
9.1 Leachate Flow ..... 121
9.2 Groundwater Flow ..... 121
9.3 Groundwater Quality ..... 122
9.4 Domestic Well Water Quality ..... 123
9.5 Surface Water Quality ..... 123
9.6 Leachate Quality ..... 124
9.7 Quality Assurance and Quality Control ..... 124
9.8 Compliance with Operating Certificate and Waste Discharge Authorization ..... 125
9.8.1 Groundwater ..... 125
9.8.2 Surface Water ..... 125
9.8.3 Leachate ..... 125
10. Recommendations ..... 126
11. Qualifications of the Authors ..... 128
12. References. ..... 129
Figures
Figure 1-1. Site Location Map ..... 18
Figure 4-1. Groundwater Elevations and Flow Directions in Bedrock ..... 50
Figure 4-2. Groundwater Flow in Cross Section A-A' ..... 51
Figure 4-3. Groundwater Flow in Cross Section B-B' ..... 52
Figure 4-4. Groundwater Elevations East of Phase 1 ..... 53
Figure 4-5. Leachate and Groundwater Elevations within Phase 1 ..... 54
Figure 4-6. Groundwater Elevations Surrounding the North Purge Wells ..... 55
Figure 4-7. Groundwater Elevations in South Purge Wells ..... 56
Figure 4-8. Water Elevations within the Leachate Conveyance System and Surrounding the Phase 2 Basin ..... 57
Figure 4-9. Groundwater Elevations in North Ridge Area ..... 58
Figure 4-10. Groundwater Elevations in North of Phase 2 Landfill ..... 59
Figure 5-1. Electrical Conductivity Plan ..... 68
Figure 5-2. Electrical Conductivity in Cross Section A-A' ..... 69
Figure 5-3. Electrical Conductivity in Cross Section B-B' ..... 70
Figure 5-4. Groundwater Quality North of Phase 1 ..... 71
Figure 5-5. Groundwater Quality North of Willis Point Road ..... 74
Figure 5-6. Groundwater Quality North of Phase 2 - Landfill Leachate Impacts ..... 77
Figure 5-7. Groundwater Quality North of Phase 2 - Aggregate Impacts (2021 to Present) ..... 83
Figure 5-8 Groundwater Quality North of Hartland North Pad ..... 86
Figure 5-9 Groundwater Quality South of Landfill ..... 87
Figure 5-10 Groundwater Quality Southeast of Landfill. ..... 91
Figure 5-11 Groundwater Quality East of Landfill ..... 92
Figure 6-1. Domestic Well Locations. ..... 94
Figure 7-1. Surface Water Bodies and Sampling Locations ..... 102
Figure 7-2. $\quad$ Surface Water Quality North of Phase 1 ..... 105
Figure 7-3 Aggregate Surface Water Quality Impacts North of Hartland ..... 110
Figure 7-4. Surface Water Quality North of Phase 2 ..... 111
Figure 7-5. Surface Water Quality Downstream of the Hartland North Pad ..... 112
Figure 7-6 Surface Water Quality South of Landfill ..... 113
Figure 8-1. Hartland Valve Chamber Leachate Chemistry (Conductivity, Ammonia and Chloride) ..... 118
Figure 8-2. Hartland Valve Chamber Leachate Chemistry (Sulphide, BOD and COD) ..... 119

## Tables

Table 3-1. Groundwater Quality QA/QC - Relative Percent Difference ..... 27
Table 3-2. Surface Water Quality QA/QC - Relative Percent Difference ..... 33
Table 3-3. Hartland Valve Chamber Leachate Chemistry QA/QC - Relative Percent Difference ..... 37
Table 5-1. Groundwater Quality Exceedances ..... 64
Table 7-1. Surface Water Quality Exceedances ..... 97
Table 9-1. Surface Water Quality Compliance at Property Boundary Stations ..... 125
Table 10-1. Summary of Recommendations ..... 126

## Appendices

## Appendix A. Monitoring Station and Groundwater Level Data

A1. Monitoring Well Co-ordinates
A2. Monitoring Well Details
A3. Groundwater Elevations
A4. Surface Water Station Details
Appendix B. Water Quality Data
B1. Groundwater Quality
B2. Domestic Well Quality
B3. Surface Water Quality
B4. Monthly Leachate Quality Data - Hartland Valve Chamber
B5. Quarterly Leachate Quality - Trace Organics
B6. Monthly Leachate Quality - Phase 2 Cleanout
B7. Monthly Leachate Quality - North Purge Wells
B8. Monthly Leachate Quality - Controlled Waste Drainage
B9. Monthly Leachate Quality - South Purge Wells
B10. Monthly Leachate Quality - West Face Drainage
B11. Monthly Leachate Quality - Cell 3 Pipe Outlet
B12. Monthly Leachate Quality - Emerging Contaminant
Appendix C. Climate Data
C1. Daily Rainfall Data - Hartland Landfill Weather Station
C2. Monthly Rainfall Data - Hartland Landfill Weather Station
Appendix D. Leachate Pipeline Flow Data
Appendix E. Hartland Landfill Site Plan and Sampling Locations
E1. Hartland Landfill Site Plan
E2. Groundwater Level Monitoring Locations
E3. Groundwater Quality Monitoring Locations
E4. Surface Water Quality Monitoring Locations
E5. Leachate Quality Monitoring Locations
Appendix F. Hartland Landfill Leachate Pipeline Plan
Appendix G. Results of Statistical Analysis
G1. Groundwater
G2. Surface Water
G3. Leachate

## 1. Introduction

Hartland Landfill is located at the end of Hartland Avenue approximately 14 kilometres (km) north of Victoria (Figure 1-1). Filling with waste commenced at the site in the 1950s. The site was owned and operated by a private company until 1975 when the property was purchased by the Capital Regional District (CRD). The landfill is currently owned and operated by the CRD and is the primary solid waste disposal site for the 13 member municipalities and three electoral areas of the Capital Region.

The CRD initiated a surface water and groundwater monitoring program at the landfill in 1983. Since 1988, annual monitoring reports have been prepared and issued by Gartner Lee and AECOM. The present Hartland Monitoring Program is part of an Amended Operational Certificate (\#12659) that is required and approved by the BC Ministry of Environment and Climate Change Strategy (ENV) and was last amended January 27, 2010.

The Hartland Landfill is divided into two distinct areas referred to as Phase 1 and Phase 2. Initially, waste was deposited in Phase 1, which reached capacity in 1996 and was capped in 1997. Phase 2 actively receives waste and is engineered and operated as a "hydraulic trap" landfill with the maintenance of upward groundwater gradients below the base of the Phase 2 landfill providing leachate containment. Filling of Phase 2 Cell 1 was completed in 2004, and Cell 1 was capped in 2011. During the summer of 2004, the west face of Phase 2 Cell 1 was capped with a geomembrane to reduce passive gas venting and provide an internal leachate collection system for future development of Phase 2 Cell 2. This area is referred to as the West Face closure. In 2016, Phase 2 Cell 3 was developed on the west side of Phase 2 along with new leachate collection infrastructure.

Leachate and surface runoff from the active landfill areas are directed to two leachate lagoons at the north end of the landfill. The leachate is transported by a pipeline through the northwest trunk sewer system and ultimately to the McLoughlin Point Wastewater Treatment Plant. Leachate discharge to sewer is authorized by CRD Regional Source Control Program Waste Discharge Authorization SC97.001, last amended on March 1, 2011, and is subject to the CRD Sewer Use Bylaw (Bylaw No. 2922).

This report presents our interpretation of water quantity and quality data collected between April 2022 and March 2023 to meet the following objectives:

1. Evaluate the effectiveness of the leachate containment and collection systems.
2. Assess the potential impact of landfill leachate and operational activities on groundwater and surface water quality.
3. Assess temporal trends in leachate, groundwater and surface water flow and quality.

Leachate, surface water and groundwater data collected beyond the monitoring timeframe was not considered in this report.


## 2. Site Description

### 2.1 Physiography

Hartland Landfill is located in the Tod Creek watershed, in the bedrock highlands of the Gowlland Range northwest of Victoria. The terrain is moderately rugged with relief of up to 446 metres $(\mathrm{m})$ in the area. Undeveloped CRD property (about 320 hectares (ha) in total) lies to the west and south of the landfill site. Mount Work Regional Park lies to the west and the Department of National Defence rifle range is situated north of the landfill. Durrance Lake Regional Park is located northwest of the landfill, and Killarney Lake Regional Park is located to the southeast. Private residential properties are located to the east and southeast of the landfill.

Durrance Lake and Killarney Lake are two major surface water bodies located within one kilometre of the landfill footprint. A lined perimeter ditch collects and diverts natural runoff from upslope areas around the landfill footprint. The water is discharged into ephemeral surface water features that report to Durrance Lake to the northwest and Killarney Lake to the southeast.

A Residuals Treatment Facility (RTF) was constructed at Hartland North to treat effluent from the McLoughlin Point Wastewater Treatment Plant. The RTF became operational in September 2020. Although the RTF is located within the Hartland Landfill Operating Area Boundary, it is managed by CRD as a separate site.

The Centrate Return Line (CRL) is used to return centrate from the RTF and leachate from the Hartland Landfill to the McLoughlin Point Wastewater Treatment Plant. The CRL was activated on December 30, 2020, and the formal switch from the leachate pipeline to the CRL occurred in March 2021.

A water main was installed along the East Perimeter Road with a water service installed between leachate lagoons in the northeast corner of the Hartland property. The new water main connects the RTF with Saanich's new water reservoir built on Hartland's eastern boundary. The new water service also supplies water to the newly built Pump Station No. 4, which is part of the Wastewater Treatment Project's Residual Solids Conveyance Line (RSCL). The RSCL conveys processed wastewater from the McLoughlin Point Wastewater Treatment Plant to the RTF. Construction of the RSCL was completed in the summer of 2020.

### 2.2 Geology

### 2.2.1 Surficial Geology

A thin veneer of glacial till composed of silty, gravelly sand, with interspersed cobbles and boulders mantles the bedrock in areas of gentle slopes and in valley bottoms. Locally surrounding the landfill, fluvial deposits consisting of well-sorted sands and gravels are observed in bedrock depressions and channels, including the channel that drained Heal Lake prior to development of the Phase 2 landfill. At the landfill, the bedrock is mantled by a thin, discontinuous veneer of glacial till (Vashon Drift) composed of silty, gravelly sand, with interspersed cobbles and boulders. Vashon Till is generally absent from the North Ridge and the Hartland North area.

In 2022, AECOM supervised a drilling program at Hartland. During the drill program, a gully infilled with unconsolidated material was identified north of the Phase 2 landfill, between Phase 2 and the Northwest Sedimentation Pond (NWSP) (AECOM 2023). The subsurface materials consisted of dense, sandy, silty diamicton and till, with interspersed cobbles.

### 2.2.2 Bedrock Geology

The bedrock geology in the area surrounding the landfill mainly comprises Wark Diorite Gneiss with Colquitz Gneiss outcropping in the northern and eastern margins of the landfill site with a thin overlying veneer of glacial till. The Wark Diorite Gneiss is dark green to black in colour. It is competent, except locally in shear zones, where it has been chloritized and weathered into soft, sand-size particles and clay. Discontinuities, including joints, shear zones and altered veins have been observed on the bedrock outcrops and noted in borehole logs. Geologic mapping of an exposed bedrock outcrop north of the Phase 2 landfill (AECOM 2018) characterized the undulating bedrock surface and confirmed the presence a complex network of lithologic contacts, faults and fractures observed in exposed bedrock excavations (sub-vertical quarry walls) including the Highland Fault and other faults that were previously inferred based on lineaments identified on aerial photographs (Thurber 1987). Evidence supporting the identification of a major oblique slip fault proximal to the trace of the Highland Fault was
identified in this study and further validated by a drill investigation in the fall of 2019 (AECOM 2019a). Permeability testing and water level measurements have determined that faults and shear zones locally behave as conduits and/or barriers to groundwater flow which imparts a degree of anisotropy in the hydrogeological properties of the bedrock (AECOM 2018).

### 2.3 Climate

The climate of this area is classified as "cool Mediterranean". Long-term (1981-2010) average climatic data is available for the Victoria International Airport Climatological Station (Victoria International Airport CS) located approximately 9 km from the landfill. Average annual temperature is $10.0^{\circ} \mathrm{C}$ and mean monthly values range from a low of $4.0^{\circ} \mathrm{C}$ in December to a high of $16.9^{\circ} \mathrm{C}$ in July. Mean annual precipitation is 882.9 mm .

In 1994, the CRD established a climate station (Victoria Hartland CS) at the landfill office. The original Hartland climate station was replaced in 2009/10 with new equipment at a location on top of Phase 1. The new weather station records temperature, precipitation, wind direction, wind speed, barometric pressure, and relative humidity directly to CRD's SCADA system.

Hartland daily precipitation measurements for 2022/23 are provided in Appendix C. The precipitation measured at Hartland Landfill from April 2022 to March 2023 was 837.4 mm, approximately 5\% lower than the 30-year average reported for Victoria International Airport CS ( $882.9 \mathrm{~mm} /$ year) , and 33\% lower than the total precipitation recorded in 2021/22 at Victoria Hartland CS (1279.5 mm). Precipitation during the 2022/23 wet season (i.e., October to February) was 528.0 mm , contributing over $63 \%$ of the total annual precipitation.

Climate change projections for the CRD were developed by the Pacific Climate Impacts Consortium to better understand the details of how climate may change in the Capital Region (CRD 2017). Overall, the results indicate that the region can expect warmer winter temperatures, fewer days below freezing, more extreme hot days in the summer, longer dry spells in summer months, more precipitation in fall, winter and spring, and more intense, extreme weather events. A water surplus occurs primarily in the cool, wet winter months (November, December, and January) with water deficit conditions occurring in the warm, dry summer months (July, August, and September).

### 2.4 Significant Site Activities in 2022/23

CRD staff maintain a log of activities at the landfill that are relevant to landfill operations, leachate containment, and compliance monitoring activities. Activities that are considered important from a compliance monitoring perspective are paraphrased below:

- Preparation of Leachate Line Decommissioning: The leachate pipeline was deactivated in March 2021 and plugged in April 2021 to prepare for decommissioning. The leachate line has been filled with super-chlorinated water since March 2021. In June 2021, the Regional Source Control program (RSCP) discharge permit (SC97.001) was amended to reflect construction of the CRL and decommissioning of the Markham Valve Chamber. In March 2023, to facilitate the Fortis construction near its alignment, the super-chlorinated water in the leachate line was permitted to drain into the Upper Lagoon.
- Deposition of Wastewater Treatment Residuals (WWTR) in Hartland Landfill: Between April 2022 and March 2023, approximately 11,596 tonnes of dried products ( $95 \%$ solid content) were received from the RTF. These biosolids are mixed volumetrically $1: 1$ prior to landfilling to prevent combustion. As of November 4,2022 , the practice of utilizing a blend of biosolids and overburden as a cover material for controlled waste and asbestos has been discontinued, and a landfilling approach has been adopted.
- Aggregate Storage Development: Aggregate production and blasting for air space, road building, and construction in the north area of Hartland continued in 2022/23. Several major storage areas were established to support aggregate storage and management, including the Northwest Stockpile, the Northeast Stockpile, the Triangle Stockpile, and the South Toe Stockpile. As of March 6, 2023, the current stockpile volume was estimated to be $691,252 \mathrm{~m}^{3}$, with a maximum capacity of $978,128 \mathrm{~m}^{3}$.
- Surface Water Quality at Compliance Location SW-N-05: Since November 2021, surface water at compliance monitoring station SW-N-05 has exhibited elevated and non-compliant nutrient concentrations. AECOM was retained by CRD to implement a multiphase assessment to determine the cause of elevated nutrient concentrations within and downstream of the Northwest Sedimentation Pond (NWSP) and Heal Creek. Between February 2022 and September 2022, three (3) drive-point piezometers, two (2) purge wells (P11 and P12), and 16 groundwater monitoring wells (GW-95-1-1 to GW-110-1-1) were installed north of the Phase 2 landfill. The new wells and drive points were monitored for
groundwater levels and groundwater quality, and then used to identify contaminant sources and migration pathways. AECOM determined that water quality north of Phase 2 was being impacted by dilute leachate and nitrate-rich contact water emanating from the Northwest Stockpile (AECOM 2023). The report identified several remedial strategies and options for managing aggregate stockpiles and leachate discharge, which are being implemented by CRD to manage aggregate stockpiling and leachate discharge, including development of an Aggregate Management Plan and updates to the existing groundwater and surface water monitoring plan. Additional design upgrades have also been incorporated into the Phase 4/5/6 landfill design.
- Establishment of New Surface Water Monitoring Stations: Between December 2022 and February 2023, a total of nine (9) surface water stations (SW-N-57, SW-N-58, SW-N-59, SW-N-60, SW-N-61, SW-N-62, SW-N-63, SW-N-64, and SW-N-65) were established north of the landfill. These stations aim to monitor the effects on surface water quality due to various landfill activities, such as road construction, blasting, and aggregate stockpiling. The data they generate supports evaluation of impacts on the downstream environment.
- Installation of Leachate Sump and French Drain System: Between June and October 2022, a French drain and a leachate sump were installed to mitigate the impact of leachate on surface water in the Phase 2 area. This system was strategically designed to reduce the perched, localized leachate mounding observed near the northern boundary of the area.
- Maintenance/Verification of North and South Purge Well Systems: Flow tests were regularly conducted at P1, P7 and P8 to monitor pump efficiency. Pumps installed in several leachate collection wells ( P 3 and P 10 ) were replaced in response to pump failure.


### 2.5 Applicable Regulatory Criteria

The Hartland Landfill is required to operate in accordance with the monitoring requirements outlined in the following:

- Amended Operational Certificate (\#12659) approved by ENV, last amended on January 27, 2010.
- Waste Discharge Authorization SC97.001 issued by the CRD Regional Source Control Program, last amended on March 1, 2011, and subject to the CRD Sewer Use Bylaw.
- Legally enforceable standards defined by the British Columbia Contaminated Sites Regulation (CSR).

The Stage 10 (Omnibus) and Stage 11 (Housekeeping) amendments to the CSR became effective on November 1, 2017. Stage 12 amendments to the CSR were issued January 24, 2019, and included clerical errors related to the Stage 10 and 11 amendments. The Stage 13 amendments came into force on February 1, 2021, and Stage 14 amendments came into effect on March 1, 2023. References to CSR standards in this report consider the latest amendments up to March 1, 2023.

As part of the update of Stage 10 (Omnibus) and Stage 11 (Housekeeping) amendments, a number of emerging contaminants were added to the water and soil schedules due to their persistence in the environment, toxicity, and relevance to contaminated sites in BC. In April 2017, AECOM conducted a focused review on the applicability of emerging contaminants at the Hartland Landfill based on Schedule 2 activities in the CSR. In 2018, AECOM conducted a detailed emerging contaminant review, to confirm regulatory compliance and to recommend an approach for future monitoring (AECOM 2018). It was recommended to continue sampling and analyzing of emerging contaminants of concern (CECs) at the Hartland Valve Chamber on a quarterly basis in conjunction with the trace organics sampling program, and the review should be conducted every five years to ensure effectiveness and relevance. The review of 2018 to 2022 data is currently on-going. In consideration of the five-year review process, quarterly sample collection in 2022/23 for CEC was momentarily paused, and only one sample was collected in May 2023 for selected CECs.

### 2.5.1 Groundwater

In February 2016, CSR Protocol 21 (P21), Water Use Determination, became legally enforceable. For the purposes of this report, Protocol 21 (Version 2.0, effective October 31, 2017) is considered when determining applicable water use standards at Hartland Landfill.

Groundwater quality data were compared to CSR Schedule 3.2 Generic Numerical Water Standards Column 3 Aquatic Life Use (AW) and Column 6 Drinking Water Use (DW), as required by P21.

Freshwater AW standards apply to groundwater because the landfill is "within 500 m of a surface water body containing aquatic life" (i.e., Heal Creek, Durrance Creek, Durrance Lake, Killarney Creek and Killarney Lake).

Future DW standards apply because the aquifer underlying the landfill has:

- A hydraulic conductivity greater than $10^{-6} \mathrm{~m} / \mathrm{s}$ and a yield greater than $1.3 \mathrm{~L} / \mathrm{min}$.
- Groundwater with natural total dissolved solids (TDS) concentrations of less than $4,000 \mathrm{mg} / \mathrm{L}$.
- An average confined aquifer saturated thickness greater than 1 m .
- An overlying silt or clay aquitard that is less than 5 m thick.

DW standards for iron and manganese do not apply to municipal landfills based on the Stage 8 Amendments to the CSR that came into effect on January 24, 2013, as presented on CSR Schedule 3.2.

The Stage 13 amendments to the BC CSR were effective as of February 1, 2021. As part of the Stage 13 amendments, Protocol 9, which includes the list of regional groundwater background concentrations, was updated. Regional background concentrations are provided in Table 1 of Protocol 9. Regional background concentrations for the Southern Vancouver Island have been considered when assessing groundwater quality data collected at the Hartland Landfill. In several instances, regional background concentrations are greater than the applicable CSR Drinking Water Use standard and/or the most stringent CSR Freshwater Aquatic Life Use standard. Groundwater that contains a substance concentration above the applicable CSR standard but below the regional background concentration for that substance is not considered contaminated under the CSR. All reference to CSR standards in this report consider Stage 14 amendments and included amendments up to March 1, 2023.

### 2.5.2 Surface Water

Surface water quality data are compared to the British Columbia Approved and Working Water Quality Guidelines (BCWQG) for the protection of AW and DW, because CSR Technical Guidance 15 (April 2013) indicates that groundwater in receiving environments (i.e., within 10 m of the high water mark of a surface water body) should be compared to the BCWQG. The approved and working BCWQGs were last updated in August 2023 and February 2021, respectively and the updated guidelines were considered when assessing data compliance in this report. Furthermore, there are two sets of BCWQG criteria for AW use:

- long-term chronic (LTC)
- $\quad$ Short-term acute (STA)

The purpose of the long-term chronic (LTC) guideline is to protect the most sensitive species and life stages against chronic exposure. To properly compare data against LTC guidelines, the LTC concentration is typically calculated by averaging the results of five or more samples collected within a 30-day period.

Short-term acute (STA) WQGs are established to protect the most sensitive species and life stages against short-term exposure (i.e., < 96 hours). BCWQGs for STAs are typically used to assess risks associated with short-term exposures.

The dissolved copper data evaluation has been updated to reflect the revised BCWQG AW criteria. The dissolved copper BCWQG varies with hardness, pH , dissolved organic carbon (DOC) and temperature, and is calculated using the Biotic Ligand Model (BLM). BLM is a series of linked equations that predict the toxicity of dissolved copper to aquatic life under specific conditions. In this report, only dissolved copper concentrations with paired DOC, pH and temperature data were compared to BCWQG AW criteria.

The approved BCWQG AW criteria for total aluminum, total arsenic and dissolved zinc were also updated in August 2023 version, and the changes have been considered in this report.

The BCWQGs apply to surface water on site and on adjacent sites under provincial jurisdiction. The Federal Water Quality Guidelines (CCME) apply on the property owned by the federal government located north of the site. Both the BCWQG and CCME criteria are guidelines and do not have the same regulatory force as the CSR Omnibus standards.

### 2.5.3 Leachate

Discharges from the leachate pipeline are subject to the CRD Regional Source Control Program (RSCP Waste Discharge Authorization (SC97.001) authorizing the discharge of non-domestic waste to the sanitary sewer in accordance with the CRD's

Sewer Use Bylaw 2922. In June 2021, the discharge authorization was amended to reflect operation of the CRL and decommissioning of the Markham Valve Chamber. Sampling is conducted on monthly basis.

The Hartland Landfill leachate compliance monitoring location is the Hartland Valve Chamber (flow detection chamber) located near the Lower Lagoon. Discharge limits are identified within the authorization.

## 3. Methods and Quality Assurance

### 3.1 Field Techniques

Sampling locations are shown on Figure 4-1. Boreholes and monitoring wells are identified using a standard system adopted by the CRD, consisting of three numbers (e.g., GW-02-02-01). The first number refers to the monitoring site, the second to the borehole at that site (there may be more than one) and the third number refers to the monitoring well in that borehole (there may be two or three at different depths in older installations). If the third number is a zero, it indicates an open borehole where no PVC monitoring well has been installed. Additionally, several leachate purge wells have been installed at the site. Purge wells are denoted with a "P" in front of the purge well number (e.g., P1).

Monitoring well construction details including location coordinates and elevations are summarized in Appendix A.1. Appendix A. 1 also lists the status of all the groundwater monitoring wells at the site together with comments describing any problems associated with each monitoring well. Monitoring wells are categorized as active (fully functioning) or inactive (non-functioning or destroyed). In 2022/23, there were 148 active groundwater monitoring wells at 94 locations in the vicinity of Hartland Landfill, of which ten (10) are active purge wells and 13 are landfill gas wells that were regularly used to measure leachate levels in Phase 1. South purge well P1 was replaced and recommissioned in November 2018.

The methods used to develop and sample each monitoring well are indicated in Appendix A.2. The Standard Operating Practice (SOP) for groundwater sampling is periodically updated. A variety of techniques are used depending on the depth of the monitoring well, the groundwater level in the monitoring well, and the permeability of the surrounding geologic formation. A number of dedicated submersible pumps have been installed by CRD in the deeper monitoring wells and open boreholes at the landfill to facilitate more efficient sampling and have resulted in improved data quality. Sampling events were scheduled to avoid heavy rainfall events to avoid sample dilution.

In 2022/23, the monitoring program consisted of the following:

- Groundwater level measurements four times per year.
- Continuous water level monitoring with pressure transducers at north end of the Phase 2 landfill.
- Continuous water level monitoring with pressure transducers at the north and south purge well systems.
- Continuous water level monitoring with pressure transducers east of Phase 1 landfill and Hartland North.
- Quarterly monitoring of wells near the property boundary and key locations to assess the effectiveness of leachate containment.
- Semi-annual monitoring stations with relatively stable long-term historical data.
- Annual sampling of 19 residential wells, including 14 wells within a 2 km radius of the landfill and five (5) domestic wells located north of the Hartland North Pad.
- Quarterly sampling of surface water stations at property boundary points and other selected monitoring locations.
- Quarterly testing of leachate discharge for trace organic compounds.
- Monthly testing of leachate for conventional parameters and metals at the point of discharge and select locations within the leachate collection system.
- One sampling event (May 2022) for select emerging contaminants at the Hartland Valve Chamber.

As in previous years, CRD staff carried out surface water, groundwater and leachate sampling and groundwater level measurements at the locations shown in tables in Appendix B and on figures in Appendix E. Further information on the monitoring program field procedures is contained in the CRD Monitoring Procedure Manual.

The CRD has adopted a continuous improvement and quality assurance program. The existing Leachate, Groundwater, and Surface Water Monitoring Plan for Hartland was established in 2016. Following a review, modifications were made to the groundwater, surface water, and leachate monitoring program, and these changes were implemented in the 2019/20 monitoring year. AECOM is in the process of updating the program to ensure that it remains effective in monitoring the impacts of various landfill operations, including aggregate production, stockpiling, leachate discharge and on-going construction.

### 3.2 Quality Assurance

In 2022/23, routine surface water, groundwater and domestic well water laboratory analyses were performed by Bureau Veritas Laboratory (BV) in Victoria and Burnaby, British Columbia. BV also analyzed leachate chemistry samples, which included analysis of trace organic compounds.

A quality assurance program is in place assess the validity of the chemical analysis results. This has involved the submission of randomly selected field replicate, trip blank and "reference" Victoria municipal water samples to the laboratory for analysis. Table 3-1 and Table 3-2 presents quality assurance of groundwater and surface water, respectively. There were 25 surface water and 42 groundwater samples submitted in duplicate between May 2022 and April 2023. Table 3-3 presents quality assurance results for the Hartland Valve Chamber. In this report, each set of replicates were taken from the same source and/or site, and under the same conditions. In all cases, the field replicates were submitted 'blind' to the laboratory. This resulted in duplicate sampling frequencies of $13 \%$ ( $42 / 330$ samples) for groundwater, $29 \%$ ( $25 / 87$ samples) for surface water, $42 \%$ ( $5 / 12$ samples) for the Hartland Valve Chamber compliance point, and $14 \%$ ( $9 / 64$ samples) for the overall leachate sampling program. Overall, duplicate samples were collected at a frequency of approximately $16 \%(81 / 493)$, which exceeded the targeted duplicate sampling rate of $10 \%$.

Submission of duplicate samples provides an estimate of total uncertainty associated with the data. Typically, one duplicate sample is collected for every ten samples (10\%) as part of quality control measures. Total uncertainty is the variability (precision plus bias) associated with the sample collection and sample analyses. In addition, the Limit of Quantification (LOQ) should be considered because the precision of analytical results just above the Method Detection Limit (MDL) are known to be poor. The LOQ is the range of concentrations between the MDL and five times the MDL for each parameter, as outlined in Part E of the 2013 British Columbia Field Sampling Manual.

The CRD has used several different statistical methods for checking the precision and accuracy of its monitoring program. The CRD uses the relative percent difference (RPD) method, as recommended by ENV, which uses duplicate analyses to determine precision of the analytical results. This method expresses percent of difference between two values as the ratio of their absolute difference to the average value of the sample and the duplicate, as shown in the equation below:
$R P D=\left[\left(\left|x_{1}-x_{2}\right|\right) /\left(\left(x_{1}+x_{2}\right) / 2\right)\right]$
Where:
$\mathrm{X}_{1}$ is the initial sample concentration ( $\mathrm{mg} / \mathrm{L}$ )
$\mathrm{x}_{2}$ is the duplicate sample concentration ( $\mathrm{mg} / \mathrm{L}$ )
Parameters with RPD values exceeding the RPD criteria should be interpreted with caution. Alarm limits were set as per the BC Ministry of Environment Environmental Laboratory Manual. Duplicate samples with RPD values within 30\% for general inorganics and metals and within $45 \%$ for organic compounds are considered to meet the Data Quality Objectives (DQOs).

### 3.2.1 Groundwater and Surface Water

Table 3-1 and Table 3-2 present calculated RPD values for replicate groundwater and surface water samples collected near the landfill. In both tables, RPD values were highlighted with an "a" if they were above $30 \%$ and it was noted with a "b" if one or more of the parameter concentrations were below the LOQ.

Table 3-1 indicates the following for groundwater samples collected at the landfill in 2022/23:

- A total of 42 duplicate samples were collected and analyzed for 57 parameters. Thirty-nine (39) samples had RPD values greater than $30 \%$ when all concentrations were above the LOQ.
- $\quad$ RPD values for pH (1 sample), ammonia (2 samples), dissolved aluminum ( 30 samples), antimony ( 2 samples), arsenic ( 1 sample), barium ( 2 samples), beryllium ( 1 sample), bismuth ( 1 sample), cadmium ( 11 samples), chloride ( 2 samples), copper (14 samples), iron (10 samples), lead (14 samples), magnesium (1 sample), manganese (2 sample),
molybdenum ( 1 sample), nickel (13 samples), phosphorus (6 samples), potassium (1 sample), selenium (1 sample), sodium ( 1 sample), sulphur ( 1 sample), thallium ( 5 samples), titanium ( 1 sample), uranium ( 4 samples), vanadium ( 5 samples), and zinc ( 23 samples), were greater than the $30 \%$ target when all concentrations were above the LOQ.
- Groundwater field replicates showed good precision for most parameters and the majority of RPD values for the 2022/23 monitoring year were within the acceptable range, except for dissolved aluminum, cadmium, copper, iron, lead, nickel, and zinc. These seven (7) dissolved metal concentrations had over $20 \%$ of the duplicate samples above the alarm limit of $30 \%$.
- Overall, groundwater quality data is acceptably precise for the purpose of this report. However, dissolved aluminum, cadmium, copper, iron, lead, nickel, and zinc concentrations should be interpreted with caution. Field blanks should be collected to further investigate/evaluate if the filters are the source of metal contamination. It is recommended that inline filters be flushed with sample water for at least 30 seconds (or 500 mL ) to remove any trace metal particulate in advance of sampling where possible.

Table 3-2 indicates the following for surface water samples collected at the landfill in 2022/23:

- A total of 25 duplicate samples were collected and analyzed for 52 parameters. Twenty-three (23) samples had parameters with RPD values greater than $30 \%$ when all concentrations were above the LOQ.
- RPD values for pH (1 sample), conductivity (9 samples), temperature (1 sample), TSS (3 sample), dissolved organic carbon ( 6 samples), total ammonia ( 1 sample), total aluminum ( 6 samples), total arsenic ( 2 samples), total cadmium ( 9 samples), total cobalt ( 3 samples), total copper ( 3 samples), total iron ( 9 samples), total lead ( 7 samples), total manganese ( 7 samples), total nickel ( 5 samples), total selenium ( 1 sample), total zinc ( 4 samples), dissolved aluminum (3 samples), dissolved cadmium ( 7 samples), dissolved copper ( 3 samples), dissolved iron ( 9 samples), dissolved lead (6 samples), dissolved manganese (3 samples), dissolved molybdenum (1 sample), dissolved nickel (1 sample), and dissolved zinc (4 samples) were greater than $30 \%$ on at least one occasion when all concentrations were above the LOQ.
- Overall, surface water field replicates showed good precision for most parameters and the majority of RPD values for the 2022/23 monitoring year were within the acceptable range. Some of the high RPD values (i.e., iron, aluminum, etc.) may be due to disturbance of sediments while sampling under low flow conditions in the drier months.


### 3.2.2 Leachate

Five (5) duplicate leachate samples were collected during April 2022, August, November, and February events and analyzed for various parameters. As shown in Table 3-3, four (4) samples and twenty-four (24) parameters had calculated RPD values that exceeded alarm limits when concentrations in both replicates were above the LOQ. Duplicate sample collected in February 2023 had most of parameters exceeded alarm limits with most total metals above the alarm limit (30\%) possibly due to issues with the filtration and collection process during sampling.

Because landfill leachate is a complex analytical matrix, leachate field replicates showed good precision for most parameters and the majority of RPD values for the 2022/23 monitoring year were within the acceptable range. Leachate quality data is acceptably precise for the purposes of this report.

### 3.3 Statistical Assessment of Temporal Trends

Seasonal variability in water quality parameters can mask the overall trend of parameters in groundwater, surface water and leachate quality data. To better understand long-term trends in water quality, a non-parametric statistical analysis has been employed to evaluate trends in water quality data at Hartland Landfill since 2005. This test, known as the Mann-Kendall test, was used to evaluate temporal trends in contaminant concentrations. The analysis of trends is intended to promote early detection of statistically significant trends (at the $95 \%$ confidence level) in groundwater chemistry at each sampling location. This method does not require normally distributed data and allows for missing data (non-detects) and irregularly spaced measurement periods in the dataset. Non-detect measurements are assigned the value of the detection limit for the purposes of the statistical analysis. Water and leachate quality data at Hartland Landfill often includes concentrations below detection limits and samples are not always collected at regularly spaced intervals. The hypothesis of both increasing and decreasing trends are tested at the same time. The Mann-Kendall test can be used for virtually any water quality or leachate parameter and provides a quantitative means of determining if a given parameter is changing (increasing or decreasing) in a statistically significant manner over time.

A statistical analysis was conducted utilizing data collected between February 2018 and March 2023. A five-year time frame is consistent with the time interval over which most water quality data is plotted within this report and is anticipated to provide a good balance between seasonal variation and the detection of long-term trends in water quality. This analysis is used to identify areas of the landfill where water quality is degrading and/or improving. The time frame over which trend analysis is conducted should be carefully evaluated in conjunction with management and operational changes.

All field and lab replicates were removed from the data set prior to the analysis. Conductivity, ammonia, and chloride concentrations were tested for trends using the Mann-Kendall test as they are considered indicators of leachate at Hartland Landfill. Trends in sulphate and nitrate concentrations were also evaluated because high concentrations have been observed in relation to aggregate generated and used at the site.

The analysis was conducted using data collected from a total of 80 groundwater monitoring wells ( 36 compliance and 44 routine), 8 leachate purge wells, 21 surface water monitoring stations ( 5 compliance and 16 routine), and the leachate compliance monitoring station (Hartland Valve Chamber). The rest of the locations are either inactive or do not have sufficient data for Mann-Kendall analysis. The results of the statistical trend analysis for groundwater, surface water and leachate are provided in Appendices G-1, G-2, and G-3, respectively, and are discussed in relevant sections of this report.

## Table 3-1. Groundwater Quality QA/QC - Relative Percent Difference 2022-2023

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## Table 3-1. Groundwater Quality QA/QC - Relative Percent Difference 2022-2023



## Table 3-1. Groundwater Quality QA/QC - Relative Percent Difference 2022-2023

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| Ssation | Smmot ipe | Cmpane wen | Datesmpmed | comments |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 21:1-1 | ${ }_{\text {fR1 }}$ | $r$ | ${ }^{20220527}$ | Cbarand coumeses | ${ }_{10} 10$. | ${ }_{138}$ | 0044 | 0.43 | 661 | 001 | 0.005 | ${ }_{512}$ | 0088 | 492 | ${ }_{76}$ | 0.1 | 0.079 | 0.323 | 150. | 156 | 0.005 | 0.5 | ${ }_{687}$ | ${ }^{924}$ | 0.522 | 0015 | 0.005 |
| $21.1-1$ | FR2 | $r$ | 20220527 |  | ${ }^{12}$. | 1.62 | 0.99 | 0.45 | 671 | 001 | 0.005 | 519. | 0.038 | 49. | 7.6 | 0.1 | 0.074 | 0.08 | 14. | ${ }^{16}$ | 0.063 | 0.5 | 6.47 | ${ }^{225}$ | 0.516 | 0.015 | 0.005 |
| 771.1 | frm |  | 2022.0527 | Menortupulates | 140. | ${ }_{1.5}$ | 0046 | 0.4245 | 668 | 0.01 | 0.005 | 515.5 | 0088 | 49.15 | 76 | 0.1 | 0.769 | 02015 | 1995 | 15.8 | ${ }_{0}^{000565}$ | 0.5 | 6.52 | ${ }^{245}$ | 0.519 | 0015 | 0.005 |
| 71.1 .1 | кро |  | ${ }^{202205027}$ | Reooftupleates | 0.0\% | 160\% | 10.8\% | 4.5\% | 1.5\% |  |  | 1.9\% | 0.0\% | 02\% | 0.0\% |  | 57\%\% | 1206\% | 0.7\% | 25\% |  |  | 15\% | 0.1\% | 12\% |  |  |
| 721.1 | FR1 | r | 20220927 | Cbearand ooviness | 150. | 1.11 | 0.088 | 0.332 | 7.55 | 0.01 | 0.005 | ${ }_{43} 4$. | 0.0057 | ${ }_{55}$. | ${ }^{8.3}$ | 0.11 | 0.024 | 0.05 | ${ }_{168} 6$ | ${ }^{38} 3$ | 0.0053 | 0.5 | 73 | ${ }^{932}$ | 0.44 | 0.015 | 0.005 |
| 71-1.1 | ¢R2 | $r$ | ${ }^{202209297}$ |  | 150. | ${ }_{26} 6$ | 0092 | 0.354 | 771 | 001 | 0.005 | 44. | 00073 | ${ }_{55}$ | ${ }_{8} 1$ | 0.1 | 0096 | 0.11 | ${ }_{18} 8$. | 31.1 | 0.007 | 0.5 | ${ }^{734}$ | ${ }^{95}$ | 0.451 | 0.02 | 0.005 |
| 771.1. | frm |  | 20220927 | Menortapolates | 150. | 187 | 0.09 | ${ }_{0} 388$ | 763 | 0.0 | 005 | 40. | 0006 | ${ }_{55}$. | 82 | 0.05 | 0035 | 0.1075 | ${ }_{168}$ | ${ }_{34} 4.7$ | 0006 | 0.5 | 738 | ${ }^{941}$ | 04465 | 0.021 | 0005 |
| 771.1 | RPO |  | ${ }_{20220927}$ | RPootapurases | 00\% | 813\% | $44 \%$ | $34 \%$ | 21\% |  |  | 32\% | 24.9\% | 00\% | 24\% |  | $24 \%$ | 47\% | 00\% | $207 \%$ | 36\% |  | 0.5\% | ${ }_{19 \%}$ | $20 \%$ |  |  |
| 721-1. | FR1 | r | ${ }^{2023} 30 \cdot 1 / 2$ | Cbearad olountess | 150. | 281 | 0.1 | 0.46 | ${ }^{835}$ | 0.0 | 0.005 | 487. | 0.081 | ${ }_{56} 6$ | ${ }_{8} 8$ | 0.1 | 0.127 | ${ }^{0.054}$ | ${ }^{173}$ | ${ }^{37,5}$ | 0.005 | 0.5 | ${ }_{7} 7.4$ | 148. | 0.527 | ${ }_{1} .5$ | 0.005 |
| $77_{1-1 / 1}$ | FR2 | $r$ | $20230 \cdot 1.12$ |  | 180. | 321 | 0.11 | 0.43 | 86 | 0.01 | 0.005 | 48. | 0.023 | 57.4 | 89 | 0.16 | 0.128 | 005 | ${ }^{175}$. | ${ }_{39} 8$ | 0.005 | 0.5 | 777 | ${ }_{16} 18$. | 0.53 | 097 | 0005 |
| 721.1 | frM |  | 20230.1 .12 | Menorituplates | ${ }_{155}$. | 301 | 0.105 | 04495 | 8875 | 0.0 | 0.005 | 48. | 0.2075 | 57. | ${ }_{8} 8$ | 0.13 | 0.1275 | 0.052 | 174. | 3865 | 0.054 | 0.5 | 77.75 | 147. | 0.52 | 1235 | 0005 |
| 710.1-1 | 8po |  | 202301.12 | Reoortuplatase | 65\% | 133\% | 95\% | 1.7\% | 29\% |  |  | 48\% | 255\% | ${ }_{14 \%}$ | ${ }^{23 \%}$ |  | 0.8\% |  | 1,1\% | 60\% |  |  | 0.4\% | 14.46 | 0.8\% | 429\% |  |
| 71/1-1 | ${ }_{\text {fr1 }}$ | r | 2023030.07 | Chara ad coouruss | ${ }^{140}$ | 567 | 0.056 | 049 | 774 | 001 | 0.005 | ${ }^{47}$. | 00096 | 479 | ${ }_{6}$ | 0.1 | 0.14 | 0.078 | ${ }^{188}$ | ${ }_{35}$ | 0.0074 | 0.53 | 693 | ${ }^{117}$ | 0.95 | 0015 | 0005 |
| 71-1.1 | FR2 | $r$ | 202303097 |  | ${ }^{12}$. | 154 | 0.114 | 0.401 | 808 | 0.0 | 0.005 | ${ }^{43}$. | 0.462 | ${ }^{48} 4$ | ${ }_{68}$ | 0.1 | 0.148 | 0.531 | ${ }_{150 .}$ | ${ }_{46} 6$ | 0.0387 | 0.55 | 702 | ${ }^{120}$ | 0.492 | 0.015 | 0.005 |
| $71.1-1$ | frm |  | ${ }^{2023} 30307$ | Menorituplates | ${ }^{14 .}$ | 10.35 | 0.095 | 0.41 | 7.9 | 0.0 | 0.005 | 48. | 0235 | 48.15 | ${ }_{685}$ | 0.1 | 0.129 | ${ }_{0}^{0.3945}$ | 199. | 409 | 0.0275 | 0.54 | 6.975 | 1185 | 0.9935 | 0.015 | 0.05 |
|  | вpo |  | 20230307 | Reootuturates | 0.0\% | 924\% | ${ }^{682 \%}$ | 4.4\% | 4.1\% |  |  | 0.8\% | 1919\% | 10\% | 1.5\% |  | 223\% | 1488\% | 1.3\% | 279\% | 1297\% | 3,7\% | 1.3\% | 25\% | 0.6\% |  |  |
| 220.1. | ${ }_{\text {FR1 }}$ | $r$ | ${ }^{20220.527}$ | Clarand coouness | 18. | 3.1 | 0.04 | 0.129 | ${ }^{123}$ | 0.02 | 0.01 | 1700. | 0.01 | 88.3 | 71. | 02 | 0.03 | 0.1 | 26. | 33. | 001 | 1. | 13.1 | 117. | 0.61 | 0.02 | 0.005 |
| 22:1.1 | FR2 | r | ${ }^{202202527}$ |  | 16. | ${ }_{16}$ | 004 | 0.14 | ${ }_{117}$ | 002 | 001 | 1800. | 001 | ${ }_{85}$ | $\pi$ | 02 | 0.039 | 0.1 | 207. | 372. | 001 | - | 13.1 | 17. | 0.68 | 0015 | 0005 |
| 22:1.1 | fam |  | ${ }^{202209527}$ | Menordotuplates | 16. | ${ }_{235}$ | 004 | 0.136 | 12. | 002 | 001 | 1895 | 001 | ${ }^{85} 8$ | 74. | 02 | 0.036 | 0.1 | ${ }^{268}$. | 376. | 001 | 1. | ${ }_{13.1}$ | 17. | 0.85 | 00185 | 0.005 |
| 22, 2.1 | Reo |  | 202205.57 | Reoortaplatase | 00\% | 683\% |  | 110\% | 50\% |  |  | 06\% |  | 12\% | 8,1\% |  | 167\% |  | 0.7\% | 248 |  |  | 0.0\% | 0.0\% | 79\% |  |  |
| 22:1.1 | fR1 | $r$ | ${ }^{2022} 21.0 .03$ | Cbearand coumess | 150. | 128 | 002 | 0.138 | ${ }^{128}$ | 0.01 | 0.005 | 1560. | 0.005 | ${ }^{83} 9$ | ${ }^{69}$ | 0.1 | 0.982 | 0.05 | 22. | ${ }_{362}$ | 0.005 | 0.5 | ${ }^{128}$ | ${ }^{13}$. | $00^{087}$ | 0.015 | 0.005 |
| 72, ${ }^{2} \cdot 1$ | FR2 | r | 2022.10 .03 |  | 150. | ${ }_{1}{ }^{3} 8$ | 0.02 | 0.134 | ${ }^{128}$ | 0.01 | 0.005 | 159. | 0.005 | 842 | ${ }^{6}$ \% | 0.1 | 0.996 | 0.05 | ${ }_{26} 23$. | 364. | 0.005 | 0.5 | ${ }^{128}$ | ${ }^{115}$ | 0.69 | 0016 | 0.005 |
| 22:1.1 | нем |  | 2027.10 .03 | Meanoftuploctes | 150. | ${ }^{138}$ | 002 | 0.136 | ${ }^{128}$ | 0.0 | 0.005 | 156. | 0.05 | 8805 | 675 | 0.1 | 0.889 | 0.05 | 2225 | ${ }_{36} 6$. | 0.05 | 0.5 | 128 | 14. | ${ }_{0} .864$ | 0.195 | 0.05 |
| 22, ${ }^{2}$ | RPo |  | ${ }^{2022.10 .03}$ | Reoortuplates | 0.0\% | $7.5 \%$ |  | $29 \%$ | 00\% |  |  | 32\% |  | 0.48 | 15\% |  | $29 \%$ |  | 04\% | 06\% |  |  | 0.0\% | ${ }_{188}$ | ${ }_{188}$ |  |  |
| 22:1.1 | ${ }_{\text {FR1 }}$ | $\checkmark$ | 202301.11 | chara and coouness | 150. | 12 | 004 | 0087 | ${ }^{129}$ | 002 | 001 | 1550. | 001 | ${ }^{815}$ | ${ }_{65}{ }^{\text {6 }}$ | 02 | 0.088 | 0.1 | 25. | ${ }_{35}$. | 001 | 1. | 134 | 10. | 064 | 0015 | 0.005 |
| 22:1.1 | FR2 | $\checkmark$ | ${ }^{202300111}$ |  | 150. | 288 | 002 | 0.094 | ${ }^{13}$ | 001 | 0.005 | 150. | 0.005 | ${ }^{825}$ | ${ }^{6}$. | 0.12 | 00708 | 0.05 | 26. | ${ }_{352}$ | 0.005 | 0.5 | 134 | ${ }^{113 .}$ | 0.70 | 0015 | 0005 |
| 22:1.1 | frm |  | $20230 \cdot 1.11$ | Menoritupleats | 150. | ${ }^{203}$ | 003 | 0.095 | 1295 | 0015 | 00075 | ${ }^{1525}$ | 00075 | ${ }^{82}$ | ${ }_{65}{ }^{6}$ | 0.16 | 0.0794 | 0.075 | 20. | ${ }^{351}$ | 0.075 | 0.75 | ${ }^{13} 4$ | 1105 | 0.8705 | 0015 | 0.005 |
| 72.1.1 | Rpo |  | ${ }^{202300111}$ | Reoortuplates | 00\% | 81,8\% |  | ${ }_{7} 7.70$ | ${ }^{\text {08\% }}$ |  |  | ${ }^{33 \%}$ |  | ${ }^{12 \%}$ | 15\%\% |  | $21.7 \%$ |  | 0.8\% | 0.6\% |  |  | 0.0\% | ${ }_{4}^{45 \%}$ | 9, \% |  |  |
| 22:1.1 | FR1 | $\checkmark$ | 20230306 | Cbearand couveress | 150. | 5.55 | 0.02 | 0.142 | 128 | 0.01 | 0.005 | 1800. | 0.005 | ${ }^{838}$ | ${ }_{66}$ | 0.12 | 0.026 | 0.077 | 26. | 37. | 0.006 | 0.56 | 134 | ${ }^{12}$. | 0.652 | 0.015 | 0.005 |
| ${ }^{2} 2.1 .1$ | ${ }_{\text {FR2 }}$ | $\checkmark$ | ${ }^{202303060}$ |  | 150 <br> 150 <br> 150 | 1.7 <br> 1.65 | ${ }_{0}^{0.09}$ | 0.142 <br> 0.42 | 13 129 129 | 0.02 0005 | ${ }^{0.00}$ | 1600 <br> 18.95 <br> 1 | 0.09 <br> 0.005 <br> 0 | 829 <br> 885 <br> 8. | \% ${ }_{\text {\% }}^{66}$ | 0.2 0.16 | 0.063 <br> 0.088 | ${ }_{\substack{0.1 \\ 0.085}}$ | 263 <br> 2635 <br> 205 | 369 <br> 370.5 | ${ }^{0.018}$ | ${ }_{0}^{1 .}$ | 137 <br> 1385 <br>  <br> 135 | ${ }^{112}{ }^{1125}$ | 0.68 0.811 | ${ }^{0.015}$ | 0.005 <br> 0.005 |
| 22,-1. <br> $\substack{\text { 22,-1 }}$ | ${ }_{\text {Rem }}$ |  | 20230306 | Meno of duplease Reootupuease | 150 <br> 0.0 | ${ }_{\substack{3025 \\ 1028 \%}}$ | 0.03 | 0.142 <br> $0.0 \%$ | ${ }^{129}{ }_{16 \%}$ | 0015 | 00075 | $\underset{\substack{1845 \\ 55 \%}}{\substack{\text { che }}}$ | 00075 | 83, <br> 1, 1.85 | ${ }_{\text {ck }}^{66}$ | 0.16 | ${ }_{\text {coser }}^{0.068}$ | 0.885 | ${ }_{\text {2035 }}^{2045}$ | 330.5 <br> $0.8 \%$ <br> 0. | 0.008 | 078 | 13.55 <br> $22 \%$ <br> 2. | 1125 0.98 0.0 | 0.641 <br> a 48 | 0015 | 0.005 |
| 73.1 .1 | ${ }_{\text {FR1 }}$ | $r$ | ${ }^{202205525}$ | sionly sily and coer | 170. | 0.7 | 0.024 | 0.198 | ${ }^{136}$ | 0.01 | 0.005 | 110. | 00237 | ${ }_{64}^{64}$ | ${ }^{19}$ | 0.11 | 0.365 | 0.41 | 20. | $\bigcirc$ | 0.014 | 0.5 | ${ }^{11.6}$ | ${ }^{12}$ | 0.94 | 0015 | 0.005 |
| 73.11 | fr2 | $r$ | ${ }^{202200525}$ |  | 170. | 0.5 | 0.025 | 0212 | 14. | 001 | 0.005 | 111. | 0.023 | ${ }_{65} 1$ | ${ }^{19}$ | 0.1 | 0381 | 0.4 | 210. | 15 | 00138 | 0.5 | ${ }^{11.5}$ | ${ }^{12}$ | 0.96 | 0.02 | 0.005 |
| 733.14 | frm |  | ${ }^{20220.5} 525$ | Menorituricaes | 170. | ${ }_{0} 0$ | 0025 | 0.25 | ${ }^{138}$ | 0.01 | 0005 | 110.5 | 0.225 | ${ }_{64} 9$ | 19. | 0.105 | 0.353 | 0905 | 2095 | 125 | 0014 | 0.5 | 11.55 | 112. | 0943 | 00185 | 0.005 |
| 73.1 .1 | RPD |  | ${ }_{2020.0525}$ | RPoof fivirases | 00\% | 333\% | 4.16\% | ${ }_{68 \%}$ | 29\% |  |  | 0.9\% | ${ }_{65 \%}$ | 06\% | 00\% |  | 1,198 | $27 \%$ | 0.5\% |  | 4,3\% |  | 09\% | 0.0\% | ${ }_{0.6 \%}$ |  |  |
| 733.1. | ${ }_{\text {FR1 }}$ | $\checkmark$ | 20220927 | siont widid geer | 18. | ${ }^{32}$ | 0.45 | ${ }^{023}$ | ${ }^{15}$. | 0.01 | 0.005 | ${ }^{123 .}$ | 0.509 | ${ }_{65} 5$ | 20. | 0.12 | 0.41 | 0.682 | 210. | ${ }_{88}$ | 0.027 | 0.5 | ${ }^{11.3}$ | 88. | 0.85 | 0.015 | 0.005 |
| 73.1 .1 | ¢R2 | $\checkmark$ | ${ }^{2022092927}$ |  | 16. | 1.71 | 0.92 | 0228 | 158 | 0.0 | 0.005 | ${ }_{118 .}$ | 0.984 | ${ }_{658}$ | 2. | 0.12 | 0.427 | 0.832 | 212 | 52 | 0021 | 0.5 | 11.5 | ${ }_{88}^{88}$ | 0.986 | 0015 | 0.005 |
| 733.1. | нем |  | 20220297 | Menorotupuales | 180. | 247 | 0045 | 0235 | 154 | 001 | 0005 | 1205 | 02787 | ${ }_{6565}$ | ${ }^{20}$ | 0.12 | 0.4185 | 0.662 | 21. | 7. | 0.025 | 0.5 | 11.4 | ${ }^{88} 4$ | 0895 | 0015 | 0.005 |
|  | Rpo |  | ${ }_{2}^{202202027}$ | RPootavicases | 00\% | 615\% | 69\% | 22\% | $52 \%$ |  |  | 4.96 | ${ }_{1653 \%}$ | 0.5\% | 00\% | 0.0\% | $4.19 \%$ | 9,1\% | 09\% | 514\% | 278\% |  | ${ }_{18 \%}$ | 0.7\% | 0.18 |  |  |
| 733.1.1 | ${ }_{\text {fR1 }}$ | $r$ | ${ }^{2023} \mathbf{2} \cdot 1.12$ | Clearand coumess | 180. | 1.08 | 0.034 | 0204 | 14.5 | 0.01 | 0.005 | ${ }^{19}$. | 0.275 | ${ }^{71,3}$ | ${ }^{21}$ | 0.1 | 0.355 | 0.458 | 230. | 4. | 0.046 | 0.5 | 12.5 | 21.5 | 0.904 | 0.015 | 0.005 |
| 73, 71 | FR2 | r | $20230 \cdot 1.12$ |  | 160. | ${ }_{1.12}$ | 0.033 | 021 | 143 | 001 | 0.005 | 122. | 0.902 | 71.8 | 21. | 0.12 | 0.354 | 0.44 | 23. | ${ }_{65}$ | 0.013 | 0.5 | 123 | 21.1 | 0.955 | 0.015 | 0.005 |
| 73.14 | ${ }^{\text {rem }}$ |  | $20230 \cdot 1.12$ | Meanotapopatas | 180. | 109 | 0.035 | 02075 | ${ }^{144}$ | 001 | 0.005 | ${ }^{2205}$ | 0.02835 | ${ }^{1,55}$ | ${ }^{21}$ | 0.11 | 0.345 | 0.495 | 23. | ${ }_{525}$ | 00138 | 0.5 | ${ }^{124}$ | ${ }^{1,3}$ | 0.0295 | 0015 | 0005 |
| 73,1.1 | ${ }^{\text {RPO }}$ |  | ${ }^{20230.12}$ | ${ }^{\text {RPPofotuluases }}$ | 00\% | ${ }_{5}^{5 \%}$ | 30\% | ${ }^{3.4 \%}$ | ${ }^{145 \%}$ |  |  | 25\% | ${ }^{13,1 \%}$ | 07\% | 00\% |  | ${ }^{0.3 \%}$ | 38\% | 00\% | 47\%\% | 118\% |  | 16\% | 0.6\% | ${ }^{\text {55\% }}$ |  |  |
| ${ }_{7}^{73,1.1}$ | ${ }_{\text {FR2 }}$ | $r$ | 202303037 2023037 20, | Coberatcolouress | 180. 170. | ${ }_{4}^{426}$ | 0 | 0.375 0203 | ${ }^{151}$ | ${ }^{0.01}$ | ${ }^{0.005}$ |  | 00181 <br> 103 <br> 103 | 798 <br> 65 <br> 8. | ${ }^{39}$ | 0.11 | 0.128 <br> 0.285 <br> 0 | ${ }^{0.328} 0$ | ${ }_{\text {227. }}^{27 .}$ | 217 | $c011900274$ | 0.54 | 676 <br> ${ }_{121}$ <br> 1 | ${ }_{88,5}^{88,}$ | ${ }_{0}^{0.057}{ }_{0}$ | ${ }^{0.015}$ | 0.005 <br> 0.0082 |
| 73, ${ }^{3}$ | ${ }_{\text {frm }}$ |  | 202303037 | Menorotuplatas | 180. | 9,13 | 0.033 | 0289 | ${ }_{1445}$ | 0.0 | 0.005 | 2005 | 0.5295 | 125 | 30. | 0.05 | 0.1965 | 04735 | 20. | 1935 | 0.01965 | 0.52 | 943 | 828 | ${ }_{0}^{080875}$ | 0015 | ${ }^{00006}$ |
| 73.1 .1 | RPo |  | 20230307 | Reof ofupleatse | 11, 10 | 1067\% | 2428 | $59.5 \%$ | 90\% |  |  | 1209\% | 193\%\% | 20,9\% | 60\% |  | $697 \%$ | 615\% ${ }^{\text {a }}$ | ${ }_{64 \%}$ | 243\% | 789\% ${ }^{\text {a }}$ | $7.69 \%$ | 56\%\% | 14.48 | 373\% ${ }^{\text {a }}$ |  | 48.50 |


|  |  |  |  | max Accopatabe RPo | 30\% | 30\% | 30\% | 30\% | 30\% | 30\% | ${ }^{30 \%}$ | 30\% | 30\% | 30\% | 30\% | 30\% | 30\% | 30\% | 30\% | 30\% | 30\% | 30\% | 30\% | 30\% | ${ }^{30 \%}$ | 30\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | method | Dotection Limit Mou) | 0.02 | 0.02 | 0.02 | 2 | 0.05 | 0.04 | 50 | 0.005 | ${ }_{0} 0.05$ | ${ }_{0} 0.05$ | 1 | 3 | 0.002 | 0.2 | 0.5 | 0.002 | 02 | 0.1 | 0.1 | 0.1 | 1 | 0.1 |
|  |  |  | Limto | $\xrightarrow{\text { a }}$ | Nitite | Netrie 0 Nitrate |  | ${ }_{\text {Phosphous }}^{\text {in }}$ | ${ }_{\text {Pooassum }}^{\text {O2, }}$ |  |  |  |  | ${ }_{\text {Sto }}^{\text {Stonimm }}$ | ${ }_{\text {Suphate }}$ | ${ }_{\text {L }}^{\text {Lutur }}$ | $\underset{\substack{\text { coum } \\ \text { Thalum }}}{\text { ate }}$ | ${ }_{\text {Tm }}$ | ${ }_{\text {Thanium }}^{\text {25 }}$ | U.0.1 | adum | $\frac{0.5}{2 \text { 2ne }}$ | ${ }_{\text {zincosum }}^{0.5}$ | ${ }_{\text {¢ }}^{0.5}$ | ${ }_{\text {conductury }}^{5}$ | ${ }_{\text {Tempersure }}^{0.5}$ |
|  |  |  |  | Fraction | os | ois | ois | ${ }^{1 / 5}$ | dis | 01s | dis | ois | ds | os | os | ois | ois | ois | ois | os | os | os | os | тот | тот | тот |
|  |  |  |  | Unit | man | mal | 上er | yon | mol | yon |  | yon | mon | ner | mor | мя | S | м92 | м92 | vor | ma | 上2 | ข2 | ${ }^{\text {о }}$ | ${ }_{\text {uscm }}$ |  |
| Station | Sample Type | compance wen | Date Sampead | Commens |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 18-1.1 | ${ }_{\text {er1 }}$ | $r$ | ${ }^{20220.0513}$ | Clararand couviress | 0.02 | 0.02 | 0.379 | ${ }_{5} 7$ | 0.25 | 0.04 | 8 800. | 0.005 | 4.79 | ${ }^{90,7}$ | ${ }^{26}$ | ${ }^{8,3}$ | 0.002 | 02 | 0.5 | 0.135 | 0.61 | ${ }_{1.16}$ | 0.1 | ${ }_{7} 74$ | ${ }_{350}$. | 102 |
| 8, 81 | FR2 | $r$ | 202205:13 |  | 0.02 | 002 | ${ }_{0} 381$ | 6. | 0234 | 0.04 | 767. | 0.005 | 483 | 90.1 | ${ }^{27}$ | 7.5 | 0.002 | 02 | 0.5 | 0.47 | 0.63 | 0.34 | 0.1 | 743 | ${ }_{350}$ | 102 |
| 18,-11 | ${ }^{\text {FRM }}$ |  | ${ }^{20220.0513}$ | Meno ofotuplats | 0.02 | 0.0 | 0.38 | 5.95 | 0242 | 0.04 | 7835 | 0.05 | 4.81 | ${ }^{904}$ | 26.5 | 79 | 0.002 | 0.2 | 0.5 | 0.141 | 0.62 | 0.75 | 0.1 | ${ }_{743}$ | ${ }_{30}{ }^{\text {a }}$ | 102 |
| 18.11 | Rpo |  | ${ }^{20220.0513}$ | RPpootupuratases |  |  | 0.5\% | 5.17 | 66\% |  | ${ }_{42 \%}$ |  | 0.8\% | 0.7\% | 38\% | 10.1\% |  |  |  | 85\% | ${ }^{32 \%}$ | 1093\% |  | 0.0\% | 00\% | 0.0\% |
| $181 / 1$ | ${ }_{\text {erl }}$ | $r$ | ${ }^{202300.125}$ | Clarand coouruss | 0.02 | 002 | 0.427 | 5. | 0.34 | 0.4 | 8280. | 0.005 | 569 | ${ }_{918}$ | ${ }^{25}$. | ${ }^{7} 7$ | 0.002 | 02 | 0.5 | 0.146 | 0.49 | ${ }^{133}$ | 0.1 | 7.76 | ${ }^{245}$. | ${ }^{98}$ |
| 18.1 .1 | ${ }_{\text {FR2 }}$ | $r$ | ${ }_{20230.125}$ |  | 0.02 | 0.02 | 0.394 | 5.1 | 0.316 | 0.04 | 882. | 0.005 | 5.71 | 89.1 | 25. | ${ }_{68}$ | 0.002 | 02 | 0.5 | 0.46 | 0.47 | ${ }_{1} .35$ | 0.1 | 7.76 | 245. | ${ }_{9} 9$ |
| 1891-1 | frM |  | ${ }_{202301.25}$ | Meno otupleates | 0.02 | 0.02 | 0.4105 | 505 | 0.315 | 0.04 | ${ }^{8350}$. | 0.005 | ${ }_{5.7}$ | ${ }_{90,45}$ | 25. | ${ }_{7} 25$ | 0.002 | 0.2 | 0.5 | 0.146 | 0.48 | ${ }_{1} .34$ | 0.1 | 7.76 | 245. | ${ }_{98}$ |
| 18,1-1 | Rpo |  | ${ }^{20230.1 .25}$ | RPDofiduplatas |  |  | 80\% | 20\% | 06\% |  | 1.7\% |  | 0.4\% | 30\% | 0.0\% | 124\% |  |  |  | 0.0\% | $42 \%$ | 1.5\% |  | 0.0\% | 0.0\% | 0.0\% |
| 18.1 .1 | ${ }_{\text {FR1 }}$ | $r$ | 20230307 | Clearand coluress | 0.02 | 0.02 | 0.406 | 82 | 0235 | 0.04 | 800. | 0.005 | ${ }_{5} 52$ | ${ }^{913}$ | - | 69 | 0.026 | 02 | 0.87 | 0.154 | 0.51 | 1.34 | 0.1 | ${ }_{7} 7.7$ | 250. | 10. |
| 18.10 | fr2 | $r$ | 202303007 |  | 0.02 | 0.02 | 0.258 | ${ }^{7} 3$ | ${ }_{0} 224$ | 0.04 | 840. | 0.005 | 504 | ${ }^{91.6}$ | ${ }^{25}$. | ${ }_{64}$ | 0.002 | 02 | 0.5 | 0.154 | 0.36 | 0.42 | 0.1 | 7.76 | 250. | 10. |
| 18.1-1 | frM |  | 20230307 | Meno ofuplolatas | 0.02 | 0.0 | 0.332 | 7.75 | 02495 | 0.04 | 822. | 0.005 | 5.13 | 9145 | ${ }^{25}$. | 6.65 | 0002 | 02 | 0.685 | 0.154 | 0.435 | 0.88 | 0.1 | ${ }_{7} 7.7$ | 250. | 10. |
| 18.1-1 | кро |  | 2023030, | RPpoftuplicates |  |  | $44.6 \%$ | 11.6\% | 4.46\% |  | $5.4 \%$ |  | 35\% | 0.3\% | 0.0\% | 7,5\% |  |  |  | 0.0\% | 34.5\% | 1095\% |  | 0.0\% | 0.0\% | 0.0\% |
| 21-1-1 | ${ }_{\text {eri }}$ | $r$ | ${ }^{2022.0520}$ |  | 0.02 | 0.02 | 0.1 | 10. | 025 | 02 | 19500. | 0.025 | ${ }^{831}$ | ${ }_{425}$ | 12. | 15. | 0.01 | 1. | 25 | 0.021 | 1. | 198 | 0.5 | 832 | ${ }^{127}$. | ${ }^{11.4}$ |
| 21-1.1 | FR2 | r | 20220.5.20 |  | 0.02 | 0.02 | 0.1 | 10. | 025 | 02 | 19800. | 0.025 | ${ }^{873}$ | 416. | 12. | 15. | 0.01 | 1. | 25 | 0.015 | 1. | 0.5 | 0.5 | 8.32 | ${ }^{127 .}$ | 11.4 |
| $21.1-1$ | frM |  | ${ }_{20220.5}^{20}$ | Meno ofoplicates | 002 | 0.02 | 0.1 | 10. | 025 | 0.2 | 19650. | 0.02 | 8.52 | 420.5 | 12. | ${ }_{15} 5$ | 0.01 | 1. | 25 | 0.018 | 1. | 10.15 | 0.5 | 8.32 | 127. | 11.4 |
| $21.1-1$ | ${ }^{\text {Rpo }}$ |  | ${ }^{20220.0520}$ | RPpootupuricases |  |  |  |  |  |  | 15\% |  | 4.9\% | 2,1\% | 0.0\% |  |  |  |  | 333\% |  |  |  | 0.0\% | 0.0\% | 0.0\% |
| 21.14 | ${ }_{\text {FR1 }}$ | $r$ | 2022:0921 |  | 002 | 002 | 0.063 | ${ }^{3.1}$ | 0219 | 0.04 | 19400. | 0.005 | 897 | ${ }_{45} 5$. | 11 | ${ }^{34}$ | 0.002 | 02 | 0.5 | 0.0214 | 025 | 0.53 | 0.1 | 8.11 | ${ }^{13}$. | ${ }^{128}$ |
| $21.1-1$ | FR2 | $r$ | 2022.0921 |  | 0.02 | 0.02 | 0.032 | 25 | 0215 | 0.04 | 19800. | 0.005 | ${ }_{88} 8$ | 453. | 1. | ${ }_{36}$ | 0.002 | 02 | 0.5 | 0.027 | 022 | 0.32 | 0.1 | 708 | ${ }^{139}$. | 128 |
| 21-1-1 | frM |  | ${ }^{202209021}$ | Meno otupleates | 0.02 | 0.02 | 0.047 | ${ }^{28}$ | 0.217 | 0.04 | 19800. | 0.005 | 892 | 455. | ${ }^{11}$ | ${ }_{3} 5$ | 0.002 | 02 | 0.5 | 0.0275 | 0235 | 0.425 | 0.1 | 7.595 | ${ }^{13}$. | ${ }^{128}$ |
| $21.1-1$ | RPD |  | ${ }^{2022} 20.921$ | RPooftioficates |  |  | 66,3\% | 22.48 | ${ }_{1.8 \%}$ |  | 0.0\% |  | 1,198 | 1,19\% | 0.0\% | ${ }_{5} 57 \%$ |  |  |  | ${ }_{14 \%}$ | 128\% | $499 \%$ |  | 138\% | 0.0\% | 0.0\% |
| 217.1 | ${ }_{\text {FR1 }}$ | $r$ | ${ }^{2022} 2 \cdot 2 \cdot 15$ |  | 0.02 | 0.02 | 0.147 | 162 | 023 | 0.08 | 2880. | 0.0 | ${ }^{8.58}$ | 49. | ${ }^{11}$ | 6. | 0.0108 | 0.4 | 1. | 0.058 | 0.4 | 0.55 | 02 | ${ }^{8.51}$ | 13. | 102 |
| $21+1.1$ | fr2 | $r$ | ${ }^{2022 \cdot 2 \cdot 12 \cdot 15}$ |  | 0.021 | 0.021 | 0.4 | ${ }^{47}$. | 025 | 0.2 | 20400. | 0.025 | 879 | ${ }_{4}^{43}$. | 11. | 15. | 0.046 |  | 25 | 0.24 | 1. | 3.71 | 0.5 | ${ }^{8.51}$ | 133. | 10.2 |
| $221-1$. | frm |  | 2022.21 .15 | Mean ofutuluates | 0.025 | 0.205 | 02735 | ${ }^{31.6}$ | 024 | 0.14 | 2060. | 0.175 | 88.85 | 467. | 1. | 10.5 | 0.024 | 0.7 | 175 | 0.1518 | 0.7 | 2.13 | 0.35 | 8.51 | 13. | 102 |
| $21.1-1$ | RPD |  | ${ }^{2022} 212.15$ | RPDo ofuplofates |  | 4.9\% | 925\% | 975\% |  |  | 19\% |  | 24\% | 10.3\% | 0.0\% |  | 123\% |  |  | 1228\% |  | $1884 \%$ - |  | 0.0\% | 0.0\% | 0.0\% |
| $21.1-1$ | ${ }_{\text {fri }}$ | $r$ | ${ }^{2023} 30310$ |  | 0.02 | 0.02 | 0.1 | 31. | 025 | 0.2 | 1900. | 0.025 | 827 | 416. | 10. | 15. | 0.01 | 1. | 25 | 0.018 | 1. | 0.79 | 0.5 | 8.53 | ${ }^{125 .}$ | 10.7 |
| 21.1.1. | FR2 | r | 202303:30 |  | 0.025 | 0.025 | 0.1 | ${ }^{25}$. | 025 | 02 | 1870. | 0.025 | 822 | 417. | 10. | 15. | 0.01 | 1. | 25 | 0.013 | 1. | 109 | 0.5 | 8.53 | ${ }^{125 .}$ | 10.7 |
| 21.1 .1 | frм |  | 2023:3910 | Meano futuriatas | 0.025 | 0.025 | 0.1 | ${ }^{28}$. | 025 | 0.2 | 18850. | 0.025 | 8245 | 4165 | 1. | 15. | 0.01 | 1. | 25 | 0.0155 |  | 0.94 | 0.5 | ${ }_{8} .53$ | ${ }^{125 .}$ | 10.7 |
| $221.1-1$ | ${ }_{\text {RPD }}$ |  | ${ }^{2023} 30310$ | RPDot ofulicase | 22\%\% | 222\% |  | 21.46 |  |  | 16\% |  | 0.8\% | 02\% | 0.0\% |  |  |  |  | 323\% |  | 319\% |  | 0.0\% | 0.0\% | 0.0\% |
| 221.12 | ${ }_{\text {fri }}$ | $r$ | 202205.19 | Clararat coluruss | 0.02 | 0.02 | 252 | - | ${ }_{5} .89$ | 0.04 | 12800. | 0.005 | ${ }^{25}$. | 39. | ${ }^{18}$ | - | 0.0037 | 02 | 0.5 | 0.056 | 0.35 | 022 | 0.1 | 7.17 | ${ }^{36}$. | ${ }^{11.3}$ |
| 21.1 .2 | ${ }_{\text {FR2 }}$ | $r$ | ${ }^{20220.05 .19}$ |  | 0.02 | 0.02 | 1.45 | 6.3 | ${ }_{5.76}$ | 0.04 | 13800. | 0.005 | ${ }^{25}$ | ${ }^{38}$. | 24. | 64 | 0.003 | 02 | 0.5 | 0.0583 | 0.3 | 022 | 0.1 | 7.17 | 316. | ${ }^{11.3}$ |
| 21.1 .2 | frM |  | ${ }_{20220.0519}$ | Meanotupluates | 0.02 | 0.02 | 1.995 | 6.15 | 5.225 | 0.04 | 13100. | 0.05 | 25. | 389.5 | 21. | 62 | 0.0035 | 0.2 | 0.5 | 0.0572 | 0.325 | 022 | 0.1 | 7.17 | 316. | ${ }^{11.3}$ |
| 21.1 .2 | Rpo |  | 202205:19 | RPDorfuploates |  |  | 539\% | 49\% | 22\% | 0.0\% | 46\% |  | 0.0\% | 0.8\% | 286\% | 6.5\% | 209\% |  |  | 38\% | 154\% | 0.0\% |  | 0.0\% | 00\% | 0.0\% |
| 21.1 .2 | ${ }_{\text {FR1 }}$ | $r$ | 20220930 | Clarand coouruss | 002 | 0.02 | 1.96 | $\stackrel{7}{ }$ | 6.17 | 0.065 | 13 130. | 0.005 | 249 | 48. | 17. | ${ }_{5} 5$ | 0.0027 | 02 | 0.5 | 0.0732 | 0.24 | 1.18 | 0.1 | 7.58 | 48. | 14.9 |
| 21-1.2 | fr2 | $r$ | 20220920 |  | 0.02 | 0.02 | 309 | 59 | 623 | 0.056 | 12200. | 0.005 | 25. | 486. | 16. | 54 | 0.0023 | 0.2 | 0.5 | 0.0719 | 0.25 | 0.48 | 0.1 | 7.48 | 48. | 149 |
| 21-1.2 | frm |  | 20220.920 | Mean ofupuratas | 002 | 0.02 | 2525 | ${ }^{645}$ | ${ }^{6} 2$ | 0.0065 | 13000 | 0.005 | $22^{25}$ | 487.5 | 16.5 | ${ }_{5.35}$ | 0.025 | 02 | 0.5 | 0.0725 | 0.245 | 0.82 | 0.1 | 753 | ${ }^{489}$ | 149 |
| 21.1 .2 | Rpo |  | 20220920 | RPDofidulicase |  |  | 4488 | 17, \%\% | 1.0\% | 149\% | 1.5\% |  | 36\% | 0.8\% | 6.1\% | 1.9\% | 160\% |  |  | ${ }^{18 \%}$ | 4.1\% | 878\% |  | 13\% | 0.0\% | 0.0\% |
| 21-1.2 | ${ }_{\text {eri }}$ | $r$ | ${ }^{2022} 212 \cdot 14$ | Clarand coouruss | 002 | 0.02 | ${ }_{365}$ | ${ }_{9} 8$ | 628 | 0.04 | 14900. | 0.005 | 26. | 59. | 17. | 59 | 0.0043 | 02 | 0.63 | 0.0801 | 0.54 | 1.75 | 0.1 | ${ }_{69}$ | 410. | 10.9 |
| 21.1 .2 | ${ }_{\text {FR2 }}$ | $r$ | ${ }^{2022} 2.21 .14$ |  | 002 | 002 | ${ }^{23}$ | 8.1 | 622 | 0.04 | 14700. | 0.005 | 264 | ${ }_{552}$ | 17. | ${ }^{6}$ | 0.0058 | 02 | 0.5 | 0.0873 | 0.38 | 0.56 | 0.1 | ${ }^{69}$ | 410. | 10.9 |
| 21-1.2 | frм |  | 2022.12 .14 | Meno futupratas | 0.02 | 0.02 | 3.075 | ${ }^{895}$ | ${ }^{624}$ | 0.04 | 14800. | 0.005 | 26.65 | 550.5 | 17. | ${ }_{595}$ | 000055 | 02 | 0.565 | 0.087 | 0.46 | 1.155 | 0.1 | ${ }_{69} 9$ | 410. | 10.9 |
| 21-1.2 | ${ }^{\text {Rpo }}$ |  | 2022.12 .14 | RPDoftupleases |  |  | 50.46 | 190\% | 0.6\% | 0.0\% | 1.4\% |  | 1.9\% | 0.5\% | 0.0\% | 1.7\% | 29,7\% |  |  | 86\% | 34.88 | 1030\% |  | 0.0\% | 0.0\% | 0.0\% |
| 21.1 .2 | ${ }^{\text {FR1 }}$ | $r$ | ${ }^{2023230310}$ | Clararad coourusess | 002 | 0.02 | ${ }^{1.46}$ | 7.1 | ${ }_{5}^{529}$ | 0.04 | ${ }^{12500}$. | ${ }^{0.005}$ | ${ }^{236}$ | ${ }^{399}$ | 18. | ${ }_{58}$ | ${ }^{0.0867}$ | 02 | 0.56 <br> 25 | ${ }^{0.0495}$ | ${ }^{0.38}$ | ${ }_{0}^{0.5}$ | 0.1 | 7.14 | ${ }^{318}$ | ${ }^{11.1}$ |
| ${ }_{\text {2l-1.2 }}^{21+1}$ | ${ }_{\text {FR2 }}^{\text {FR }}$ | $r$ | 2023.39 .10 20230310 |  | 002 <br> 002 <br> 0 | 0.02 002 | 1.58 149 | ${ }_{8,}^{82}$ | 508 5.175 | 0.04 <br> 0.04 | 1230. <br> 12200 <br> 1 | 0.005 <br> 0.005 | 228 238 238 | 372. <br> 3805 | ${ }_{18}^{18 .}$ | 46 <br> 52 | 0.0097 <br> 0.028 <br> 0 | 02 | 255 <br> 1555 <br> 1 | $\ldots$ | O.85 0.59 | ${ }^{0.8}$ | ${ }^{0.1}$ | 7.14 <br> 7 <br> 14 | ${ }^{318}$ | ${ }^{11.1}$ |
| 21-1.2 | ${ }_{\text {Rep }}$ |  | ${ }_{\text {20230303030 }}$ | Mean ofuplatass | 0.02 | 0.02 | 1.49 $40 \%$ 4 | 258\% | S.476 |  | (1200\% | 0.005 | 238 <br> $3.4 \%$ | ${ }_{45 \%}$ | 0.0\% | 23.16 |  |  | ${ }^{1250 \%}$ | -0.4\%4 | ${ }^{88,1 \%} 9$ | 462\% | 0.1 | 2.4\% <br> $0.0 \%$ | 318 310\% 0.0 | ${ }_{\substack{11.1 \\ 0.0 \\ 0.0}}$ |
| 21.2 .1 | FR1 | $r$ | 2022.05,19 | silty and orave | 002 | 002 | 143 | 76 | ${ }_{567}$ | 0.04 | 12700 | 0.005 | ${ }^{24,3}$ | 37. | ${ }^{24}$ | ${ }^{63}$ | 0.0026 | 02 | 0.5 | ${ }^{0.0587}$ | 029 | 0.32 | 0.1 | 683 | 34. | 10.9 |
| 212.2.1 | FR2 | $r$ | 202205:19 |  | 0.02 | 0.02 | 1.46 | ${ }_{6} 6$ | 568 | 0.04 | 12800. | 0.005 | 25. | ${ }_{332}$ | ${ }^{24}$ | 6.5 | 0.032 | 02 | 0.5 | 0.583 | 0.3 | 0.8 | 0.1 | 6.83 | 34. | 10.9 |
| 21.2 .1 | frM |  | 202205:19 | Menor ofupleatas | 0.02 | 002 | 1.445 | ${ }^{6} 9$ | 5675 | 004 | 12750. | 0.005 | ${ }_{2465}$ | 3305 | 24. | ${ }_{6} 6$ | 0.029 | 02 | 0.5 | 0.0575 | 0.295 | 0.56 | 0.1 | 683 | 34. | 10.9 |
| 21.2 .1 | кро |  | 202205:19 | RPpoftuplicates |  |  | 21\% | 2036\% | 02\% |  | 0.8\% |  | 28\% | 0.8\% | 0.0\% | 3.1\% | 207\% |  |  | 28\% | 3.4\% | $8557 \%$ |  | 0.0\% | 0.0\% | 0.0\% |
| 21.2 .1 | ${ }_{\text {FR1 }}$ | $r$ | ${ }_{20220920}$ | Clarand colouress | 0.023 | 0.02 | 218 | 5.1 | 6.18 | 0.051 | 1280. | 0.005 | 27.5 | 488. | 13. | ${ }_{4}{ }^{3}$ | 0.0021 | 02 | 0.5 | 0.1 | 023 | 0.42 | 0.1 | 7.54 | ${ }^{483}$. | 142 |
| 21.2 .1 | FR2 | r | 202209020 |  | 0.021 | 0.021 | 224 | 62 | ${ }_{6.36}$ | 0.046 | 12900. | 0.005 | 28.2 | ${ }_{493}$ | 14. | 4.5 | 0.0023 | 02 | 0.5 | 0.101 | 0.28 | ${ }^{1.14}$ | 0.1 | 6.1 | 483. | 142 |
| 21.2 .1 | frм |  | 20220920 | Meanotupuprats | 0.02 | 0.02 | 221 | 565 | ${ }^{627}$ | 0.045 | 12850. | 0.005 | 2785 | 480.5 | 13.5 | 4.4 | 0.0022 | 02 | 0.5 | 0.1005 | 0.245 | 0.78 | 0.1 | 682 | 483. | 142 |
| 21.2 .1 | ${ }^{\text {Rpo }}$ |  | 20220920 | RPDoftupleates | 9.1\% | 9.1\% | 27\% | 19.9\% | 29\% | 10.3\% | 0.8\% |  | 25\% | 1.0\% | ${ }^{7} 4.4$ | 4.5\% | 9.1\% |  |  | 1.0\% | 122\% | ${ }^{\text {223\% }}$ |  | 21.18 | 00\% | 0.0\% |
| ${ }^{212.2 .1}$ | ${ }^{\text {FR1 }}$ | $r$ | ${ }^{2022921215}$ | Clarand colouress | 0.02 | 002 | ${ }^{226}$ | ${ }^{78}$ | ${ }_{6}^{63}$ | 0.064 | ${ }^{13300}$ | ${ }^{0.005}$ | ${ }^{313}$ | ${ }_{4}^{483}$ | 17. | ${ }_{58}^{58}$ | ${ }_{0}^{0.0077}$ |  | 0.5 | ${ }^{0.1}$ | ${ }^{0.42}$ | 0.42 | 0.1 | ${ }^{688}$ | ${ }^{404}$ | ${ }^{10.7}$ |
| ${ }_{\text {21:2, }}^{2 \cdot 12 \cdot 1}$ | ${ }_{\text {FR2 }}$ | r | ${ }^{2022} 2.21 .15$ | Meanotapurates | 0.02 <br> 0.02 | 0.02 0.02 | 226 226 | 74 | 64 <br> 6.36 | $\xrightarrow{0.04}$0.052 | (13800. | 0.005 <br> 0.005 | 31.6 <br> 31.45 | s00. 49.5 | ${ }^{18 .} 17.5$ | 59 <br> 5.5 | ${ }_{0}^{0.0062}$ | - 02 | 0.5 | 0.0825 <br> 0.0925 <br> 0 | 0.37 <br> 0.355 | 0.56 0.49 | 0.1 | 6.88 <br> 6.88 <br> 8 | 404. <br> 404. | 10.7 <br> 10.7 |
| 2 21.2.1 | Rpo |  | 2029212.15 | RPooftuplorates |  |  | 00\% | 10\%\% | 13\% |  | 0.7\% |  | 10\% | 35\% | 5.7\% | 1.7\% | $21.8 \%$ |  |  | 192\% | 127\% | 286\% |  | -0.0\% | -0.0\% | 0.0\% |






## sw..54



[^0]RPD Realive enerent differeneoo of fied replicaes


Table 3-2. Surface Water Quality QA/QC - Relative Percent Difference 2022-2023

|  |  |  |  | max Accoprabe RPo | 30\% | 30\% | 30\% | ${ }^{30 \%}$ | 30\%\% | 30\%\% | 30\%\% | 30\%6 | 30\%6 | 30\%6 | ${ }^{30 \%}$ | 30\% | 30\% | 30\%6 | 30\%6 | 30\%\% | 30\%6 | ${ }^{30 \%}$ | 30\%\% | ${ }^{45 \%}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | ${ }_{0}^{0.5}$ | 0 | $\bigcirc$ | $\bigcirc$ | ${ }_{2,5}^{0.5}$ | ${ }_{0}^{0.1}$ | ${ }_{50}^{10}$ | ${ }_{0}^{0.0025}$ | O.0.05 | 0.05 <br> 0.25 <br> 0.0 | 0.5 2.5 2 | $\frac{1}{5}$ | ${ }^{0.005}$ | 0.05 <br> 0.25 | ${ }_{\text {cose }}^{0.05}$ |  |  |  | ${ }_{\substack{0.5 \\ 0.5 \\ 20.0}}$ | ${ }^{0.5}$ |
|  |  |  |  | $\underset{\substack{\text { Parameater } \\ \text { Fration }}}{\text { ate }}$ | $\underset{\substack{\text { zinc } \\ \text { Tor }}}{\text { cot }}$ | ${ }_{\text {Prer }}^{\text {Pror }}$ | Conductivit | Temporatue | dis | $\substack{\text { Assenic } \\ \text { ois }}$ | coich | ${ }_{\text {caimium }}^{\text {cois }}$ | Cobat | Copper |  | $\substack{\text { loon } \\ \text { lis }}$ | Llad | Manganes | Mopboen | Nickel | ${ }_{\text {Solemilem }}^{\text {Sols }}$ | ${ }_{\substack{\text { Silwer } \\ \text { Sis }}}$ | $\underset{\substack{\text { zinc } \\ \text { Dis }}}{ }$ | ado orgic |
|  |  |  |  |  | war | ${ }^{\text {pH }}$ |  |  | ${ }^{190}$ | ${ }_{\text {M9\% }}$ | ${ }_{\text {col }}^{\text {cos }}$ | ${ }_{\text {M92 }}$ | L9\% | $\xrightarrow{190}$ | ${ }_{\text {mgl }}$ |  | ${ }^{190}$ | $\stackrel{102}{ }$ | ${ }_{\text {cı }}^{19}$ | $\stackrel{191}{19}$ | vol | $\xrightarrow{\text { Log }}$ | ${ }_{\text {cos }}$ | mgl |
| Staton | ${ }_{\text {Sample }}^{\substack{\text { sampe } \\ \text { Type }}}$ | (compane | Iee samplea | Comments |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| sw-N/4151 | FR1 | $r$ | ${ }^{2023.0105}$ | Claar. colouless. | 1.2 | 7.67 | 360. | ${ }^{84}$ | 5.14 | 0.161 | ${ }^{17}$ | 0.005 | ${ }^{0.0463}$ | 0.263 | 167. | 26.1 | 0.005 | 33.5 | 0.26 | 0.083 | 0.134 | 0.005 | 0.54 | 2. |
| sw-N/4/151 | fr2 | $r$ | 2023.0105 | clear, colouress. | ${ }^{1.1}$ | 7.87 | ${ }^{360}$. | ${ }^{8.4}$ | 5.44 | 0.166 | 17. | 0.005 | 0.0465 | 0.257 | ${ }^{165 .}$ | 27.7 | 0.005 | ${ }^{33} 2$ | 0.259 | 0.087 | 0.124 | 0.005 | 0.54 | 1.9 |
| sw-N/41s1 | fRM |  | ${ }^{2023.010 .05}$ | Mean oftupiciates | 1.15 | 7.87 | ${ }^{360}$. | ${ }^{8.4}$ | 5.29 | ${ }^{0.1835}$ | 17. | 0.005 | 0.0464 | 0.26 | 166. | 26.9 | 0.005 | ${ }^{33,35}$ | 0.295 | 0.085 | 0.129 | 0.005 | 0.54 | 1.95 |
| sw-N/4151 | RPD |  | ${ }^{2023.010 .05}$ |  | 8.7\% | 0.0\% | 0.0\% | 0.0\% | 5.7\% | 3.1\% | 0.0\% | 0.0\% | 0.4\% | 23\% | 1.2\% | 5.9\% | 0.0\% | 0.9\% | 0.4\% | 4.7\% | 7.8\% | 0.0\% | 0.0\% | 5.1\% |
| sw-N/4151 | fr1 | $r$ | ${ }^{2023.022 .21}$ | Low fow, claearan colouress. | 0.15 | 7.99 | ${ }^{360}$. | 7.2 | 4.04 | 0.17 | 19. | 0.005 | 0.0415 | 0.246 | ${ }^{176}$ | ${ }^{23.1}$ | 0.005 | ${ }^{328}$ | 0.276 | 0.093 | 0.091 | 0.005 | ${ }^{0.37}$ | 1.6 |
| sw-N/4/1s1 | fr2 | $r$ | $2023.02 \cdot 21$ | Low fow, claar and colouress. | 1.8 | 7.98 | ${ }^{236}$. | 7.2 | 4.33 | 0.18 | 20. | 0.005 | 0.0418 | 0.444 | 174. | ${ }^{24,3}$ | 0.0124 | ${ }_{33,2}$ | 0.261 | 0.098 | 0.085 | 0.005 | 0.66 | 1.8 |
| sw-N/4151 | fRM |  | $2023.02 \cdot 21$ | Meano ofuplicates | 0.975 | 7.98 | 298. | 7.2 | 4.185 | 0.175 | 19.5 | 0.005 | ${ }^{0.04165}$ | 0.345 | ${ }^{175}$ | ${ }^{23.7}$ | 0.0087 | ${ }^{33}$ | 02865 | 0.0955 | 0.088 | 0.005 | 0.515 | 1.7 |
| sw-N/4/151 | RPD |  | ${ }^{2023.30221}$ |  | 1692\% a | 0.0\% | 41.6\% a | 0.0\% | 6.9\% | 5.7\% | 5.1\% | 0.0\% | 0.7\% | 574\% | 1.19\% | 5.1\% | 85.1\% b | 1.2\% | 5.9\% | 52\% | 6.9\% | 0.0\% | 56.3\% a | 11.9\% |
| sw-N/4281 | fr1 | $r$ | ${ }^{2022} 2.50 .04$ |  | 1.0 | ${ }^{7.91}$ |  |  | 6.99 | 0.072 | ${ }^{43}$ | 0.005 | ${ }^{0.0511}$ | 0.358 | 166. | ${ }^{16.4}$ | 0.005 | ${ }^{19,3}$ | 0.508 | 0.108 | 0.044 | 0.005 | 0.4 | 1900. |
| sw-N/4281 | fr2 | r | 2022.50 .04 |  | 1.0 | 7.89 |  |  | 5.75 | 0.063 | ${ }_{45}$ | 0.005 | 0.0544 | 0.373 | 167. | ${ }^{13,2}$ | 0.005 | 19.1 | 0.527 | 0.113 | 0.04 | 0.005 | 0.44 | 200. |
| sw-N/42si | fRM |  | ${ }^{2022.050 .04}$ | Mean of tuplicates | 1.0 | 7.9 | ${ }^{276}$. | ${ }^{9.8}$ | ${ }_{6,37}$ | ${ }_{0}^{0.0675}$ | 44. | 0.005 | ${ }^{0.05275}$ | 0.3655 | 166.5 | ${ }^{14.8}$ | 0.005 | 19.2 | ${ }^{0.5175}$ | 0.1105 | 0.042 | 0.005 | 0.42 | 2050. |
| sw-N/42si | RPD |  | ${ }^{2022.050 .04}$ |  | 0.0\% |  |  |  | 195\% | 13,3\% | 4.5\% | 0.0\% | 6.3\% | 4.1\% | 0.6\% | 21.\% | 0.0\% | 1.0\% | 3.7\% | 4.5\% | 9.5\% | 0.0\% | 9.5\% | 14.6\% |
| sw-N/42si | ${ }_{\text {FR1 }}$ | $r$ | ${ }^{2022} 2.1 .109$ |  | ${ }_{5} 5$ | ${ }^{8}$ | 305. | ${ }^{4.3}$ | 5.63 | 0.47 | 59. | 0.0057 | ${ }^{0.0427}$ | 0.421 | 25. | ${ }^{9.4}$ | 0.0115 | 7.08 | 0.628 | 0.159 | 0.079 | 0.005 | 1.34 | ${ }^{30}$. |
| sw-N/4281 | FR2 | r | ${ }^{2022}$ |  | 4.1 | 7.07 | ${ }^{305 .}$ | ${ }^{4.3}$ | 6.19 | 0.148 | ${ }^{60}$ | 0.005 | 0.045 | 0.448 | 250. | ${ }^{8.5}$ | 0.0101 | 6.74 | 0.654 | 0.164 | 0.069 | 0.005 | 1.65 | 1700. |
| sw-N/42s1 | frM |  | ${ }^{2022}$ | Meano of tupiciates | 4.9 | 7.535 | ${ }^{305}$ | ${ }^{4.3}$ | 5.9 | 0.1475 | ${ }_{59} 9$ | 0.00535 | 0.0436 | ${ }^{0.4345}$ | 250.5 | ${ }^{8.95}$ | 0.0108 | 6.91 | 0.841 | 0.1615 | 0.074 | 0.005 | 1.495 | 1005. |
| sw-N/42si | RPD |  | ${ }^{2022} 21.1 .09$ |  | 327\% a | 123\% | 0.0\% | 0.0\% | 9.5\% | 0.7\% | 1.7\% | 13.1\% | 4.1\% | 62\% | 0.4\% | 10.1\% | 130\% | 4.9\% | 4.1\% | 3.1\% | 13.5\% | 0.0\% | 20.7\% | 138.3\% |
| sw-N/42si | fr1 | $r$ | 2023.00 .05 | clarar colouress. | 25 | ${ }^{7} 24$ | 293. | 7.2 | 4.87 | 0.088 | ${ }^{60}$ | 0.005 | 0.047 | 0.508 | ${ }^{188 .}$ | ${ }^{8.3}$ | 0.0062 | ${ }^{6.75}$ | 0.489 | 0.11 | 0.072 | 0.005 | 0.61 | 24. |
| sw-N/42si | FR2 | r | ${ }^{2023.010 .05}$ | clear. colouness. | ${ }^{1.1}$ | 7.05 | 450. | 7.2 | 4.43 | 0.083 | ${ }^{62}$ | 0.005 | 0.0481 | 0.509 | ${ }_{187}$ | 8.2 | 0.0064 | 6.79 | 0.492 | 0.133 | 0.072 | 0.005 | 0.57 | 25. |
| sw-N/4281 | frM |  | ${ }^{2023.010 .05}$ | Meano oftupiciates | 1.8 | 7.145 | 371.5 | 7.2 | 4.65 | 0.0855 | ${ }^{61}$ | 0.005 | 0.047 | ${ }_{0}^{0.5085}$ | 187.5 | ${ }^{8.25}$ | 0.0063 | 6.77 | 0.4905 | 0.1215 | 0.072 | 0.005 | 0.59 | 245 |
| sw-N/4281 | RPD |  | ${ }^{202300105}$ |  | 77.8\% a | 2.7\% | 423\% a | 0.0\% | 9.5\% | 5.8\% | 3.3\% | 0.0\% | 0.9\% | 0.2\% | 0.5\% | 1.2\% | 32\% | 0.6\% | 0.8\% | 18.9\% | 0.0\% | 0.0\% | 6.9\% | 4.1\% |
| sw-N/42si | FR1 | $r$ | ${ }^{2023.02021}$ | Low fow, claar and colouress. | ${ }^{1.2}$ | ${ }^{7,33}$ | ${ }^{292}$ | ${ }^{6.3}$ | 4.08 | 0.085 | ${ }^{61}$. | 0.005 | ${ }^{0.0421}$ | ${ }^{0.36}$ | ${ }^{203 .}$ | ${ }^{9.6}$ | ${ }^{0.0057}$ | ${ }^{7} 34$ | ${ }^{0.533}$ | 0.122 | 0.054 | 0.005 | ${ }^{0.47}$ | ${ }^{27}$ |
| sw-N/4281 | FR2 | $r$ | 2023.022 .21 | Low fow, cleara and colouress. | ${ }^{1.3}$ | ${ }^{7} .33$ | 450. | ${ }_{6}{ }^{3}$ | 4.1 | 0.084 | 56. | 0.005 | 0.0419 | 0.396 | 205. | ${ }^{8.8}$ | 0.0058 | 7.57 | 0.549 | 0.115 | 0.057 | 0.005 | 0.65 | ${ }^{28}$ |
| sw-N/42si | fRM |  | ${ }^{2023.022 .21}$ | Mean ofupupleates | 1.25 | ${ }^{7} 33$ | 37. | ${ }_{6} 6$ | 4.09 | 0.0845 | 58.5 | 0.005 | 0.042 | 0.378 | 204. | ${ }^{92}$ | 0.00575 | 7.455 | 0.541 | 0.1185 | ${ }^{0.0555}$ | 0.005 | 0.56 | 275 |
| sw-N/42si | RPD |  | 2023.02221 |  | 8.0\% | 0.0\% | 426\% a | 0.0\% | 0.5\% | 1.2\% | 8.5\% | 0.0\% | 0.5\% | 9.5\% | 1.0\% | 8.7\% | 1.7\% | 3.1\% | 30\% | 59\% | 54\% | 0.0\% | 32.1\% a | 3.6\% |
| sw-N.57 | ${ }_{\text {FR1 }}$ |  | 2023.20 .07 | Field charaetesisisis noterecriced. | ${ }^{5} 5$ | 7.94 | ${ }^{423}$. | ${ }^{8.3}$ | ${ }^{18.1}$ | 0.158 | ${ }^{36}$. | 0.006 | 0.64 | 1.12 | 29. | ${ }^{17.2}$ | 0.005 | 29.4 | 1.74 | ${ }^{123}$ | 1.03 | 0.005 | 1.28 | 28. |
| sw-N.57 | FR2 |  | 2023.020 .07 | Field charatesesisiss not eecocted. | 6.6 | 0. | 0. | 0. | 18.2 | 0.177 | 40. | ${ }_{0}^{0.0057}$ | 0.652 | 1.04 | 292 | 16.2 | 0.0061 | 30.4 | 1.79 | ${ }^{1.31}$ | 1.04 | 0.005 | ${ }^{1.1}$ | 330. |
| sw.N.57 | frM |  | 2023.020 .07 | Meano of tupicates | 6.05 | 3.97 | 21.5 | 4.15 | 18.15 | 0.1675 | ${ }^{38}$ | 0.00585 | 0.466 | 1.08 | 29. | 16.7 | 0.0055 | 29.9 | 1.765 | 1.27 | 1.035 | 0.005 | 1.19 | 305. |
| sw-N.57 | Rро |  | ${ }^{2023.320 .07}$ |  | 182\% | 200.0\% ${ }^{\text {a }}$ | $200.0 \%$ a | 200.0\% 2 | 0.6\% | 113\% | 10.5\% | 5.1\% | 1.9\% | 7.4\% | 0.7\% | 6.0\% | 19.8\% | 3.3\% | 28\% | 6.3\% | 1.0\% | 0.0\% | 15.1\% | 164\% |
| sw.s.04 | ${ }_{\text {FR1 }}$ | $r$ | ${ }^{2022.0505}$ |  | ${ }^{22}$ |  |  |  | $2{ }^{213}$ | 0.128 | 60. | 0.005 | 0.0979 | $2{ }^{278}$ | ${ }^{76,7}$ | ${ }^{70.3}$ | 0.0459 | 2.15 | 0.264 | 0.529 | 0.041 | 0.005 | 2.55 | 1700. |
| sw.so4 | FR2 | r | 2022.50 .05 |  | ${ }^{25}$. |  |  |  | 25.5 | 0.138 | $6_{61}$ | 0.0052 | 0.0964 | 275 | ${ }^{76.4}$ | 51. | ${ }^{0.0507}$ | 2.17 | 0.272 | 0.473 | 0.044 | 0.005 | 248 | 1900. |
| sw.s.04 | frM |  | 2022.0505 | Meano futuplates | 23.5 | 7.42 | 157. | 9.5 | 23.4 | 0.133 | ${ }^{60.5}$ | 0.0051 | ${ }^{0.097} 15$ | 27.75 | ${ }^{7} .65$ | 60.65 | 0.0483 | 2.16 | 0.268 | 0.501 | 0.0425 | 0.005 | 2.515 | 1800. |
| sw.s.04 | Rpo |  | 2022.50 .05 | Rpp Catuation | 128\% |  |  |  | ${ }^{17.9 \%}$ | 7.5\% | 1.7\% | 3.9\% | 1.5\% | 1.17\% | 0.4\% | 31.8\% a | 9.9\% | 0.9\% | 30\% | 11.2\% | 7.1\% | 0.0\% | 2.8\% | 11.1\% |
| sw.s.04 | ${ }_{\text {FR1 }}$ | $r$ | ${ }^{2022} 2.1 .108$ |  | 201 | ${ }^{7} 24$ | ${ }^{253 .}$ | ${ }^{9} 9$ | 10.5 | 0.123 | ${ }^{74}$ | 0.0128 | ${ }_{0}^{0.063}$ | 1.61 | 12. | ${ }^{15,1}$ | ${ }_{0}^{0.0063}$ | 0.34 | 0.758 | 0.359 | 0.077 | 0.005 | 1.77 | 1100. |
| sw.So4 | FR2 | r | ${ }^{2022.1 .1 .08}$ |  | 1.86 | ${ }_{6.83}$ | 253. | ${ }_{9} 96$ | ${ }_{9} .62$ | 0.112 | ${ }^{2}$ | 0.0117 | ${ }_{0}^{0.0673}$ | ${ }_{1} .55$ | ${ }^{121 .}$ | 3. | 0.0063 | 0.263 | 0.71 | ${ }^{0.34}$ | 0.084 | 0.005 | ${ }_{1}^{1.78}$ | 88. |
| sw.s.04 | frM |  | ${ }^{2022}$ | Meano ofuplicates | 1.935 | 7.035 | 25. | ${ }^{9} .6$ | 10.06 | 0.1175 | ${ }^{73}$ | 0.01225 | 0.0668 | 1.58 | 120.5 | 9.05 | 0.0063 | 0.3015 | 0.734 | 0.3495 | 0.085 | 0.005 | 1.775 | 995. |
| sw.s.04 | Rpo |  | ${ }^{2022} 21.1 .108$ | RPDCacauation | 7.8\% | 5.8\% | 0.0\% | 0.0\% | 8.7\% | 9.4\% | 27\% | 9.0\% | 1.5\% | 3.8\% | 0.8\% | 133.7\% a | 0.0\% | 25.5\% | 6.5\% | 5.4\% | 8.7\% | 0.0\% | 0.8\% | 21.19\% |
| sw.so4 | FR1 | $r$ | 040012023 | Claar, colouress. | 3.76 | ${ }^{6.83}$ | 174. | 7.1 | 16.1 | 0.102 | ${ }^{54}$ | 0.005 | ${ }^{0.0504}$ | 1.61 | 89.7 | ${ }^{13,3}$ | 0.005 | 0.186 | 0.127 | ${ }^{0.3}$ | 0.077 | 0.005 | 326 | ${ }^{3.7}$ |
| sw.s.04 | FR2 | $r$ | 040012023 | cliar, colouress. | 3.74 | 6.85 | 280. | 7.1 | ${ }^{17.3}$ | 0.113 | 56. | 0.0059 | ${ }^{0.0507}$ | ${ }^{1.61}$ | 89.5 | ${ }_{18,3}$ | 0.0095 | 0.353 | 0.131 | 0.313 | 0.076 | 0.005 | 3.17 | ${ }^{3.8}$ |
| sw.S.04 | frM |  | 04011223 | Meanot fupleates | ${ }^{3}, 75$ | 6.84 | 227. | 7.1 | 16.7 | 0.1075 | 55. | 0.00545 | ${ }^{0.05055}$ | 1.01 | 89.6 | 15.8 | 0.00725 | 0.2695 | 0.129 | 0.3065 | ${ }_{0}^{0.765}$ | 0.005 | ${ }^{3215}$ | ${ }_{3} .75$ |
| sw.s.04 | RPD |  | ${ }^{2023.010 .04}$ | RPD Caluulion | 0.5\% | 0.3\% | 46.7\% a | 0.0\% | 72\% | 102\% | 3.6\% | 16.5\% | 0.6\% | 0.0\% | 0.2\% | 31.6\% a | $62.1 \%$ | 620\% ${ }^{\text {a }}$ | 3.1\% | 42\% | 1.3\% | 0.0\% | 28\% | 2.7\% |
| sw.s.04 | FR1 | $r$ | ${ }^{2023.02 .06}$ | Modearat fow, clear and colouress. | 2.93 | ${ }^{7}, 37$ | 202 | ${ }^{7} 5$ | 12.9 | 0.101 | ${ }^{68}$ | 0.0106 | ${ }^{0.0793}$ | ${ }^{1.75}$ | 102 | 206. | 0.0118 | 1.15 | 0.438 | 0.369 | 0.055 | 0.005 | 1.76 | 7.3 |
| sw. 04 | FR2 | $r$ | ${ }^{2023.020 .06}$ | Verl low fow Clar, couuress. | 2.96 | ${ }^{7} 37$ | 202 | 7.5 | 11.2 | 0.101 | ${ }^{63}$ | 0.0098 | 0.072 | ${ }_{1} 1.68$ | 103. | ${ }^{7} 3$ | 0.0078 | 0.398 | 0.163 | 0.338 | 0.052 | 0.005 | 1.82 | 190. |
| Sw. 04 | FRM | $r$ | ${ }^{2023} 2020.06$ | Mean ofuplicates | 2.945 <br> $1.0 \%$ | 7.37 <br> $0.0 \%$ | 202 <br> $0.0 \%$ | 7.5 <br> $0.0 \%$ | 12.05 14.15 10 | 0.101 $0.0 \%$ | ${ }^{65.5}$ | (0.0.02 | ${ }_{0}^{0.0758}$ | 1.715 <br> $4.15 \%$ | 1025 <br> $1.0 \%$ <br> 1 | 106.65 <br> $188.3 \%$ | 0.0098 <br> $40.8 \%$ | 0.774 <br> $972 \%$ | ${ }_{0}^{0.3005}{ }^{9.5 \%}$ a | (0.3535 | 0.055 $0.5 \%$ 0.6 | 0.005 <br> $0.0 \%$ | 1.79 <br> $.4 \%$ | ${ }_{\text {18852\% }}^{\text {98, }}$ |

## sw..54

| Noles |  |
| :---: | :---: |
| ss | Single sample |
| FR2 | Fiedr rem |


| State | Parameter | MDL | LOQ | $\begin{aligned} & \text { Lab } \\ & \text { Units } \end{aligned}$ | $\begin{array}{\|c\|} \text { Maximum } \\ \text { Acceptable RPD } \end{array}$ | Hartand Valve Chamber | $\begin{array}{c}\text { Hartland Valve } \\ \text { Chamber }\end{array}$ | RPD \% | Hartland Valve Chamber | Hartland Valve Chamber | RPD \% |  | Hartland Valve <br> Chamber <br> FR1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | FR1 | FR2 |  | FR1 | FR2 |  |  |  |
|  |  |  |  |  |  | SAT00110-FR1 | SAT00110-FR2 |  | SAT00110-FR1 | SAT00110-FR2 |  |  | SAT00110-FR1 |
|  |  |  |  |  |  | 26-Apr-2022 | 26-Apr-2022 |  | 05-Aug-2022 | 05-Aug-2022 |  |  | 22-Aug-2022 |
| Conventionals |  |  |  |  |  |  |  |  |  |  |  |  |  |
| тот | Total Sulphide | 0.0018 | 0.009 | mg/L | 30\% | < 0.045 | < 0.045 |  | 0.085 | 0.06 | 34.5\% | a | 0.067 |
| DISS | Dissolved Sulphide | 0.009 | 0.045 | mg/L | 30\% | 0.045 | 0.045 |  | 0.075 | 0.065 | 14.3\% |  | 0.0057 |
| TOT | TSS | 1. | 5. | mg/L | 30\% | 19. | 19. |  | 20. | 22. | 9.5\% |  | 12. |
| тот | BOD | 2. | 10. | mg/L | 30\% | 20. | 20. |  | 30. | 34. | 12.5\% |  | 27. |
| тот | COD | 10. | 50. | mg/L | 30\% | 9420. | 8610. | 9.0\% | 561. | 560. | 0.2\% |  | 127. |
| DISS | Chloride | 1. | 5. | mg/L | 30\% | 330. | 320. | 3.1\% | 520. | 520. |  |  | 410. |
| diss | Sulphate | 1. | 5. | mg/L | 30\% | 130. | 100. | 26.1\% | 17. | 20. | 16.2\% | < | < 1 . |
| тот | Oil \& Grease, Total | 1. | 5. | mg/L | 30\% | 1. | 1. |  | 1. | 1. |  | < | < 1 . |
| тот | Oil \& Grease, Mineral | 2. | 10. | mg/L | 45\% | 2. | 2. |  | 2. | 2. |  | < | < 2. |
| тот | Cyanide - SAD (total) | 0.0005 | 0.0025 | mg/L | 30\% | 0.013 | 0.013 |  | 0.012 | 0.0106 | 12.4\% | $<$ | < 0.01 |
| тот | Cyanide - WAD | 0.0005 | 0.0025 | mg/L | 30\% | 0.01 | 0.012 | 18.2\% | 0.0081 | 0.0036 | 76.9\% | a < | < 0.01 |
| тот | Phenols | 0.015 | 0.075 | mg/L | 45\% | 0.03 | 0.03 |  | 0.0076 | 0.0082 | 7.6\% | $<$ | < 0.015 |
| тот | Ammonia | 1.5 | 7.5 | mg/L | 30\% | 250. | 250. |  | 280. | 290. | 3.5\% |  | 310. |
| тот | Nitrite | 0.005 | 0.025 | mg/L | 30\% | 1.17 | 1.21 | 3.4\% | < 0.5 | 0.5 |  |  | 0.115 |
| тот | Nitrate | 0.02 | 0.1 | mg/L | 30\% | 13.9 | 15.3 | 9.6\% | 2. | 2. |  |  | 0.683 |
| тот | pH | 0.01 | 0.05 | pH | 30\% | 7.1 | 7.1 |  | 8.14 | 8.14 |  |  | 8.43 |
| тот | Conductivity | 0.1 | 0.5 | $\mu \mathrm{S} / \mathrm{cm}$ | 30\% | 3569. | 3569. |  | 3633. | 3633. |  |  | 4994. |
| тот | Temperature | 0.1 | 0.5 | ${ }^{\circ} \mathrm{C}$ | 30\% | 16.4 | 16.4 |  | 24.6 | 24.6 |  |  | 22.5 |
| тот | Dissolved Oxygen | 0.01 | 0.05 | mg/L | 30\% | 1.37 | 1.37 |  | 1.57 | 1.57 |  |  |  |
| тот | ORP | 1. | 5. | mV | 30\% | 113. | 113. |  | 49. | 49. |  |  | - 43. |
| Total Metals |  |  |  |  |  |  |  |  |  |  |  |  |  |
| тот | Arsenic | 0.04 | 0.2 | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 6.8 | 6.74 | 0.9\% | 9.34 | 9.45 | 1.2\% |  | 8.48 |
| тот | Cadmium | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 0.04 | 0.052 | 26.1\% | 0.053 | 0.06 | 12.4\% |  | 0.034 |
| тот | Chromium | 0.2 | 1. | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 44. | 49.8 | 12.4\% | 57.4 | 59.2 | 3.1\% |  | 63.7 |
| тот | Cobalt | 0.02 | 0.1 | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 13.3 | 13. | 2.3\% | 18.8 | 19.7 | 4.7\% |  | 17.6 |
| тот | Copper | 0.2 | 1. | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 10.2 | 10.5 | 2.9\% | 11.8 | 12.3 | 4.1\% |  | 9.08 |
| тот | Iron | 10. | 50. | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 2140. | 2120. | 0.9\% | 3010. | 3100. | 2.9\% |  | 3040. |
| тот | Lead | 0.04 | 0.2 | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 0.667 | 0.664 | 0.5\% | 0.72 | 0.87 | 18.9\% |  | 0.65 |
| тот | Manganese | 0.2 | 1. | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 674. | 656. | 2.7\% | 751. | 753. | 0.3\% |  | 839. |
| тот | Mercury | 0.019 | 0.095 | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 0.019 | < 0.019 |  | 0.038 | 0.038 |  | < | 0.019 |
| тот | Molybdenum | 0.1 | 0.5 | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 1.98 | 2.35 | 17.1\% | 4.04 | 4.24 | 4.8\% |  | 3.36 |
| тот | Nickel | 0.2 | 1. | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 39.8 | 38.4 | 3.6\% | 53.1 | 54.6 | 2.8\% |  | 51.6 |
| тот | Selenium | 0.08 | 0.4 | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 0.42 | 0.46 | 9.1\% | 0.48 | 0.56 | 15.4\% |  | 0.49 |
| тот | Siver | 0.02 | 0.1 | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 0.02 | 0.02 |  | 0.05 | 0.05 |  | < | < 0.05 |
| тот | Zinc | 2. | 10. | $\mu \mathrm{g} / \mathrm{L}$ | 30\% | 16.8 | 15.5 | 8.0\% | 15.2 | 15.1 | 0.7\% |  | 13.1 |
| BTEX |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DISS | Benzene | 0.4 | 2. | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.6 | 0.51 | 16.2\% | 0.48 | 0.4 | 18.2\% | < | < 0.4 |
| DISS | Toluene | 0.4 | 2. | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 1.5 | 1.1 | 30.8\% | 0.4 | 0.4 |  | < | 0.4 |
| diss | Ethylbenzene | 0.4 | 2. | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.4 | < 0.4 |  | 0.85 | 0.48 | 55.6\% | $\mathrm{a}<$ | < 0.4 |
| diss | Xylenes | 0.4 | 2. | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 1.2 | 1. | 18.2\% | 1.6 | 1.3 | 20.7\% |  | 0.78 |
| РАН |  |  |  |  |  |  |  |  |  |  |  |  |  |
| тот | Total PAH | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 6.6 | 7.3 | 10.1\% | 3.7 | 4.6 | 21.7\% |  | 1.2 |
| тот | Acenaphthene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 2.9 | 3.1 | 6.7\% | 0.86 | 0.91 | 5.6\% |  | 0.3 |
| тот | Acenaphthylene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.055 | 0.056 | 1.8\% | 0.024 | 0.021 | 13.3\% |  | 0.016 |
| тот | Anthracene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.17 | 0.18 | 5.7\% | 0.047 | 0.056 | 17.5\% |  | 0.035 |
| тот | Benzo(a)anthracene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.042 | 0.047 | 11.2\% | 0.013 | 0.019 | 37.5\% |  | 0.022 |
| тот | Benzo(a)pyrene | 0.005 | 0.025 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.0078 | 0.0081 | 3.8\% | < 0.005 | 0.0072 | 36.1\% |  | 0.0066 |
| тот | Benzo(b,j) fluoranthene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.011 | 0.012 | 8.7\% | < 0.01 | 0.013 | 26.1\% | < | < 0.01 |
| TOT | Benzo(g,h,i)perylene | 0.02 | 0.1 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.02 | < 0.02 |  | < 0.02 | 0.02 |  | $<$ | 0.02 |
| тот | Benzo(k)fluoranthene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.01 | < 0.01 |  | 0.01 | 0.01 |  | < | 0.01 |
| тот | Chrysene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.037 | 0.042 | 12.7\% | 0.017 | 0.023 | 30.0\% |  | 0.021 |
| тот | Dibenzo(a, , )anthracene | 0.02 | 0.1 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < 0.02 | < 0.02 |  | 0.02 | 0.02 |  | < | 0.02 |
| тот | Fluoranthene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.56 | 0.62 | 10.2\% | 0.16 | 0.19 | 17.1\% |  | 0.17 |
| тот | Fluorene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 1.8 | 2. | 10.5\% | 0.41 | 0.45 | 9.3\% |  | 0.32 |
| тот | Indeno(1,2,3--c, d) pyrene | 0.02 | 0.1 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.02 | < 0.02 |  | < 0.02 | 0.02 |  | $<$ | < 0.02 |
| тот | Naphthalene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.059 | 0.076 | 25.2\% | 0.8 | 1.5 | 60.9\% | a | 0.12 |
| тот | Phenanthrene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.3 | 0.32 | 6.5\% | 0.14 | 0.16 | 13.3\% |  | 0.038 |
| тот | Pyrene | 0.01 | 0.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | 0.31 | 0.39 | 22.9\% | 0.12 | 0.14 | 15.4\% |  | 0.14 |

Table 3-3. Hartland Valve Chamber Leachate Chemistry QA/QC - Relative Percent Difference 2022-2023

| State | Parameter | MDL | LOQ | $\begin{array}{\|l\|l} \text { Lab } \\ \text { Units } \end{array}$ | $\begin{gathered} \text { Maximum } \\ \text { Acceptable RPD } \end{gathered}$ |  | Hartland Valve Chamber |  | Hartland Valve Chamber | RPD \% | Hartland Valve Chamber | Hartland Valve Chamber | RPD \% | Hartland Valve <br> Chamber <br> FR1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | FR1 |  | FR2 |  | FR1 | FR2 |  |  |
|  |  |  |  |  |  |  | SAT00110-FR1 |  | SAT00110-FR2 |  | SAT00110-FR1 | SAT00110-FR2 |  | SAT00110-FR1 |
| Chlorinated Phenols |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| тот | Chlorinated phenols (total) | 0.41 | 2.05 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.41 |  | 0.41 |  |  |  |  |  |
| тот | Total Dichlorophenols | 0.22 | 1.1 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.22 | < | < 0.22 |  |  |  |  |  |
| тот | Total Monochlorophenols | 0.09 | 0.45 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.09 | < | < 0.09 |  |  |  |  |  |
| тот | Total Nonchlorinated phenols | 1.6 | 8. | $\mu \mathrm{g} / \mathrm{L}$ | 45\% |  | 7.3 |  | 11. | 40.4\% |  |  |  |  |
| тот | Total Phenolic Compounds | 0.17 | 0.85 | \#N/A | 45\% |  | 0.21 |  | 0.21 |  |  |  |  |  |
| тот | Total Tetrachlorophenols | 0.17 | 0.85 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.24 | < | < 0.24 |  |  |  |  |  |
| тот | Total Trichlorophenols | 0.24 | 1.2 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.05 | < | < 0.05 |  |  |  |  |  |
| тот | 2-Chlorophenol | 0.05 | 0.25 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.05 | < | < 0.05 |  |  |  |  |  |
| тот | 2,4 + 2,5 Dichlorophenol | 0.1 | 0.5 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.1 | < | < 0.1 |  |  |  |  |  |
| тот | 2,4,6-Trichlorophenol | 0.1 | 0.5 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.1 | < | < 0.1 |  |  |  |  |  |
| тот | Pentachlorophenol | 0.1 | 0.5 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.1 | < | < 0.1 |  |  |  |  |  |
| тот | 2,3,4,5-Tetrachlorophenol | 0.1 | 0.5 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.1 | < | < 0.1 |  |  |  |  |  |
| TOT | 2,3,4,6-Tetrachlorophenol | 0.1 | 0.5 | $\mu \mathrm{g} / \mathrm{L}$ | 45\% | < | 0.1 | < | < 0.1 |  |  |  |  |  |

Notes:
FR1, FR2 - Field replicates 1 and 2.
FRM - Mean of field replicates.
MDL - Method detection limit.
cov - Coefficient of Variation
LOQ - Limit of quantification.
RPD - Relative percent difference
na - Not applicable, some replicates less than the detection limit
a - Relative Standard Difference greater than $30 \%$ for general inorganic parameters/metals and $45 \%$ for organic parameters and all replicates greater than the limit of quantitation.

- Relative Standard Difference greater than $30 \%$ for general inorganic parameters/metals and $45 \%$ for organic parameters, with some replicates less than the limit of quantitation


| State | Parameter | MDL | LOQ | $\begin{aligned} & \text { Lab } \\ & \text { Units } \end{aligned}$ | Hartland Valve Chamber | RPD \% | Hartland Valve Chamber | Hartland Valve Chamber | RPD \% | Hartland Valve Chamber | Hartland Valve Chamber | RPD \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | FR2 |  | FR1 | FR2 |  | FR1 | FR2 |  |
|  |  |  |  |  | SAT00110-FR2 |  | SAT00110-FR1 | SAT00110-FR2 |  | SAT00110-FR1 | SAT00110-FR2 |  |
| Chlorinated Phenols |  |  |  |  |  |  |  |  |  |  |  |  |
| тот | Chlorinated phenols (total) | 0.41 | 2.05 | $\mu \mathrm{g} / \mathrm{L}$ |  |  |  |  |  |  |  |  |
| тот | Total Dichlorophenols | 0.22 | 1.1 | $\mu \mathrm{g} / \mathrm{L}$ |  |  |  |  |  |  |  |  |
| тот | Total Monochlorophenols | 0.09 | 0.45 | $\mu \mathrm{g} / \mathrm{L}$ |  |  |  |  |  |  |  |  |
| тот | Total Nonchlorinated phenols | 1.6 | 8. | $\mu \mathrm{g} / \mathrm{L}$ |  |  |  |  |  |  |  |  |
| тот | Total Phenolic Compounds | 0.17 | 0.85 | \#N/A |  |  |  |  |  |  |  |  |
| тот | Total Tetrachlorophenols | 0.17 | 0.85 | $\mu \mathrm{g} / \mathrm{L}$ |  |  |  |  |  |  |  |  |
| TOT | Total Trichlorophenols | 0.24 | 1.2 | $\mu \mathrm{g} / \mathrm{L}$ |  |  |  |  |  |  |  |  |
| тот | 2-Chlorophenol | 0.05 | 0.25 | Mg/L |  |  |  |  |  |  |  |  |
| тот | 2,4+2,5 Dichlorophenol | 0.1 | 0.5 | $\mu \mathrm{g} / \mathrm{L}$ |  |  |  |  |  |  |  |  |
| тот | 2,4,6-Trichlorophenol | 0.1 | 0.5 | Hg/L |  |  |  |  |  |  |  |  |
| тот | Pentachlorophenol | 0.1 | 0.5 | $\mu \mathrm{g} / \mathrm{L}$ |  |  |  |  |  |  |  |  |
| тот | 2,3,4,5-Tetrachlorophenol | 0.1 | 0.5 | $\mu \mathrm{g} / \mathrm{L}$ |  |  |  |  |  |  |  |  |
| tot | 2,3,4,6-Tetrachlorophenol | 0.1 | 0.5 | Mg/L |  |  |  |  |  |  |  |  |

FR1, FR2 - Field replicates 1 and 2.
FRM - Mean of field replicates.
MDL - Method detection limit.
cov - Coefficient of Variation
LOQ - Limit of quantification.
RPD - Relative percent difference
na - Not applicable, some replicates less than the detection limit.
a - Relative Standard Difference greater than $30 \%$ for general inorganic param
b - Relative Standard Difference greater than $30 \%$ for general inorganic paramı

### 3.4 Summary

In summary, the 2022/23 quality assurance (QA) analysis indicates the following:

- Duplicate sampling frequencies of $13 \%$ ( $42 / 330$ samples) for groundwater, $29 \%$ ( $25 / 87$ samples) for surface water, 42\% ( $5 / 12$ samples) for the Hartland Valve Chamber compliance point, and $14 \%$ ( $9 / 64$ samples) for the overall leachate sampling program. Overall, duplicate samples were collected at a frequency of approximately $16 \%$ ( $81 / 493$ ), which exceeded the targeted duplicate sampling rate of $10 \%$.
- Groundwater sampling and laboratory analysis have produced reliable results. The QA results indicated a total of 42 samples, and 39 analytical results exceeded the RPD alert limits when all parameter concentrations were above the LOQ. Dissolved aluminum, cadmium, copper, iron, lead, nickel, and zinc concentrations should be interpreted with caution, as these seven (7) dissolved metal concentrations had over $20 \%$ of the duplicate samples above the alarm limit where all parameter concentrations were above the LOQ, indicating potential contamination during the sample handling or filtration process.
- Surface water sampling and laboratory analysis have produced reliable results. The QA results indicated a total of 23 samples and 26 analytical results exceeded the RPD alert limits where all parameter concentrations were above the LOQ.
- Leachate sampling and laboratory analysis have produced reliable results. A total of four (4) samples and twenty-four (24) parameters had calculated RPD values that exceeded RPD alert limits when concentrations in both replicates were above the LOQ.
- A Mann-Kendall statistical trend analyses was conducted on water quality data collected from 80 groundwater monitoring wells, 8 leachate purge wells, 21 surface water stations and 1 leachate monitoring point for parameters that are known indicators of leachate and aggregate influences to evaluate temporal trends in water quality at the landfill. The results of trend analyses are discussed in the groundwater and surface water quality sections of this report.


## 4. Groundwater Flow

### 4.1 Data

A review of the sampling program was undertaken in early 2016 (AECOM 2016), and recommendations for modifications to the number, location, and sampling frequency of compliance monitoring stations were implemented throughout the 2020/21 monitoring year. Groundwater elevations at the landfill were measured on a quarterly basis. The groundwater flow interpretation presented in this section is based on the following data:

- Continuous water level and leachate elevation monitoring using the SCADA system for the lower leachate lagoon, upper leachate lagoon, Phase 2 basin, wells GW-36-1-1 and GW-37-1-1., and one monitoring well located north of Phase 1 (GW-40-1-1).
- Continuous water level monitoring in five purge wells south of Phase 1 (P1, P2, P3, P4 and P10).
- Continuous water level monitoring in four monitoring wells east of Phase 1 (GW-17-1-1, GW-18-1-1, GW-54-1-1 and GW-76-1-1).
- Continuous water level and flow monitoring in or near three north purge wells (GW-80-1-0-P8, GW-81-1-0-P9 and GW-52-4-0-P7) and two monitoring wells (GW-52-3-0 and GW-52-1-1) located north of Phase 1.
- Continuous water level monitoring in 10 wells north of Phase 2 (GW-41-1-1, GW-43-1-1, GW-44-1-1, GW-62-1-1, GW-77-1-1, GW-78-1-1, GW-87-1-1, GW-87-2-1, GW-88-1-1 and GW-88-2-1).
- Continuous water level monitoring in four wells in the Hartland North Ridge area (GW-91-1-1, GW-92-1-1, GW-93-1-1 and GW-94-1-1).
- Presence and elevation of topography, refuse, engineered covers, temporary tarps, ditches, and surface water features.
- Daily precipitation data.

Manually measured groundwater elevations for 2022/23 are presented in Appendix A.3. Groundwater flow patterns were interpreted based on groundwater elevations measured in September 2022. The data indicate that there are two separate groundwater flow systems at the Landfill. One is a regional groundwater flow system in the bedrock surrounding and underlying the landfill. The second is a perched system contained within Phase 1. A similar system has not been observed within the Phase 2 landfill thus far, although low permeability liners, tarps, and other barriers to vertical percolation may promote development of localized perched leachate systems during wet weather and over time. Although the two flow systems are separate, the presence of the leachate mound within the waste influences groundwater flow in the bedrock underlying the waste. Understanding these two flow systems is important for evaluating the effectiveness of leachate control and containment measures.

Monitoring wells GW-79-1-1, GW-79-2-1, GW-74-1-1, and GW-74-2-1 were decommissioned in June 2018 to accommodate blasting and site preparation for aggregate storage. The pressure transducer installed in GW-79-1-1 was removed from the well prior to decommissioning. Four (4) new leachate monitoring wells (GW-89-1-1, GW-89-2-1, GW-90-1-1 and GW-90-2-1) were installed in Phase 1 and Phase 2 in 2018/19 to verify leachate capture and support management decisions. In April 2020, four (4) pressure transducers were installed at locations GW-91-1-1, GW-92-1-1, GW-93-1-1 and GW-94-1-1 to continuously monitor water levels in the Hartland North Ridge area.

In 2022/23, monitoring well GW-93-1-1 was destroyed due to blasting activities, and pressure transducers installed in these wells could not be retrieved. All these wells will need to be decommissioned to meet the well abandonment requirements of the British Columbia Groundwater Protection Regulation.

### 4.2 Regional Groundwater Flow in the Bedrock

Figure 4-1 presents an interpretation of regional groundwater flow patterns based on groundwater elevations in bedrock wells and deep wells completed in refuse, as observed in September 2022. Within the landfill footprint, several wells (GW-75-1-1, GW-82-1-1, VLGW-02-D, VLGW-03-D, VLGW-08-D, VLGW-15-D, VLGW-16-D and VLGW-17-D) are screened at or near the bottom of the waste and their water levels are interpreted as being representative of the regional groundwater flow system within bedrock underlying the landfill. Geological structures including the Highland Fault and another inferred fault near
monitoring well GW-77-1-1 are shown in plan on Figure 4-1 and in cross section in Figure 4-2 and Figure 4-3. Measured groundwater elevations and hydrogeologic testing indicates that these structures influence local-scale groundwater flow patterns. Figure 4-2 is a cross-section that extends from north of the leachate lagoons to the south of Phase 1 and depicts groundwater flow in a north-south direction. Figure 4-3 shows groundwater flow in an east-west cross-section extending from the bedrock ridge northwest of Phase 2 to the eastern property boundary.

Groundwater flow patterns were consistent with previous years, with some differences in the North Ridge area. Regional groundwater flow near the landfill is influenced by bedrock structures, topographic relief, and the presence of surface water features. The regional groundwater flow direction is from southwest to northeast from Mount Work toward the north-south trending valley that underlies the northern portions of the Phase 1 and Phase 2 landfills. Groundwater flow in the bedrock valley underlying the landfill is predominantly northward, as shown in Figure 4-1. Most of the leachate-impacted groundwater in the bedrock below the landfill flows northward to the lower leachate lagoon via the Phase 2 basin leachate collection system, micro-tunnel, leachate springs and purge wells (GW-52-4-0-P7, GW-80-1-0-P8, and GW-81-1-0-P9) located north of Phase 1.

As shown on Figure 4-1, there is an inferred groundwater flow divide located near the south end of the landfill and a small portion of groundwater below the south portion Phase 1 flows towards the southeast. This flow divide trends roughly with a bedrock high in the valley floor beneath the landfill. Southeastward groundwater flow below the landfill is constrained by a constructed clay berm/bedrock grout curtain that was installed at the south end of the landfill in the 1980s, and by five purge wells (P1, P2, P3, P4 and P10) that commenced pumping leachate in 2001. Leachate collected by the south purge well network is pumped from the south pumping station northward to the lower leachate lagoon.

Near the North Ridge and Hartland North Pad, located northwest of Phase 2, groundwater flows radially from a topographic high situated north of Phase 2. The water predominantly flows northeast towards Heal Creek and northwards in the direction of Durrance Lake. Throughout 2022/23, continued blasting operations along the North Ridge led to notable topographic and hydraulic alterations, particularly to the north and west of Phase 2. Notably, groundwater elevations in wells 87 and 88, situated south of the Northwest Stockpile, declined by approximately 4-10 m compared to previous years. Although this reduced the magnitude of groundwater gradients, inward hydraulic gradients were maintained, and the Phase 2 hydraulic trap remained intact.

Groundwater elevations associated with the Highland Fault have changed. Groundwater elevations typically increase during wet fall, winter, and spring months when precipitation inputs increase infiltration and raise groundwater elevations. Historically, groundwater elevations during the wet season have been several metres higher west of the Highland Fault than they are immediately east, implying that the Highland Fault is a barrier to west-east groundwater flow. In 2022/23, seasonal fluctuations in groundwater elevations were observed in this area, but they have become less pronounced. Groundwater elevations should be continuously monitored to observe any changes during the drier months, and consideration should be given to characterizing the hydrogeology of the bedrock mass upslope (west) of the Phase 2 landfill to confirm the impact of ongoing quarrying on leachate containment and the surrounding groundwater flow system.

Along the bedrock ridge (trending roughly north-south between the Hartland Landfill and Kiowa Place Road) east of the landfill boundary, groundwater flows inwards toward the northern portion of the Phase 1 landfill. Further east of the landfill boundary, the topography begins to slope eastward towards Tod Creek valley and the groundwater flow direction is likely eastward towards Tod Creek.

### 4.3 Leachate Elevations in Phase 1 and Phase 2

### 4.3.1 Phase 1

Groundwater monitoring wells installed at varying depths in Phase 1 allow for measurement of leachate levels and interpretation of flow directions within the refuse. In addition to the monitoring wells, water level data is collected manually from selected landfill gas wells in Phase 1.

The leachate mound in the Phase 1 landfill is depicted along a north-south cross section in Figure 4-2, based on groundwater and leachate level measurements collected in September 2022. During the operating years of the Phase 1 landfill, the leachate mound within Phase 1 reached an elevation of approximately 160 m asl (above mean sea level). Final cover that incorporated a geomembrane liner was installed on Phase 1 in 1997 to limit infiltration. As shown on Figure 4-2, the leachate mound in the refuse is above the regional bedrock groundwater flow system, and the hydraulic gradient is downward. The downward gradient in the central portion of the landfill reverse to upward gradients due to pumping at the leachate collection
purge wells in the north and south areas of the landfill. In 2022/23, leachate elevations within Phase 1 were generally below 155 m asl, reflecting an approximate 5 m decrease in the elevation of the leachate mound since capping of the Phase 1 landfill.

Figure 4-5 presents leachate levels in 13 landfill gas monitors in the Phase 1 landfill. Although it is difficult to accurately measure leachate levels in landfill gas wells, they provide additional landfill leachate level information. Like previous years, leachate levels in the shallow landfill gas monitors typically show minor (i.e., $\pm 1 \mathrm{~m}$ ) variations, as the refuse has a relatively high porosity and exhibits relatively consistent recharge and discharge patterns. The relatively high leachate elevations in shallow gas wells screened in refuse (e.g., VLGW-21-D and VLGW-26-D) indicate downward vertical gradients from refuse to the underlying bedrock aquifer. In March 2019, a pressure transducer was installed in GW-89-2-1 to record leachate levels continuously. As shown on Figure 4-5, leachate elevations in the shallow well GW-89-2-1 were generally stable over time, ranging from 149.4 to 151.2 m asI. In April 2023, the pressure transducer from well GW-89-2-1 was damaged and data could not be retrieved. Manual leachate elevation data from well GW-89-2-1 indicates that leachates levels have remained below 152.4 m asl.

Like previous years, leachate elevations in the deep gas monitors (VLGW-02-D, VLGW-03-D, VLGW-08-D, VLGW-15-D, VLGW-16-D and VLGW-17-D) continued to fluctuate in response to seasonal variability in groundwater recharge, indicating that the lower portions of the Phase 1 landfill are in hydraulic connection with the regional groundwater flow system. Leachate monitoring well GW-89-1-1 was installed in November 2018 to replace the decommissioned well GW-74-1-1. Well GW-89-1-1 was screened at the bottom of the refuse to facilitate monitoring of the leachate mound. In April 2023, the pressure transducer from well GW-89-1-1 was damaged, and data from 2022/23 could not be retrieved. Manual leachate elevation measurements from well GW-89-1-1 indicate that the leachate elevation was 152.7 m asl on February 3, 2023.

Well GW-75-1-1 is located further downgradient near the north end of Phase 1 and it monitors the deep regional water table. Like previous years, groundwater elevations in well GW-75-1-1 ranged from approximately 128 to 130 m asl in 2022/23 (Figure 4-5). The relatively high leachate elevations in shallow monitors VLGW-21-D ( 142.7 m asl) and VLGW-26-D ( 144.7 m asI), both located within 50 m of well GW-75-1-1 and screened in refuse, indicate that strong downward vertical gradients are present in this area of the landfill.

### 4.3.2 Phase 2 Basin

The Phase 2 landfill is in a large bedrock basin situated immediately west of the north end of the Phase 1 landfill. The Phase 2 landfill is segregated into multiple cells. Cell 1 and Cell 2 are complete, and Cell 3 is active. Quarrying is underway to lower the existing bedrock surface to form the base of Cells 4,5 and 6 . Leachate from Cell 1 and Cell 2 is captured in a 350 mm diameter micro-tunnel by gravity and transported to the Lower Leachate Lagoon. After reporting to the Lower Leachate Lagoon, all leachate is discharged to the sanitary sewer via the leachate pipeline. A geomembrane liner separates Cell 2 and Cell 3. Leachate from Cell 3 is captured via the Toutle Drain, which reports to the Upper Leachate Lagoon prior to discharge into the leachate pipeline and sanitary sewer.

Figure 4-8 presents hydrographs for the groundwater monitoring wells and leachate wells located north of the Phase 2 basin. Groundwater levels north of Phase 2 need to be higher than leachate levels inside the Phase 2 basin for the hydraulic trap to operate. Leachate levels within the Phase 2 basin are typically around $113-114 \mathrm{~m}$ asl, or 8 to 10 m lower than groundwater elevations outside the basin (AECOM 2020b). In 2022/23, groundwater elevations at groundwater monitoring locations, 38, 39, 62,77 , and 78 were higher than leachate levels in the Phase 2 basin, indicating that the leachate collection system functioned effectively.

Leachate monitoring wells (GW-82-1-1, GW-83-1-1, GW-84-1-1, GW-86-1-1, and GW-90-1-1) were installed to investigate potential leachate mounding within the Phase 2 refuse. Wells GW-84-1-1 and GW-86-1-1 were damaged and subsequently decommissioned. New leachate monitoring wells (GW-90-1-1 and GW-90-2-1) were installed in 2018/19 to verify leachate capture and monitor leachate elevations. Drilling observations and subsequent monitoring indicated that the waste mass was moist but largely unsaturated. Although higher (i.e., >125 m asl) leachate levels in GW-82-1-1 and GW-83-1-1 were observed in the north area of Phase 2 (Figure 4-8), leachate levels were typically within or below the well screens, and well GW-82-1-1 was dry in September and November 2022. Higher leachate levels may be related to the presence of condensate, historical tarping, or low-permeability strata within the refuse. Wells GW-90-1-1 and GW-90-2-1 were dry in 2022/23.

Historically, leachate elevations in Phase 2 were typically about 0.8 m below the leachate elevation in the Lower Leachate Lagoon. However, since January 2022, leachate elevations in Phase 2 have increased, exceeding the leachate level in the Lower Leachate Lagoon in the spring. In 2022/23, the leachate level in the Lower Leachate Lagoon was lower than normal
operating levels, which CRD confirmed that it was likely due to calibration/ instrument drift, and were not indicative of a change in operations. In November 2022, leachate in Phase 2 reached a maximum elevation of approximately 116 m asl ( 0.5 m higher than the leachate elevation in the Lower Leachate Lagoon) in response to heavy precipitation.

Figure 4-8 shows a hydrograph of leachate levels in the lined Upper Leachate Lagoon and the unlined Lower Leachate Lagoon, based on pressure transducer readings recorded by the SCADA system. In 2022/23, leachate levels in the Upper Leachate Lagoon were generally around 124 m asl, reaching a peak elevation of 127.87 m asl $(77 \%$ full) in December 2022, in response to heavy precipitation. Due to the relatively high flow rate of the CRL, water levels gradually decreased and returned to a normal operating level of 124 m asl by January 5, 2023. Leachate levels in the Lower Leachate Lagoon fluctuated around 114.5 m asl throughout the year.

In October and November 2022, the SCADA system in the Lower Leachate Lagoon recorded considerable fluctuations (Figure $4-8)$. The CRD informed AECOM that the Upper Leachate Lagoon underwent maintenance, so the leachate level trends observed in October and November reflect the lagoon being drained periodically. Ultimately, throughout 2022/23, the Lower Leachate Lagoon assumed a substantial role as a leachate storage facility, and the available leachate storage capacity was sufficient.

As landfill development continues, it is imperative that leachate levels in each phase of the landfill are closely monitored to verify seismic stability, confirm that the Leachate Collection System and the Phase 2 hydraulic trap are functioning, and determine if additional leachate containment measures should be implemented. AECOM previously recommended investigating leachate mounding in Phase 2 on a five-year basis, with the next investigation scheduled for 2025.

### 4.4 Groundwater Flow in the Bedrock Aquifer Near the Landfill

### 4.4.1 East of Phase 1

Figure 4-4 shows groundwater elevations at monitoring locations 17, 18, 54, and 76, which are located on the bedrock ridge east of Phase 1. Dramatic changes in water levels at Site 18 have occurred occasionally since 2001. In 2016/17, groundwater elevations in the deepest wells at Site 18 increased by approximately 8 m relative to years prior, resulting in lower westward hydraulic gradients. As per the 2018/19 annual monitoring report recommendations (AECOM 2019a), pressure transducers were installed in wells GW-18-1-1, GW-76-1-1, GW-17-1-1, and GW-54-1-1 to continue monitoring the groundwater elevations and confirm groundwater flow is toward the landfill.

As shown on Figure 4-4, groundwater levels in GW-18-1-1 have returned to historical levels since 2018, indicating that westward hydraulic gradients restabilized. In 2022/23, groundwater levels in GW-18-1-1 remained approximately 6 to 7 m lower than those in GW-76-1-1. Similarly, groundwater elevations at GW-54-1-1 are consistently higher than those in GW-17-$1-1$, indicating a westward component of groundwater flow at these locations. Like previous years, groundwater elevations at monitoring station 76 indicate strong downward hydraulic gradients. However, groundwater elevations at station 18 in 2022/23 indicate low, downward to neutral gradients. Groundwater levels should continue to be monitored in these locations to verify groundwater gradients are directed inward toward the landfill.

### 4.4.2 North of Phase 1

Groundwater quality data collected from wells downgradient of the North Purge Wells indicate that the purge well system has a mitigating effect on the northward migration of leachate. In July 2016, monitoring well GW-81-1-0-P9 became operational in an ongoing effort to increase leachate collection capacity upgradient of well 40-1-1. The influence of pumping the North Purge Wells (GW-80-1-0-P8, GW-52-4-0-P7, and GW-81-1-0-P9) on groundwater flow is illustrated in plan on Figure 4-1, with the 115 m and 120 m water table contours deflecting southward due to the drawdown of the water table surrounding the North Purge Wells. In 2022/23, a total of $14,277 \mathrm{~m}^{3}$ of leachate was removed by the North Purge Wells, which was lower than in 2021/22 (18,140 $\mathrm{m}^{3}$ ) and 2020/21 (24,790 $\left.\mathrm{m}^{3}\right)$. The highest leachate discharge rate was observed in late March 2023, with a peak flow of $102 \mathrm{~m}^{3} /$ day.

Figure 4-6 presents precipitation and groundwater elevation data for monitoring wells located near the Phase 1 North Purge Well System (GW-40-1-1, GW-52-1-1, GW-52-3-0, GW-52-4-0-P7, GW-80-1-0-P8, and GW-81-1-0-P9). Leachate discharge rates from the North Purge Wells (GW-52-4-0-P7, GW-80-1-0-P8 and GW-81-1-0-P9) are presented to illustrate the volume of leachate extracted in response to precipitation events. Monitoring well GW-52-3-0 is the original purge well that operated between 1995 and 1998 and is located within 2 m of well GW-52-4-0-P7. Water levels in well GW-52-3-0 are affected by the pumping rate in GW-52-4-0-P7 and seasonal variations in groundwater recharge. Pressure transducers connected to SCADA
are installed in wells GW-40-1-1, GW-52-3-0, GW-52-4-0-P7, GW-80-1-0-P8, and GW-81-1-0-P9 to provide long-term monitoring of purge well performance north of Phase 1.

As shown on Figure 4-6, leachate discharge from the North Purge Wells increased in February/March 2023 in response to considerable precipitation. A total of $3,007 \mathrm{~m}^{3}$ of leachate was discharged during that period, and the daily average leachate flow was approximately $51.9 \mathrm{~m}^{3} /$ day.

The pressure transducer in well GW-52-1-1 was damaged in March 2022, so no timeseries data was recorded in 2022/23. Groundwater levels in GW-52-3-0 increased substantially in 2021/22 and remained elevated in 2022/23. In 2022/23, the average water level in GW-52-3-0 was 116.5 m asl, which was about 2-2.5 m higher than levels seen before 2021. This may indicate a reduction in the extent of the drawdown cone surrounding GW-52-4-0-P7 and/or GW-80-1-0-P8 due to formation of a biofilm on the inside of the well bore. This should be further investigated by conducting specific capacity tests on each well and comparing it to historical results to determine if well performance has degraded. If deemed to significantly impact well performance, the biofilm should be removed by rehabilitating the purge wells.

Water levels in P7, P8, and P9 also exhibited some variability, and were likely influenced by fluctuations in precipitation and pumping rates. However, the average water levels in P7, P8 and P9 were 112.0 m asl, 113.99 m asl and 118.32 m asl, respectively, where were consistent with their 2021/22 average results. Groundwater elevations in the wells surrounding the North Purge Well System should continue to be monitored to confirm that the North Purge Wells are functioning properly. Regular pump and well maintenance is required to maintain leachate capture and therefore minimize potential future leachate impacts around the lower leachate lagoon.

In August 2022, two new purge wells (P11 and P12) were installed near the edge of the ridge just south of the NWSP. P11 was installed in bedrock, and P12 was completed in overburden or across the overburden-bedrock contact. Groundwater elevations in P11 and P12 ranged from 125.3 to 128.6 m asl, and from 120.3 to 123.6 m asl, respectively. The bedrock high near P11 may direct groundwater in overburden sediments to the east-southeast before it flows north toward the NWSP. In early 2023, P11 was equipped with a pneumatic pump and has been operating intermittently. P12 is dry consistently and will likely not be equipped with a pump.

### 4.4.3 South of Phase 1

The CRD installed six leachate collection purge wells at the south end of the Phase 1 landfill in August 2000 to intercept leachate migrating south of the Phase 1 groundwater divide. Pneumatic pumps were installed in four of these wells (P1, P2, P3, and P4) and they have been in operation since September 2001 (continuously since May 2002). The remaining two wells (P5 and P6) were not outfitted with pumps due to low well yield. An additional well (P10) was installed in 2010 and outfitted with a pneumatic pump to augment pumping capacity south of the landfill. P1 has subsequently been altered and outfitted with an electric submersible pump.

Groundwater elevations measured using pressure transducers in each of the five operational south purge wells are plotted on Figure 4-7. The on/off cycling of the pumps is evident as water levels generally ranged from approximately 133 to 151 m asl. In 2022/23, water levels in all south purge wells remained within their normal ranges, except for P1, P2, and P10. A pump failure may have occurred at P10 in December 2022, when groundwater elevations in the well rapidly increased by approximately 13.6 m to 148.6 m asl. Between January and February 2023, the SCADA System recorded a gradual decline in water elevations to approximately 120 m asl at P2, followed by a rapid increase back to typical levels. The CRD determined this recording resulted from an error in the SCADA System. Generally, large and rapid changes in water levels imply the subsurface materials have a low transmissivity and bulk porosity, and the cone of depression associated with each purge well may be limited in lateral extent. Some fluctuations are related to pump maintenance events and short-term power disruptions.

Groundwater elevations in the south purge wells have fluctuated since 2007. Although elevations have remained somewhat higher than the target pumping elevations, the installation of P10 in September 2010 and the pump upgrades to P1 between 2015 and 2017 resulted in significantly improved drawdown. Unfortunately, P1 required significant maintenance due to well fouling by leachate due to the well/pump design. In November 2018, P1 was re-installed and produces approximately of 1.01 $\mathrm{L} / \mathrm{s}$ of leachate. The replacement well is screened in refuse only from 6.10 to 12.19 m below ground surface. Despite the shallower installation depth, the well removes a similar volume of leachate. In November 2020, the pump in P1 was rebuilt and the pressure transducer was re-installed. The lower set point was changed to 147.3 m asl, which is the lowest possible depth without burning out the pump.

In 2022/23, groundwater levels in P1 were approximately 147 m asl and remained consistent for most of the monitoring year. However, in November 2023, groundwater levels began to rise and have continued to increase, reaching a peak of approximately 150.6 m asl in April 2023 where the data set ends. The increase in groundwater levels in P1 likely reflects a decline in pump performance or increased infiltration following prolonged heavy precipitation events. In response to the observed elevation in water levels at P1, the pump has undergone maintenance and is now fully operational. Water levels have subsequently been restored to 147.3 m asl. Water levels in the adjacent purge well P4 remained below 136 m as throughout the 2022/23 monitoring year.

In 2022/23, a total of $30,580 \mathrm{~m}^{3}$ of leachate was collected from the South Purge Wells, approximately $13.3 \%$ less than in the previous monitoring year. The average daily flow in 2022/23 was $83.8 \mathrm{~m}^{3} /$ day. Overall, consistent leachate discharge volumes and groundwater levels observed in the South Purge Wells suggest the purge well system functioned effectively in 2022/23.

### 4.4.4 North of Phase 2

A detailed assessment of hydrogeologic conditions below the North Ridge was conducted in 2016, culminating in a hydrogeological conceptual model of the ridge (AECOM 2016). The report stated that subvertical strike-slip faults (e.g., the Highland Fault) near groundwater monitoring stations 87 and 88 and an inferred fault near station 77 behave as barriers to west-east groundwater flow, creating a compartmentalized groundwater flow system in the North Ridge area. Conversely, subhorizontal tensile fractures behave as preferential conduits for groundwater flow. It is likely that bedrock discontinuities contribute to the large fluctuations in groundwater elevations beneath the North Ridge in response to seasonal precipitation and infiltration events. Geologic mapping of exposed bedrock on the North Ridge (AECOM 2018a) revealed an undulating bedrock surface with several closed depressions that allow for surface water pooling and enhanced recharge during wet weather, which is important for maintaining the Phase 2 hydraulic trap.

Figure 4-9 presents groundwater elevations in North Ridge area. The location of the groundwater divergence below the North Ridge has important implications for maintaining the hydraulic trap as the landfill is expanding northward. Since 2006, ten (10) new wells have been installed at five separate locations (groundwater monitoring stations 77, 78, 79, 87, and 88) north of the High Level Road to investigate the groundwater divergence and direction of groundwater flow north of Phase 2. Continuous water levels have been recorded at monitoring locations GW-77-1-1, GW-78-1-1, GW-87-1-1, GW-87-2-1, GW-88-1-1, and GW-88-2-1 to better understand the temporal and spatial variability in groundwater elevations over time. In November 2019, four new groundwater monitoring wells (GW-91-1-1, GW-92-1-1, GW-93-1-1, and GW-94-1-1) were installed in the North Ridge area to increase the spatial coverage of the groundwater monitoring network and investigate the groundwater divergence. Well GW-94-1-1 is located along the inferred groundwater divergence. In 2022/23, well GW-93-1-1 was damaged due to the blasting activities and subsequently decommissioned.

In 2022, a total of 17 monitoring wells were installed north of Phase 2. Wells GW-96-1-1, GW-97-1-1, GW-98-1-1, GW-99-1-1, GW-103-1-1, GW-104-1-1, GW-105-1-1, GW-106-1-1, GW-107-1-1, GW-108-1-1, GW-109-1-1 and GW-110-1-1 were installed in bedrock, whereas monitoring wells GW-95-1-1, GW-100-1-1, GW-101-1-1, GW-102-1-1 and GW-107-1-2 were installed in overburden or across the overburden-bedrock contact. Well GW-102-1-1 was decommissioned shortly after installation.

In 2022/23, blasting performed in the western portion of Phase 2 resulted in substantive lowering of the bedrock surface of quarry and the Toutle Valley. Groundwater elevations along the North Ridge continued to exhibit seasonal fluctuations, but their intensity was less prominent as seen in Figure 4-9. Historically, groundwater levels have been notably higher to the west of the Highland Fault compared to the immediate east, particularly during the winter. However, in 2022/23, this difference in groundwater levels became less distinct during the winter months. In the summer, measurements beneath the North Ridge (from wells GW-62-1-1, GW-87-1-1, and GW-88-1-1) depicted a moderately sloping piezometric surface ranging between 160 to 165 m asl-roughly 4 m lower than the previous year's readings. It is suspected that the quarry cut through the Highland Fault, potentially facilitating the drainage of eastward flowing groundwater that was previously impeded by the fault. This is consistent with the extensive groundwater seepage observed at the base of the Cell 4/5/6 quarries and generally lower groundwater elevations in the North Ridge area. Groundwater well GW-27-1-1 was an artesian well, but water levels decreased to $4-5 \mathrm{~m}$ below ground surface in 2022/23. Even with the weakened eastward groundwater hydraulic gradient, the Phase 2 hydraulic trap remained effective.

In 2022/23, upward groundwater gradients continued to present at monitoring stations 77, 87, 88, indicating that groundwater discharges to surface over the footprint of the bedrock quarry and within Toutle Valley. Seepage faces were also observed near the base of the quarry highwall. Well GW-27-1-2 was destroyed, so the hydraulic gradient could not be assessed in this location. However, the upward hydraulic gradients at well location 27 diminished during the October 2021 and February 2022 monitoring events, which may be related to active quarrying in the Toutle Valley. Due to ongoing quarry and blasting activities,
well GW-27-1-1 is no longer sustainable and needs to be decommissioned. The well should be decommissioned and sealed properly in accordance with the Groundwater Protection Regulation, B.C. Reg. 75/2021 (last updated March 11, 2021).

Given the importance of maintaining the groundwater divergence for leachate containment, future quarrying in the Toutle Valley and the North Ridge should continue to be conducted under the direction of a qualified blasting professional to minimize the potential for blast-enhanced fracturing, with possible negative impacts on hydraulic properties, groundwater elevations, groundwater flow rates, and leachate containment north of the Phase 2 landfill. Ultimately, if blasting programs are not properly designed and implemented, the integrity of the hydraulic trap may be compromised. In circumstances where blasting might induce substantial topographic alterations or changes to the Toutle Valley's elevation, consultation with a hydrogeologist is recommended. This is because the ground surface within the quarry serves as the primary control mechanism for groundwater tables. Any major modifications to the site's topography or valley floor elevation will directly influence the potentiometric structure of the hydraulic trap.

### 4.5 Summary

### 4.5.1 Leachate Flow

In 2022/23 leachate flow patterns at Hartland were consistent with historic interpretations. Leachate mounding persisted in Phase 1, and leachate elevations were generally stable, exhibiting minor seasonal variations. The leachate mound in the upper portion of the refuse is interpreted as being 'perched' above the regional bedrock groundwater flow system, with relatively high water levels and strong downward hydraulic gradients. Between 2016 and 2023, leachate elevations in the upper portion of the refuse were generally below 155 m asl, reflecting an approximate 5 m decrease in elevation since closure of the Phase 1 landfill in 1997. Based historical data and the 2022/23 leachate flow data, AECOM made the following interpretations:

- In April 2023, the pressure transducers in wells GW-89-1-1 and GW-89-2-1 (completed in Phase 1) were damaged, so the data recorded in 2022/23 could not be retrieved. Manual leachate elevation measurements from wells GW-89-1-1 and GW-89-2-1 were consistent with values observed last monitoring year.
- The highest leachate elevations ( 155 to 157 m asl) were typically observed in the east/southeast area of the Phase 1 (GW-46-2-1, VLGW-004D and VLGW-011S), an area with elevated topography and refuse heights.
- Historically, leachate elevations in Phase 2 were typically about 0.8 m below the leachate elevation in the Lower Leachate Lagoon. However, since November 2021, leachate elevations in Phase 2 have increased, and were above the elevations of the Lower Leachate Lagoon in the wet season. CRD confirmed that this trend was likely due to calibration/ instrument drift, and were not indicative of a change in operations.
- Leachate levels observed at monitoring stations GW-82-1-1 and GW-83-1-1 (completed in Phase 2) were consistent with historical values, and well GW-82-1-1 was dry in September and November 2022.
- In 2022/23, a total of $30,580 \mathrm{~m}^{3}$ of leachate was collected from the South Purge Wells, approximately $13.3 \%$ less than in the previous monitoring year. Standard leachate discharge volumes and consistent groundwater levels observed in the South Purge Wells suggest the purge well system functioned effectively in 2022/23.
- In 2022/23, a total of $14,277 \mathrm{~m}^{3}$ of leachate was collected from the North Purge Wells, approximately $21.3 \%$ lower than in the previous monitoring year. Water levels in GW-40-1-1, GW-52-4-0-P7, GW-80-1-0-P8, and GW-81-1-0-P9 were generally variable but consistent with historical ranges. Standard leachate discharge volumes and consistent groundwater levels observed in the North Purge Wells suggest the purge well system functioned effectively in 2022/23.


### 4.5.2 Groundwater Flow

In 2022/23, groundwater flow patterns observed at Hartland were consistent with historical interpretations, with some variability in the North Ridge area. Regional groundwater flows from Mount Work northeast to the north-south trending valley that underlies the northern portions of the Phase 1 and Phase 2 landfill. Most of the northward groundwater flow in the bedrock below the landfill is captured by the Toutle Valley Underdrain, Phase 2 Basin Leachate Collection System, springs discharging to the lower lagoon, and the north and south purge well systems (wells P1, P2, P3, P4, P7, P8, P9, and P10). Based on the 2022/23 groundwater flow data, AECOM made the following interpretations:

- Groundwater monitors east of Phase 1 (e.g., GW-54-1-1, GW-76-1-1, GW-17-1-1, and GW-18-1-1) confirmed eastwest flow toward the landfill, preventing off-site migration to the east.
- Groundwater levels in GW-52-3-0 increased substantially in 2021/22 and remained elevated during 2022/23. The increase in groundwater elevations at GW-52-3-0 may be related to overall lower leachate discharge rates in North Purge Wells or reduced well efficiency because of biofilm formation in wells GW-52-4-0-P7 or GW-80-1-0-P8. This should be investigated further to confirm the purge wells remain effective.
- Leachate indicator concentrations in wells near GW-81-1-0-P9 have stabilized or decreased slowly under the current pumping configuration/settings. The CRD may consider increasing the pumping rates or adjusting set points to achieve the groundwater elevations required to maintain pumping levels below the Lower Leachate Lagoon and collect more leachate migrating from the area around the landfill gas plant.
- Closure of Phase 2 Cell 1 and the application of tarps to restrict infiltration and leachate generation appears to be slowly improving leachate containment north of the landfill.
- A small amount of groundwater flows southeastward from the south end of Phase 1 in the direction of Killarney Lake. Southeastward groundwater flow below the landfill is constrained by a constructed clay berm and bedrock grout curtain installed at the south end of the landfill and by drawdown cones associated with the South Purge Wells. In 2022/23, water levels all south purge wells remained within their normal ranges, except for P1, P2 P10. Water levels in P1 gradually increased from November 2022 reaching a peak of approximately 150.6 m asl in March 2023. In response to the observed elevation in water levels at P1, the pump has undergone maintenance and is now fully operational. Water levels have subsequently been restored to 147.3 m asl.
- In 2022/23, quarry and blasting activities carried out west of Phase 2 resulted in substantive changes to the topography and lowering of the Toutle Valley ground surface. Subsequently, groundwater levels in several wells situated south of the Northwest Stockpile, notably wells GW-87-1-1 and GW-88-1-1, declined by approximately 4-10 m compared to previous years. This decline also led to diminished eastward hydraulic gradients. Seasonal fluctuations can still be observed along the Highland Faults, but they have become less pronounced. Despite this impact on the local groundwater flow pattern and hydraulic gradients, the Phase 2 hydraulic trap remained effective. However, groundwater elevations require close monitoring as they establish a new dynamic equilibrium to confirm the performance of the hydraulic trap and inform the design of the Phase 4/5/6 landfill and its associated underdrain system.






Figure 4-4. Groundwater Elevations East of Phase 1




Figure 4-5. Leachate and Groundwater Elevations Within Phase 1




Figure 4-7. Groundwater Elevations in the South Purge Wells


Figure 4-8. Water Elevations Within the Leachate Conveyance System and Surrounding the Phase 2 Basin




## 5. Groundwater Quality Monitoring Wells

### 5.1 Compliance Groundwater Monitoring Locations

A total of 36 compliance monitoring wells have been identified at 19 different locations at the Hartland Landfill. These stations are concentrated along the south, east and northern property boundaries and are located downgradient of areas that have the potential to be impacted by leachate or runoff from the site. The monitoring wells listed below are considered Boundary Compliance Wells.

South of the Landfill (10)

- 04-3-1, 04-4-1
- 71-1-1, 71-2-1, 71-3-1
- 72-1-1, 72-3-1
- 73-1-1, 73-2-1, 73-3-1

North of the Landfill (15)

- 20-1-1, 20-1-2
- 21-1-1, 21-1-2, 21-2-1
- 28-1-0
- 29-1-1, 29-1-2
- 30-1-1, 30-1-2
- 31-1-1, 31-1-2
- 39-1-1, 39-2-1
- $53-1-1$

East of the Landfill (6)

- 17-1-1, 17-1-2, 17-1-3
- 18-1-1, 18-2-1, 18-2-2

North of the Hartland North Pad (5)

- 41-1-1
- 42-1-1
- $55-1-1$
- $56-1-1$
- $57-1-1$

Compliance is assessed in Section 5.2. All data, including the applicable standards, are provided in Appendix B. Values that exceed CSR standards are noted with footnotes. Analytical results for groundwater samples collected from monitoring wells for the reporting period are presented in Appendix B.1. Table 5-1 presents a summary of the wells that exceeded CSR standards for one or more parameters in 2022/23. Most exceedances represent groundwater samples collected from leachate collection wells or near leachate collection infrastructure. Additionally, nitrate concentrations in several groundwater wells (e.g., GW-25-1-1, GW-16-1-1, and GW-104-1-1) located downgradient of aggregate stockpiles exceeded the CSR DW standard throughout the monitoring year.

A review of the sampling program was undertaken in early 2016, and recommendations for modifications to the number, location and sampling frequency of compliance monitoring locations were implemented beginning in the 2016/17 monitoring year. Groundwater quality data at compliance wells GW-04-2-1 and GW-72-2-1 has not been collected since 2016 due to the low recharge rate in the wells and these wells are no longer considered compliance locations. AECOM is in the process of reviewing and updating the sampling program to ensure it remains effective in monitoring the impacts from ongoing landfill operation, including activities such as aggregate production, stockpiling and other activities.

Based on the Stage 8 Amendments to the CSR, DW and AW standards for iron and manganese are no longer applied to municipal landfills.

Quarterly monitoring conducted by the CRD typically includes both compliance wells and other non-compliance locations that contribute to a fulsome understanding of landfill processes and the potential for environmental risks.

### 5.2 Assessment of Groundwater Quality Impacts

The primary causes for any groundwater quality degradation at the site include leachate, road salt and aggregate production, stockpiling or use for construction purposes. Professional judgement is used to differentiate between different contaminant sources (leachate, road salt and aggregates) and to assess the nature and degree of any impacts. The authors of this report are hydrogeologists and geochemists with considerable experience at other landfills in coastal regions of British Columbia. Groundwater quality may be judged to be impacted relative to background without exceeding regulatory criteria, and therefore compliant. If concentrations exceed CSR AW or DW standards for groundwater quality at the property boundary, standard protocols for notification of affected property owners should be followed.

Relative concentrations and patterns of conductivity, ammonia, chloride, sulphate, and nitrate are compared to background concentrations to differentiate between the site's typical contaminant sources as outlined below:

- Background conductivity is typically below $500 \mu \mathrm{~S} / \mathrm{cm}$ but has been observed in some background wells at concentrations up to $1,000 \mu \mathrm{~S} / \mathrm{cm}$ immediately after well installation or following prolonged dry periods. Background ammonia concentrations are typically below $0.1 \mathrm{mg} / \mathrm{L}$, but occasionally reach $0.5 \mathrm{mg} / \mathrm{L}$ downgradient of wetland areas. Background chloride concentrations are typically below $20 \mathrm{mg} / \mathrm{L}$. Background sulphate concentrations are typically below $50 \mathrm{mg} / \mathrm{L}$ but are regularly observed at concentrations up to $100 \mathrm{mg} / \mathrm{L}$ in wells screened within weathered bedrock and near geological alteration zones.
- Groundwater is considered to be impacted by leachate when conductivity concentrations are above $1,000 \mu \mathrm{~S} / \mathrm{cm}$, ammonia concentrations are above $1 \mathrm{mg} / \mathrm{L}$, and chloride concentrations are above $20 \mathrm{mg} / \mathrm{L}$. Peak concentrations in leachate impacted wells are typically observed during the dry summer and early fall months, when there is limited dilution by precipitation.
- Groundwater is considered impacted by aggregate (e.g., production, stockpiling or site construction works) when sulphate is present at concentrations above $75 \mathrm{mg} / \mathrm{L}$ and ammonia or nitrate are present at concentrations above background levels of $0.1 \mathrm{mg} / \mathrm{L}$. Peak concentrations are typically observed during the first sampling event following the onset of wet weather in the fall months after recent blasting and aggregate stockpiling.
- Groundwater is considered impacted by road salt when both conductivity ( $>1,000 \mu \mathrm{~S} / \mathrm{cm}$ ) and chloride ( $>20 \mathrm{mg} / \mathrm{L}$ ) are elevated above background levels, but ammonia and its degradation products (primarily nitrate) are not elevated. Chloride ( Cl ) to sodium ( Na ) molar ratios are also used to assess the source of potential chloride sources. Cl/Na molar ratios in road salt are generally one to one, assuming $100 \%$ compositional purity. In 2022/23, CI/Na molar ratios in the north and south purge wells except for P9 ranged from 0.38 to 0.75 , with a median ratio of 0.54 . Road salt impacted sites must also be located downgradient (or downstream) of surfaces where road salt is known to be applied. Concentrations of conductivity and chloride typically exhibit peaks following cold weather periods when de-icing salt is often applied to roadways.


### 5.3 Electrical Conductivity

Figure 5-1 presents the electrical conductivity values in plan for samples collected at Hartland Landfill in September 2022. Figure 5-2 and Figure 5-3 present north-south and east-west cross-sections through the landfill and north of the landfill that show conductivity values in September 2022. Electrical conductivity is a good indicator of the presence of inorganic parameters and a good indicator of potential leachate contamination when elevated ammonia and chloride are also present. The highest conductivity values are typically observed during the dry season (i.e., August/ September), and were utilized to interpret the conductivity contours.

On Figure $5-1$, the $1,000 \mu \mathrm{~S} / \mathrm{cm}$ conductivity contour line is interpreted as indicating the presence of leachate in groundwater. Figure $5-1$ shows that the $1,000 \mu \mathrm{~S} / \mathrm{cm}$ contour closely resembles the outline of current refuse disposal and indicates that groundwater in these areas has been affected by leachate.

Like previous years, conductivity values in north and south purge wells were generally above $1,000 \mu \mathrm{~S} / \mathrm{cm}$ in 2022/23, except for GW-81-1-0-P9. Conductivity values continuously fell below $1,000 \mu \mathrm{~S} / \mathrm{cm}$ for well GW-81-1-0 (P9) through 2022/23. As shown on Figure 5-1, conductivity values in the north purge wells were highest in P8, which was close to $4,069 \mu \mathrm{~S} / \mathrm{cm}$ in September 2022. In $2022 / 23$, conductivity values in the south purge wells ranged from 990 to $2,600 \mu \mathrm{~S} / \mathrm{cm}$. Like previous years, the $1,000 \mu \mathrm{~S} / \mathrm{cm}$ contour runs north of the north purge well system and Lower Leachate Lagoon, and extends south of $\mathrm{P} 1, \mathrm{P} 2, \mathrm{P} 3, \mathrm{P} 4$, and P 10 of the south purge well system. Conductivity values at new purge wells P 11 and P 12 were relatively lower, ranged from 645 to $1,169 \mu \mathrm{~S} / \mathrm{cm}$. P 12 was dry most of the time, and only one sample could be taken.

The $500 \mu \mathrm{~S} / \mathrm{cm}$ conductivity contour is considered indicative of background groundwater quality. In September 2022, the $500 \mu \mathrm{~S} / \mathrm{cm}$ conductivity contour along the northern boundary of the landfill expanded beneath the Upper Leachate Lagoon, as indicated by GW-38-1-1 and new wells GW-105-1-1 and GW-108-1-1, and slightly further north from the Lower Leachate Lagoon. Additionally, a 500 $\mu \mathrm{S} / \mathrm{cm}$ contour was inferred around GW-31-1-1, and in 2022/23 extended to well GW-30-1-1. The elevated electrical conductivity values observed in this region could potentially be linked to impacts from the upgradient stockpiled aggregate located to the southeast of this area, in the Triangle Stockpile and Southeast Stockpile. Sulphate concentrations in well GW-31-1-1 and GW-31-1-2 ranged from 185 to $270 \mathrm{mg} / \mathrm{L}$, which were well above the aggregate impact threshold of $75 \mathrm{mg} / \mathrm{L}$. Chloride and ammonia concentrations remained low, that water quality is unlikely to be impacted by road salt or leachate.

Like previous years, the $500 \mu \mathrm{~S} / \mathrm{cm}$ contour did not extend beyond location 25 at the north end of the landfill. Groundwater collected from location 36 showed that conductivity values during two of the four sampling events were above $1,000 \mu \mathrm{~S} / \mathrm{cm}$ and remained slightly elevated in 2022/23. At Hartland North Pad, elevated conductivity was also observed at wells GW-43-1-1 (509 $\mu \mathrm{S} / \mathrm{cm}$ ) and GW-91-1-1 ( $510 \mu \mathrm{~S} / \mathrm{cm}$ ) Along the east boundary of the landfill, the $500 \mu \mathrm{~S} / \mathrm{cm}$ contour did not extend beyond GW-18-1-1. At the west boundary, the location of the $500 \mu \mathrm{~S} / \mathrm{cm}$ contour could not be determined due to the absence of monitoring wells on the steep slope. However, the presence of conductivity data for GW-27-1-1, coupled with eastward and upward groundwater flow in that area indicates the $500 \mu \mathrm{~S} / \mathrm{cm}$ contour remains east of GW-27-1-1. South of the landfill, the $500 \mu \mathrm{~S} / \mathrm{cm}$ contour did not extend to GW-07-1-1 but does encapsulate well GW-72-1-1.

Source control techniques should continue to be implemented throughout the landfill including minimizing the volume of aggregate stockpiles, application of covers and paving of traffic surfaces to maintain the quality of surface and groundwater in the future.

### 5.4 Overview of Groundwater Quality Exceedances

Groundwater quality data were compared to applicable BC CSR AW and DW standards, as shown in Appendix B-1. A summary of groundwater quality exceedances is presented in Table 5-1. In 2022/23, Boundary Compliance Wells and off-site monitoring wells met CSR AW and DW standards, except for one copper concentration exceedance ( $41.5 \mu \mathrm{~g} / \mathrm{L}$ ) observed at GW-21-1-1 during the May 2022 sampling event. However, the reported elevated copper concentration was derived from the mean of the parent and duplicate results (Table 3-1). While one result was below the detection limit of $0.25 \mu \mathrm{~g} / \mathrm{L}$, the other was significantly higher at $82.8 \mu \mathrm{~g} / \mathrm{L}$. This disagreement in measurements suggests the possibility of cross-contamination either during sampling or in post-sampling procedures. Therefore, it's likely that the recorded exceedance may not accurately represent true conditions.

Similar to previous years, most exceedances were present in groundwater wells in close proximity to leachate purge wells and known leachate sources. However, nitrate concentrations in several groundwater wells (e.g., GW-16-1-1, GW-25-1-1, GW-104-1-1, GW-105-1-1, GW-106-1-1, GW-107-1-1) located downgradient of aggregate stockpiles exceeded applicable CSR DW standards on one or more sampling event. In all wells except GW-106-1-1, ammonia concentrations remained below the detection limit. The nitrate exceedances are attributed to aggregate production and stockpiling at the landfill.

### 5.5 Monitoring Sites North of the Phase 1 Landfill

Figure 5-4 illustrates long-term groundwater quality trends north of the Phase 1 landfill.

### 5.5.1 Monitoring Site 58

Monitoring site 58 (not a Boundary Compliance Well) is located in the transition area between the Phase 1 and Phase 2 landfill 200 m from the property boundary. GW-58-1-0 is 19 m deep and screened in bedrock below refuse. Well GW-58-1-0 has exhibited elevated concentrations of leachate indicator parameters since 2001, shortly after landfilling upslope of the well began. Although solute concentrations have generally decreased since peak concentrations were observed from 2001 to 2004, concentrations of key leachate indicator parameters (e.g., conductivity, ammonia, chloride, and nitrate) remain highly elevated. Water quality is indicative of leachate-impacted groundwater flowing toward the Lower Leachate Lagoon and the north purge wells.

In 2022/23, four groundwater sampling events were conducted at location 58 . Similar to previous years, all four samples from well GW-58-1-0 exhibited concentrations of ammonia and cobalt that exceeded CSR AW standards. Furthermore, concentrations of chloride, cobalt, nickel, strontium, and vanadium exceeded CSR DW standards during one or more sampling events. Sodium concentrations were above the CSR DW of $200 \mathrm{mg} / \mathrm{L}$ but well below the regional background concentration of $1,700 \mathrm{mg} / \mathrm{L}$ and therefore below CSR standards. In 2022/23, mean concentrations of ammonia and chloride were $105.8 \mathrm{mg} / \mathrm{L}$ and $988 \mathrm{mg} / \mathrm{L}$, respectively, which are consistent with 2021/22 concentrations. Groundwater quality in GW-58-1-0 was generally stable with respect to all leachate indicator parameters other than conductivity. Since Q1 2020, sulphate concentrations have progressively increased from $60 \mathrm{mg} / \mathrm{L}$ to $120 \mathrm{mg} / \mathrm{L}$ and may reflect the influence of nearby aggregate storage and use to support landfill operations and closure.

Similar to previous years, dissolved cobalt concentrations ranged from 50.1 to $55.7 \mu \mathrm{~g} / \mathrm{L}$, exceeding both the CSR AW and regional background concentration of $14 \mu \mathrm{~g} / \mathrm{L}$. Continued water quality monitoring at this well over time is important as it provides an indication of the quality of leachate migrating to the Lower Leachate Lagoon and the north purge well system.

### 5.5.2 Monitoring Sites 52 (P7), 80 (P8) and 81 (P9)

Monitoring sites 52, 80 and 81 (not Boundary Compliance Wells) located in the centre of the leachate plume between the toe of the Phase 1 landfill and the lower leachate lagoon, approximately 100 m from the property boundary. Leachate flows along the bedrock/refuse interface underlying Phase 1 toward the lower leachate lagoon and the north purge wells (GW-52-4-0-P7, GW-80-1-0-P8 and GW-81-1-0-P9). Groundwater quality has been impacted heavily by leachate since monitoring began at these locations in the 1980's.

In 2022/23, groundwater samples were collected from well GW-52-1-1 (31 m deep) and leachate samples were collected from the north purge wells (GW-52-4-0-P7, GW-80-1-0-P8, and GW-81-1-0-P9). Prior to 2016/17, leachate samples had only been collected from the combined discharge at GW-52-4-0-P7 and GW-80-1-0-P8.

Concentrations of chloride in monitoring well GW-52-1-1 met the CSR DW standard of $250 \mathrm{mg} / \mathrm{L}$ during all sampling events in 2022/23. Ammonia concentrations were similar to previously reported values, exceeding the CSR AW standards on all sampling dates. Similarly, strontium concentrations exceeded the CSR DW standard on all sampling dates. In 2022/23, the mean annual conductivity in GW-52-1-1 was $1,785 \mu \mathrm{~S} / \mathrm{cm}$, which was about $7 \%$ lower than the mean value observed in 2021/22 (1,926 $\mu \mathrm{S} / \mathrm{cm}$ ). No statistically significant trend was observed in GW-52-1-1 over the past five years.

In 2022/23, conductivity values in GW-52-4-0-P7 and GW-81-1-0-P9 were consistent with those observed in 2021/22. However, in GW-80-1-0-P8, conductivity increased by approximately $15 \%$. Unlike previous years, leachate in P8 was even more concentrated than P7, with conductivity ranging from 3,061 to $6,600 \mu \mathrm{~S} / \mathrm{cm}$. Leachate quality in P8 exhibited the most significant degradation, and ammonia, conductivity, and chloride concentrations were generally 10 to $20 \%$ higher than those measured in 2021/22, which may reflect reduced infiltration following closure of Phase 2 Cell 1, and prolonged dry weather.

Statistically significant increasing trends in conductivity, chloride and ammonia concentrations were observed in GW-80-1-0P8, and a statistically significant increasing trend in conductivity was observed at GW-52-4-0-P7. Groundwater quality in GW-52-4-0-P7 exceeded CSR AW standards for ammonia on all sampling events and the CSR DW standard for chloride during three sampling events. Groundwater quality in GW-80-1-0-P8 exceeded the CSR AW standards for ammonia and chromium on all four sampling events. Moreover, arsenic, chloride, cobalt, and vanadium concentrations in GW-80-1-0-P8 exceeded the CSR DW standards on all sampling events. More frequent CSR exceedances and increased concentrations of leachate indicator parameters observed at GW-80-1-0-P8 indicate that leachate quality has continued to deteriorate throughout 2022/23. Leachate quality in GW-81-1-0-P9 has generally been stable over time, exhibiting statistically significant decreasing trends in nitrate and chloride concentrations. Although, in 2022/23, strontium concentrations observed at GW-81-1-0-P9 exceeded the CSR DW standard on one sampling event.

Overall, groundwater quality in and GW-80-1-0-P8 continued to degrade throughout 2022/23, while leachate indicator concentrations remained stable or decreased slightly in GW-52-4-0-P7 and GW-81-1-0-P9. Well GW-80-1-0-P8 exhibited the most substantive evidence of groundwater quality degradation, supported by an increased number of CSR exceedances and higher conductivity, chloride, and ammonia concentrations. Water quality at these wells should continue to be monitored to verify the effectiveness of the leachate collection system, assess temporal changes in leachate quality in response to seasonal changes in precipitation, and assess water quality impacts associated with aggregate production and stockpiling.


| ${ }^{\text {bc css }}$ |  |  |  | AW Mxixum (1) |  |  | ${ }^{9}$ | 50 | 10000 | ${ }^{2(9)}$ |  | 12000 | $0.54(5)$ |  | 1500 | $90(7)$ | 40 | ${ }^{20.900(5)}$ |  |  | ${ }^{40-160(5)}$ |  |  |  | 10000 | ${ }^{13111844(3)}$ |
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|  |  |  |  | ow Mxximum (2) |  | 9500 | 6 | 10 | 1000 | 8 |  | 5000 | 5 |  | 250 | 6000 | 1499 | 1500 |  |  | 10 | ${ }^{33} 9$ |  |  | 250 |  |
| Station | Sample | complane | Date Sampled | Paramear |  | Aumium | Anainony | Assonic | Batum | Berlium | Bismuth | Baon | ${ }_{\text {cammum }}$ | Cacaum | Choricie | chomium | Cobat | Coperer | ${ }_{\text {Hardioss }}^{\text {Cas }}$ (asas) | lon | Lead | Lithium | Magnosium | Manganse | Morobesum | Ammona |
|  |  |  |  | Fraction | тот | 115 | is | 015 | dis | 015 | 115 | dis | 015 | 015 | 015 | 015 | dis | 015 | 015 | 015 | 015 | ois | 015 | 015 | ois | ois |
|  |  |  |  | Mentrod dotection Limit Unitul | mal | ${ }_{\text {gel }}^{\text {g. }}$ | ¢0. | $\underbrace{}_{\substack{\text { gel } \\ 0.02}}$ |  |  | $\underbrace{\text { den }}_{\substack{\text { pel } \\ 0.005}}$ | ¢90ヶ |  | $\underbrace{}_{\substack{\text { mgl } \\ 0.05}}$ | m9 ${ }^{\text {mal }}$ | ${ }_{\text {gol }}^{0.1}$ | ${ }_{\text {che }}^{\substack{\text { gen } \\ \text { 0.05 }}}$ | $\underbrace{\substack{\text { g. } \\ 0.05}}_{\text {gen }}$ | ${ }_{\text {mal }}^{\text {m }}$ | ${ }_{\text {¢9 }}$ | $\xrightarrow{\text { pel }}$ | ${ }_{\text {gel }}^{0.5}$ | $\underbrace{}_{\substack{\text { mal } \\ 0.05}}$ |  |  | $\underbrace{}_{\substack{\text { mal } \\ 0.015}}$ |
| $8{ }^{81-1.00(99)}$ | ss |  | 2022.10 .05 |  | 23. | ${ }^{7} .83$ | 0.041 | 0.123 | ${ }^{37.9}$ | 0.01 | 0.005 | 707. | 0.005 | ${ }^{79.7}$ | ${ }^{92}$ | 0.55 | 0.663 | 0.089 | 32. | 79.1 | $0^{0.0113}$ | 0.81 | 30.4 | 14. | 0.44 | 5.9 |
| P10 | ss |  | ${ }^{20220.5} 5$ | Stighy sily and sighly roarge | 560. | ${ }^{236}$ | ${ }^{0.036}$ | 0.855 | 74. | 0.01 | ${ }^{0.005}$ | ${ }^{1340}$ | ${ }^{0.0183}$ | ${ }^{99} 3$ | 100. | 0.54 | 1.36 | 0.326 | ${ }^{32}$. | 504. | 0.012 | ${ }^{324}$ | 19.8 | 557. | 4.68 | ${ }^{33}$ |
| P10 | ss |  | ${ }^{2022-20.14}$ |  | ${ }^{660}$. | 22. | ${ }^{0.1}$ | ${ }^{0.94}$ | ${ }^{221 .}$ | ${ }^{0.05}$ | 0.025 | ${ }^{1990}$ | 0.025 | ${ }^{91.8}$ | ${ }^{140}$ | 0.92 | ${ }^{232}$ | 1.91 | 32. | ${ }^{609}$ | 0.223 | ${ }^{4.3}$ | 22. | 50. | ${ }^{7} 06$ | 47. |
| P10 | ss |  | ${ }^{2022-12.01}$ | Clears, sighly yelow | 710. | ${ }^{6.4}$ | 0.051 | 0.979 | 921. | 0.02 | 0.01 | 1790. | 0.01 | ${ }^{93.1}$ | ${ }^{140}$ | 0.77 | 2.56 | 0.47 | ${ }^{323}$ | 56. | 0.072 | 4.9 | 21.9 | 510. | 6.75 | 54. |
| P10 | ss |  | ${ }^{2023} 22.24$ | Vens signt yelelu, no utubiliy | 550. | 4.15 | 0.034 | ${ }^{1.12}$ | ${ }^{226 .}$ | 0.01 | 0.005 | 1130. | 0.005 | ${ }^{858}$ | 54. | ${ }^{0.54}$ | ${ }_{1}^{1.18}$ | 0.841 | 28. | 498. | 0.056 | ${ }^{3.73}$ | 16.8 | 427. | 3.74 | ${ }^{33}$ |
| P11 | ss |  | 20220.901 | Moderaey stiv, vee gey | 110. | ${ }^{3.54}$ | 1.71 | ${ }^{0.342}$ | ${ }^{25,3}$ | 0.01 | ${ }^{0.005}$ | 217. | 0.0135 | ${ }^{131 .}$ | ${ }^{9} 9$ | ${ }^{0.1}$ | ${ }^{0.369}$ | 1.03 | ${ }^{493}$ | ${ }^{5} .1$ | 0.0742 | ${ }^{0.72}$ | ${ }^{40.3}$ | ${ }^{828}$ | 4.43 | ${ }_{0}^{0.34}$ |
| P11 | ss |  | ${ }^{2022-1.129}$ | Clear and colouress | ${ }^{110}$ | 10.6 | 0.186 | 0.087 | ${ }^{16.5}$ | 0.01 | 0.005 | 187. | ${ }^{0.0387}$ | 114. | 5. | 0.19 | 248 | ${ }_{6} .93$ | ${ }^{341}$. | 21.7 | 0.0225 | 0.5 | ${ }^{13.7}$ | ${ }^{189}$ | ${ }_{1} 1.3$ | 7.9 |
| P11 | ss |  | 202303.08 | Claearand coumesess | ${ }^{88}$ | ${ }_{3} .67$ | 0.167 | 0.075 | ${ }^{8.13}$ | 0.01 | 0.005 | ${ }_{1} 35$. | 0.0321 | 155. | ${ }^{8}$ | 0.1 | 0.864 | 287 | 473. | 7.2 | 0.0084 | 0.5 | 21. | 60.2 | ${ }_{1} .32$ | 1.4 |




| $(1)$ |
| :--- |
| $(1)$ |
| $(3)$ |


${ }^{(4)}$



(7)





| ${ }^{\text {bc css }}$ |  |  |  | AW Maximum (1) | ${ }^{0.2244}$ |  | 400 |  | 400 |  | 250-1500 (5) |  |  | 20 |  | ${ }^{0.5 .155(5)}$ |  |  | 128429 (5) |  | 3 |  | 1000 | ${ }^{85}$ |  |  | 75.2000 (5) |  |  |  |  |
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|  |  |  |  | ow maximum (2) | 1 |  | 10 |  | 10 |  | 80 |  |  | 10 |  | 20 | 1700 (9) | 250 | 500 |  |  | 25 |  | ${ }^{20}$ | 20 |  | 3000 |  |  |  |  |
| Station | ${ }_{\substack{\text { Sapple } \\ \text { Type }}}^{\text {ded }}$ | Complane | Date Sampled | Paraneter | Netie | $\frac{\text { Nitate }}{\substack{\text { Dis }}}$ |  |  |  |  | $\begin{gathered} \text { Nickel } \\ \hline \text { DIS } \end{gathered}$ | Phosphous | $\begin{gathered} \hline \text { Potassium } \\ \hline \text { DIS } \end{gathered}$ | $\begin{array}{\|c} \hline \text { Seenium } \\ \hline \text { Dis } \\ \hline \end{array}$ | $\begin{gathered} \hline \text { silion } \\ \text { Dis } \\ \hline \end{gathered}$ | $\begin{array}{\|c\|} \hline \text { siver } \\ \text { ils } \\ \hline \end{array}$ | $\begin{gathered} \text { Sodium } \\ \text { Dis } \\ \hline \end{gathered}$ | Strontium | Suphate | Sulur |  |  | Thanum | Unaium | Vanadum |  | Zno | zrionium | pH | conoustiny | Temperatur |
|  |  |  |  | Fration | ${ }^{1 / 5}$ |  |  |  | Nitrite + Nitrate |  |  | Dis |  |  |  |  |  | DIS |  | dis |  |  | dis |  |  |  | $\begin{array}{r}\text { Dis } \\ \text { vg } \\ \hline\end{array}$ | $\begin{gathered} \hline \text { ois } \\ \hline \text { pol } \\ \hline 01 \\ \hline 01 \end{gathered}$ | Tor <br> ¢ ${ }_{0}$ <br> 0 | ¢от | тот |
|  |  |  |  | Unit | ${ }_{\text {mon }}^{\substack{\text { mon } \\ 0.005}}$ | ${ }_{\substack{\text { mgl } \\ 0.02}}$ |  |  | $\mathrm{c}_{\text {mol }}^{\text {mod }}$ |  | ${ }_{\substack{\text { pol } \\ 0.02}}$ | $\stackrel{\mathrm{rg}}{2}$ | ${ }_{\substack{\text { mgL } \\ 0.05}}$ |  |  | ${ }_{\text {reg }}^{\substack{\text { ren } \\ 0.005}}$ | $\underbrace{}_{\substack{\text { m9, } \\ 0.05}}$ |  | mal | ${ }_{\text {gen }}$ | $\underbrace{\text { g.0. }}_{\text {gel }}$ | \%al | ${ }_{0}^{\mathrm{ng}} 0$ |  |  |  | ${ }^{\circ} \mathrm{C}$ |  |  |  |
| 116.12 | ss |  | ${ }^{2023.03 .15}$ | White, Mightubidity | 0.005 |  | 17.9 | ${ }^{-}$ | 17.9 | ${ }^{-}$ | ${ }_{0}^{0.382}$ | 3.2 | 0.241 | 0219 | 6540. | 0.005 | 5. | 261. | ${ }^{71}$ | ${ }^{227}$ | 0.0041 | 0.2 | 0.5 | 0.124 | 0.5 |  |  | 3.9 | 0.1 | ${ }_{6.85}$ | ${ }^{376}$. | 11.2 |
| 16.2 .1 | ss |  | 20230316 | diearand columess | ${ }_{0}^{0.0212}$ |  | 14.5 | b | 14.5 | b | ${ }^{0.561}$ | ${ }_{5}{ }^{2}$ | 0.351 | 0.176 | 6790. | 0.005 | 5.88 | ${ }^{323}$ | ${ }_{5} 5$. | 19.5 | ${ }_{0}^{0.0033}$ | 0.2 | ${ }^{1.19}$ | 0.109 | 1.07 |  | ${ }^{1.69}$ | ${ }^{0.1}$ | 6.84 | 366. | 10.7 |
| 16.2 .2 | ss |  | 202303316 | dearand colurusess | 0.0123 |  | ${ }^{17.3}$ | ${ }^{\circ}$ | 17.4 | b | 1.04 | 2. | ${ }^{0.3}$ | 0.19 | 6770. | 0.005 | ${ }^{5.3}$ | 27. | ${ }^{63}$ | 22.1 | 0.0034 | 0.2 | 0.5 | 0.13 | ${ }^{0.57}$ |  | 204 | 0.1 | 6.72 | ${ }^{389}$ | ${ }^{11}$ |
| $2{ }^{21-1-1}$ | FR1 | $r$ | 20220520 |  | 0.005 |  | 0.02 |  | 0.02 |  | 0.1 | 10. | 0.25 | 0.2 | 19500. | 0.025 | ${ }^{8.31}$ | ${ }_{425}$ | 12. | ${ }^{15}$ | 0.01 | - | 25 | 0.021 | 1. |  | ${ }^{19.8}$ | 0.5 | 8.32 | ${ }^{127 .}$ | ${ }^{11.4}$ |
| 25.1 .1 | ss |  | ${ }^{2022.05417}$ | Claerand colouless | 0.013 |  | 12.9 | $\checkmark$ | 12.9 | ${ }^{5}$ | 0.128 | 3.7 | 0.298 | 0.935 | 10000. | 0.005 | 4.51 | ${ }^{358}$. | ${ }^{190}$ | ${ }_{53} 5$ | 0.002 | 0.2 | 0.5 | 0.248 | 0.44 |  | 0.39 | 0.1 | ${ }^{7} 45$ | ${ }_{505 .}$ | ${ }^{11.4}$ |
| $5{ }^{52 \cdot 1.1}$ | ss |  | 202.054 .18 | Clear and sisighly freen | 0.005 |  | 0.02 |  | 0.02 |  | ${ }^{12.1}$ | 49. | ${ }^{13.7}$ | ${ }^{0.2}$ | 40900. | 0.025 | ${ }^{229}$ | 2900. | - 1 | 15. | 0.01 | - | 2.5 | 0.01 | - |  | 1.98 | ${ }^{0.5}$ | ${ }_{6.98}$ | 1688. | ${ }^{13.8}$ |
| $5{ }^{52 \cdot 1.1}$ | ss |  | 20220920 | Clear and colouress. | 0.005 |  | 0.02 |  | 0.02 |  | ${ }^{11}$ | ${ }^{30.3}$ | 14.9 | 0.127 | 38700. | 0.01 | 24. | 3490. | - 1 | 6. | 0.004 | 0.4 | 1 | 0.004 | 0.77 |  | ${ }^{124}$ | 0.2 | ${ }_{6.57}$ | 1799. | ${ }^{153}$ |
| 52-1.1 | ss |  | ${ }^{2022.212 .16}$ | Claearand columess | 0.005 |  | 0.02 |  | 0.02 |  | 12.7 | ${ }^{37}$ | ${ }^{14.8}$ | 0.2 | 40600. | 0.025 | 242 | 3360. | - 1. | 15. | 0.01 | - 1. | 2.5 | 0.01 | 1. |  | 1.72 | 0.5 | ${ }^{7} 2$ | 1592. | ${ }^{124}$ |
| $5{ }^{52-1 / 1}$ | ss |  | 2023.03 .02 | Claearand columess | 0.005 |  | 0.02 |  | 0.02 |  | 10.6 | 25.5 | ${ }^{13,7}$ | 0.114 | 37400. | 0.01 | 219. | 3310. | - 1. | 6. | 0.004 | 0.4 | 1. | 0.004 | 0.79 |  | 0.24 | 0.2 | 7.25 | 1680. | ${ }^{13}$ |
| 58.10 | ss |  | 20220517 | Ciearand modeatey y yelow | 0.072 |  | 0.2 |  | 0.2 |  | ${ }^{84}$ | ${ }_{664}$ | 56.5 | 0.372 | 19100. | 0.01 | 513. | 4230. | 110. | 38.9 | 0.0061 | ${ }^{0.67}$ | 2.5 | 0.401 | 20.8 | $\bigcirc$ | ${ }^{328}$ | 223 | ${ }_{6.68}$ | 5174. | 20. |
| 58-10 | FR1 |  | 2020.0922 | Cliar. vey orange | 0.179 |  | 0.457 |  | 0.636 |  | 86.7 | ${ }^{2}$ | ${ }_{53,5}$ | 0.46 | 18100. | 0.025 | 52. | 3890. | 99. | ${ }^{28}$ | 0.01 | 1. | 2.5 | 0.388 | 20.9 | - | 523 | 228 | 6.64 | 5699. | 19.3 |
| 58.10 | FR2 |  | 20220922 | Clearand colouress. | 0.218 |  | 0.41 |  | 0.628 |  | ${ }^{86,8}$ | ${ }^{2} 2$ | ${ }_{54,2}$ | 0.44 | 18300. | 0.025 | 52. | 3900. | 100. | 29. | 0.01 | - 1. | ${ }^{2.5}$ | 0.375 | 20.7 | - | $5^{5.1}$ | 2.33 | ${ }^{6.64}$ | 5699. | ${ }^{193}$ |
| 58, 10 | ss |  | 2022-12:13 | Claar, very yelow | 0.0632 |  | ${ }^{123}$ |  | 1.29 |  | ${ }^{826}$ | ${ }^{83,1}$ | ${ }_{63} 6$ | ${ }^{0.437}$ | 2200. | 0.01 | 564. | 4500. | ${ }^{20}$. | 31.7 | 0.004 | 0.74 | 3.5 | 0.335 | 30. | - | 4.57 | 293 | ${ }_{6}^{6.8}$ | 5164. | 16.6 |
| 58.10 | ss |  | 2023.03.09 | vees sighty yelow. .nilubidily | 0.05 |  | ${ }_{5} .81$ |  | 5.81 |  | 824 | ${ }^{86}$ | ${ }_{55,6}$ | 0.4 | 1920. | 0.025 | 511. | 3680. | ${ }^{92}$ | ${ }^{30}$ | 0.05 | , | 4.9 | 0.415 | ${ }^{26}$ | b | 5.2 | 2.58 | ${ }_{6}^{68}$ | 5841. | 19.6 |
| 559.1 | ss |  | 202209:15 | very untid, veng gey | 0.005 |  | 0.02 |  | 0.02 |  | 46.4 | 1160. | 1.05 | 0.437 | 32100. | 0.356 | ${ }_{6.35}$ | ${ }^{196 .}$ | ${ }^{36}$ | ${ }^{23 .}$ | 0.124 | 0.57 | ${ }_{453 .}$ | 1.16 | ${ }^{103 .}$ | b | 104. | 1.4 | 7.12 | ${ }^{395}$ | ${ }^{13,3}$ |
| 1041-1 | ss |  | 2022.1129 | vere utbid. very gey | 0.005 |  | 18.4 | $\checkmark$ | 18.4 | b | 0.52 | ${ }^{4.2}$ | 0.325 | 1.02 | 5390. | 0.005 | 4.47 | 27. | ${ }^{140}$ | 44.9 | 0.0024 | 0.2 | 0.5 | 0.204 | 0.84 |  | 1.04 | ${ }^{0.1}$ | 7.52 | 437. | ${ }^{11.8}$ |
| $1041-1$ | ss |  | ${ }^{202303316}$ | SIIghyy unbic. coourress | 0.005 |  | 21.5 | - | 21.5 | - | 0.58 | 27 | 0.512 | 0.911 | 5370. | 0.005 | 36.6 | 345. | ${ }^{20}$. | 43.5 | ${ }_{0}^{0.0086}$ | 0.2 | 0.61 | 0.597 | ${ }_{0}^{0.84}$ |  | ${ }_{1}^{1.38}$ | 0.1 | ${ }_{7} 7.6$ | 713. | 9.9 |
| $105 \cdot 1 / 1$ | ss |  | 20220.0.01 | Sighly wubid, siflily ysey | ${ }^{0.0938}$ |  | ${ }^{228}$ | , | 22.9 | - | 0.881 | 4.9 | 1.54 | ${ }^{1.36}$ | 5600. | 0.005 | 5.44 | 254. | 29. | ${ }^{77}$ | 0.0112 | ${ }^{0.38}$ | 0.5 | 0.334 | 0.35 |  | ${ }^{0.36}$ | 0.1 | ${ }_{6.61}$ | ${ }^{695}$ | ${ }^{13.5}$ |
| 105-1-1 | ss |  | $2022 \cdot 112$ | Modeataey wubdi, stinly fey | ${ }_{0}^{0.0232}$ |  | ${ }^{11.2}$ | , | 11.2 | b | 0.366 | ${ }_{6} 6$ | ${ }_{1}^{1.34}$ | ${ }^{1.26}$ | 5880. | 0.005 | 4.36 | ${ }^{184 .}$ | ${ }^{180}$. | 55.2 | 0.0065 | 0.2 | 0.5 | 0.46 | 0.41 |  | 0.28 | 0.1 | ${ }^{7}$ 7.62 | 507. | ${ }^{11.8}$ |
| 105.1.1 | ss |  | 2023.03.17 | Low utbidity, brown in couur low inens | ${ }^{0.0526}$ |  | 32. | b | ${ }^{32.1}$ | - | 0.416 | ${ }^{3} 7$ | 1.52 | 1.49 | 5900. | 0.005 | ${ }^{8.56}$ | 367. | 29. | ${ }^{92} 7$ | 0.008 | 0.2 | 0.5 | 0.602 | 0.44 |  | 1.16 | 0.1 | 7.63 | ${ }^{930}$. | 11.5 |
| 108-1-1 | ss |  | 20220.0.01 | Moderatey utid. moderaley yey | 1.92 | ${ }^{-}$ | 10.8 | b | ${ }^{12.8}$ | - | 19.9 | 30.9 | 21.5 | 0.336 | 6340. | 0.005 | ${ }^{74.8}$ | 60. | 250. | 74.2 | 0.0215 | ${ }^{0.93}$ | 2.12 | 1.53 | 1.22 |  | 5.91 | 0.44 | 7.43 | 2162. | 20.9 |
| 108.1.1 | ss |  | 2023.0331 | Extemeny utidi, exteney yrey | 0.097 |  | 20.8 | b | 20.9 | - | 4.09 | 76.9 | 245 | 0.107 | 8180. | 0.0061 | ${ }_{6} 6$. | 435. | 260. | ${ }^{93} 4$ | 0.002 | 0.2 | 322 | 0.564 | 925 |  | ${ }_{6} 6$ | 0.19 | 7.56 | 2200. | ${ }^{14.1}$ |
| $107.1 / 1$ | ss |  | 20220.0.01 | Vey ututi, sigity fey | 0288 |  | 23.9 | - | 24.2 | - | ${ }^{361}$ | 7.9 | 1.21 | 0.225 | 5880. | 0.005 | ${ }^{10.1}$ | ${ }^{390}$ | ${ }^{270}$. | 85.4 | 0.0118 | 0.2 | ${ }_{0} 0.95$ | 0.887 | 0.59 |  | 0.93 | 0.1 | ${ }^{7} 7.5$ | 1083. | ${ }^{21 .}$ |
| 109.1 .1 | ss |  | 2020.0.0.01 | vey utubi, sighly sey | 0.0223 |  | 0.044 |  | 0.066 |  | 0.889 | ${ }^{322}$ | ${ }^{1.24}$ | ${ }^{247}$ | ${ }^{11900}$ | 0.005 | ${ }^{85} 9$ | ${ }^{83} 2$ | ${ }^{76}$ | ${ }^{24.8}$ | ${ }_{0}^{0.025}$ | ${ }^{0.2}$ | 0.5 | 1.52 | 5.04 |  | 1.29 | 0.1 | ${ }^{8,74}$ | 460. | 21.9 |
| $109+1$ | ss |  | 2022.1 .130 | Modeatey utbid., modealey yey | 0.024 |  | 0.02 |  | 0.053 |  | 0.439 | ${ }_{35,3}$ | 1.11 | 1.06 | 13300. | 0.005 | 101. | 109 | ${ }^{\text {89. }}$ | 27.6 | 0.0039 | 0.2 | 0.5 | 236 | 1.52 |  | ${ }_{0} .31$ | 0.1 | 7.98 | 387. | ${ }^{14.6}$ |
| P1 | ss |  | ${ }^{2022.0526}$ | Claer and colouless | 0.005 |  | 0.02 |  | 0.02 |  | ${ }^{7} 33$ | ${ }^{86,9}$ | ${ }^{37,6}$ | 0.08 | 19800 | 0.01 | ${ }^{120 .}$ | ${ }^{957}$ | - | 6. | 0.004 | ${ }^{0.43}$ | 1. | 0.024 | 0.84 |  | 7.22 | 0.2 | 7.09 | 1261. | ${ }^{16 .}$ |
| ${ }^{9}$ | ss |  | 20220914 |  | 0.005 |  | 0.02 |  | 0.02 |  | 8.45 | 164. | 58.5 | 0.2 | 21100. | 0.025 | 27. | 1080. | 1. | 15. | 0.01 | 1. | 2.5 | 0.01 | 1. |  | 206 | 0.5 | 7.11 | 1913. | 18.3 |
| P1 | ss |  | 202.21 .201 | Claerand colouless | 0.0071 |  | 0.029 |  | 0.036 |  | 146 | ${ }^{85}$ | ${ }_{532}$ | 0.2 | 2000. | 0.025 | 187. | 1000. | 4.1 | 15. | 0.01 | 1. | 25 | 0.033 | 1. |  | 50.9 | 0.5 | 7.1 | 1730. | 16.5 |
| P1 | ss |  | 20230224 | dearand coluritess | 0.005 |  | 0.02 |  | 0.02 |  | ${ }_{7} 72$ | 46.8 | 20.4 | 0.08 | 18200. | 0.01 | 44. | 1060. | 1.2 | 6. | 0.004 | 0.4 | 1. | 0.029 | 0.4 |  | ${ }^{30.8}$ | 0.2 | 7.07 | 88. | 16.1 |
| ${ }^{\text {P2 }}$ | ss |  | ${ }^{2022.0526}$ | Cliar and Sisinhy yelow | 0.05 |  | 0.29 |  | 0.29 |  | 9.45 | ${ }_{35}$ | ${ }^{65} 2$ | 0.2 | 11800. | 0.025 | ${ }^{225 .}$ | ${ }^{222}$ | 1. | 15. | 0.01 | - 1. | 2.5 | 0.018 | 22 |  | ${ }^{1.88}$ | 0.5 | 7.22 | 1924. | 16.9 |
| ${ }^{2}$ | ss |  | 20220914 |  | 0.0159 |  | 0.069 |  | 0.085 |  | ${ }_{9} 98$ | ${ }_{5}^{53}$ | 76. | 0.2 | 13000. | 0.025 | 27. | 1210. | 1 | 15. | 0.01 | 1. | 2.5 | 0.017 | 2.4 |  | 5.88 | 0.5 | 7.35 | 2157. | 16.5 |
| ${ }^{\text {P2 }}$ | ss |  | 2022.12.01 | Sighaty ubides, sighly yelow | 0.0189 |  | 0.02 |  | 0.031 |  | 9.44 | ${ }^{22}$ | ${ }^{726}$ | 0.155 | 12100. | 0.005 | 256. | 1170. | 1. | 3. | 0.002 | 0.29 | 0.5 | 0.0185 | 278 |  | 292 | 0.39 | ${ }^{7} 24$ | 1927. | ${ }^{128}$ |
| P2 | ss |  | 202302224 | Vees silg yelolow, no utubily | 0.005 |  | 0.02 |  | 0.02 |  | ${ }^{7} 23$ | ${ }^{22.7}$ | 59.1 | 0.095 | 11900. | 0.01 | 199. | ${ }^{124}$ | 1. | 6. | 0.004 | 0.4 | 1. | 0.017 | 227 |  | 297 | 0.28 | ${ }^{7}{ }^{7}$ | 1742. | 15.5 |
| ${ }^{\text {P3 }}$ | ss |  | ${ }^{2022.0526}$ | Ciearand sighty orange | 0.05 |  | 0.2 |  | 0.2 |  | 10.9 | 41. | ${ }^{79,7}$ | 0.2 | 14700. | ${ }^{0.025}$ | 27. | ${ }^{937}$. | 1. | 15. | 0.01 | 1. | 2.5 | 0.013 | 2.9 |  | 225 | 0.5 | ${ }^{7} 22$ | 2197. | 18.3 |
| ${ }^{83}$ | ss |  | 20220914 |  | 0.14 |  | 0.034 |  | 0.174 |  | 9.41 | 44. | ${ }^{78,1}$ | 0.2 | 14000. | 0.025 | 286. | 1000. | - | 15. | 0.01 | 1. | 2.5 | 0.011 | 1.9 |  | 1.92 | 0.5 | 7.13 | 2226. | ${ }^{16}$ |
| ${ }^{2}$ | ss |  | $2022 \cdot 12.01$ | Slighty ubud, sighly orange | 0.0369 |  | 0.02 |  | 0.05 |  | 7.95 | 16.7 | ${ }^{75.6}$ | 0.16 | 13200. | 0.005 | 29. | 1010. | 2.3 | ${ }^{3}$ | 0.002 | 0.2 | 0.5 | ${ }^{0.0103}$ | 224 |  | 0.9 | 0.2 | 7.16 | 1625. | ${ }^{11.6}$ |
| ${ }^{\text {P }}$ | ss |  | 2023022.24 | Claer and colouress | 0.0196 |  | 0.025 |  | 0.045 |  | 7.16 | 26.3 | 69.9 | 0.06 | 15200. | 0.01 | ${ }^{223 .}$ | ${ }^{889}$ | 1. | 6. | 0.004 | 0.4 | 1. | 0.002 | 264 |  | 287 | 0.21 | ${ }_{7} 7.36$ | 1678. | 14.4 |
| ${ }^{84}$ | ss |  | ${ }^{20220525}$ | Cliar and moderatey orange | 0.335 |  | 0.95 |  | 1.89 |  | ${ }^{7} 77$ | ${ }^{24.7}$ | ${ }^{64,5}$ | 0.153 | 14100. | 0.01 | 22. | ${ }^{926 .}$ | 1 | 6. | 0.004 | 0.4 | 1. | ${ }^{0.0124}$ | 202 |  | 1.17 | 0.2 | 7.05 | 2004 | ${ }^{15.6}$ |
| P4 | ss |  | 202.09 .14 |  | 0.0143 |  | 0.033 |  | 0.097 |  | 8.1 | ${ }^{43}$. | 60. | 0.2 | 14300. | 0.025 | 224. | 847. | 1. | 15. | 0.01 | 1. | 2.5 | 0.01 | 1.9 |  | ${ }_{6.98}$ | 0.5 | 7.09 | 2011. | ${ }^{16 .}$ |
| ${ }^{4} 4$ | ss |  | $2022 \cdot 12.01$ | Sishaty wrubd, sighly orange | 0.0156 |  | 0.02 |  | 0.032 |  | 729 | ${ }_{43,6}$ | 60.2 | 0.46 | 14200. | 0.005 | 216. | ${ }^{851 .}$ | 1. | 3. | 0.002 | 0.22 | ${ }^{0.85}$ | 0.0106 | 2.6 |  | 1.63 | 0.21 | 7.04 | ${ }^{1770}$. | ${ }^{13.1}$ |
| ${ }^{4}$ | ss |  | 2023.0224 | Claerand colouress | 0.0057 |  | 0.095 |  | 0.101 |  | 7.18 | 25.1 | ${ }_{63}$ | 0.117 | 13100. | 0.01 | 213. | 904. | . | ${ }^{6}$ | 0.004 | 0.4 | 1. | 0.015 | 208 |  | 0.55 | 0.2 | 7.09 | 1095. | 134 |
| 524.4 (P7) | ss |  | ${ }^{202205520}$ | Coiarand sighly yelow | 0.005 |  | 0.02 |  | 0.02 |  | ${ }^{13.1}$ | 302 | ${ }^{103}$ | 0.127 | 18500. | 0.01 | ${ }^{232}$ | 1070. | ${ }^{1.3}$ | ${ }_{6} 6$ | -0.004 | 0.58 | 1.9 | ${ }^{0.0228}$ | ${ }^{24}$ |  | ${ }^{6} 32$ | 0.54 | ${ }^{6.7}$ | 2959. | ${ }^{18}$ |
| ${ }_{5}^{5240}$ ( $\mathrm{P7} 7$ | ss |  | 2022:10.05 |  | 0.0061 |  | 0.02 |  | 0.02 |  | 15.1 | 29. | 109. | 0.888 | 22300. | 0.01 | ${ }^{326 .}$ | ${ }^{1370}$ | 59. | ${ }^{6}$ | 0.004 | 0.54 | 1.7 | 0.0217 | 268 |  | ${ }^{16,1}$ | 0.87 | ${ }_{6} 6.9$ | 3482. | 18.6 |
| ${ }^{5240}$ (P7) | ss |  | $202 \cdot 2 \cdot 12.01$ | Moderaiey ubbid. moderately bown | 0.005 |  | 0.02 |  | 0.02 |  | 17.2 | ${ }_{550}$ | 108. | 0.155 | 28800. | 0.01 | ${ }^{32}$. | ${ }^{1340}$. | 1. | 6. | 0.004 | 1.17 | ${ }^{8.8}$ | ${ }^{0.0237}$ | ${ }_{5}^{567}$ |  | ${ }^{3.68}$ | ${ }^{1.53}$ | ${ }_{6}^{68}$ | 340. | 15.2 |
| 52.40 (P7) | ss |  | 2023.02 .24 | Claarand colouless | 0.005 |  | 0.027 |  | 0.027 |  | 16.6 | 27. | ${ }^{115 .}$ | 0.13 | 2080. | 0.01 | 341. | 1430. | - | 6. | - 0.004 | 29.5 | ${ }^{26}$ | 0.026 | 32 |  | 19.2 | 0.75 | 7.01 | 3063. | ${ }_{13,7}$ |
| ${ }^{80-1.0 .088)}$ | ss |  | 2020.0520 | Sighty stiy and moderatey bown | 0.05 |  | 0.2 |  | 0.2 |  | 33.4 | 2560. | ${ }^{118 .}$ | 0.336 | 17400. | - 0.01 | ${ }^{340}$. | ${ }^{793 .}$ | ${ }^{18}$ | ${ }^{7} 7$ | -0.004 | 221 | ${ }_{36} 3$. | 0.427 | ${ }^{3.9}$ | b | 5.71 | 6.53 | 7.08 | 3994. | ${ }^{15.3}$ |
| $8{ }^{80.10 .0(8)}$ | ss |  | ${ }^{2022 \cdot 10.005}$ |  | 0.05 |  | 0.2 |  | 0.2 |  | ${ }_{5} 5$ | 5230. | 159. | 0.84 | 19800. | 0.025 | 43. | 97. | 4. | 15. | 0.01 | 10.2 | 104. | 0.312 | 70 | - | 9.05 | 15.8 | 729 | 4069. | 17.7 |
| ${ }^{80-1.0 .088)}$ | ss |  | $202 \cdot 2 \cdot 1201$ | Modeatey uridid, ver b bown | 0.05 |  | 0.2 |  | 0.2 |  | 3.9 | 864. | 10. | 0.442 | 16300. | 0.005 | 29. | 87. | 5. | 15.4 | 0.0026 | 227 | 31.5 | 0.531 | 44.7 | - | 9.15 | ${ }_{9.46}$ | ${ }^{7.3}$ | 3061. | 10.2 |
| 80 | ss |  | 2023.0224 | clearand columess | 0.05 |  | 0.2 |  | 0.2 |  | 699 | 2500. | 20. | 0.668 | 25600. | - 0.01 | 57. | 1250. | 5. | 18.6 | $\leqslant 0.004$ | 11.9 | 107. | 0.593 | ${ }^{93,4}$ | b | 19.7 | 19.7 | 7.56 | 5075. | 11.8 |


| be css |  |  |  | AW Maximum (1) | ${ }^{0.2244}$ | 400 |  | 400 | 250.1500 (5) |  |  | ${ }^{20}$ |  | 0.5 .15 (5) |  |  | ${ }^{128-429} 5$ |  | 3 |  | 1000 | ${ }^{85}$ |  | ${ }_{75} 52400$ (5) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | ow Mxixum (2) | 1 | 10 |  | 10 | 80 |  |  | 10 |  | 20 | 170099 | 2500 | 500 |  |  | 2. |  | 20 | 20 | 3000 |  |  |  |  |
| Station | ${ }_{\substack{\text { Sample } \\ \text { Type }}}^{\text {den }}$ | Complane | ${ }^{\text {date Sampled }}$ | Paramear | Nitite | Nitate |  | Nititie +Nitale | Nidel | Phosphous | Potassum | Stenium | silion | swer | Sodium | Stronium | Suphate | sutur | Tralium | Tin | Tranium | Uaraium | Vanadium | znc | zrronium | рн | Conoucciniv | Tempeature |
|  |  |  |  | Fration | is | ois |  | ${ }^{1 / 5}$ | 015 | 015 | 015 | 015 | dis | ${ }_{0}$ | 015 | 015 | 015 | dis | 015 | dis | ${ }_{0} 15$ | 015 | dis | ${ }^{1 / 5}$ | ois | тот | тот | тот |
|  |  |  |  | Method Dosection Limit moul | ${ }_{\text {mal }}^{\text {mo. }}$ | ${ }_{\text {mgl }}^{0.02}$ |  | ${ }_{\text {man }}^{\text {mo }}$ | $\xrightarrow{19.02}$ | $\stackrel{\mathrm{mg}}{2}$ | ${ }_{\text {mal }}^{0.05}$ | $\xrightarrow{19.04}$ | ${ }_{\text {Len }}^{\text {¢0, }}$ |  | ${ }_{\text {mgl }}^{0.05}$ | ${ }_{\substack{\text { L9 } \\ 0.05}}$ | mal | $\stackrel{19}{4}$ | $\underbrace{}_{\substack{\text { L0, } \\ 0.02}}$ | ${ }_{0.2}^{102}$ | ${ }_{\text {real }}^{\text {res }}$ |  | ${ }_{\text {rag }}^{0.2}$ | ${ }_{0}^{19.1}$ |  | $\stackrel{\text { pH }}{0}$ | ${ }_{\text {usim }}^{0}$ | \% |
| $8{ }^{81-1.0 .099)}$ | ss |  | ${ }^{2022 \cdot 10.05}$ |  | 0.005 | 0.02 |  | 0.02 | 6.26 | ${ }^{228}$ | 285 | ${ }^{18.7}{ }^{\text {b }}$ | 17300. | -0.005 | ${ }_{36}$ | 1080. | 75. | 76.6 | 0.002 | 0.21 | 0.5 | ${ }_{0}^{0.0297}$ | 0.027 | 0.17 | 0.1 | 7.63 | ${ }^{838 .}$ | ${ }^{17.7}$ |
| P10 | ss |  | ${ }^{202205526}$ | Stighy sily and sighty orange | 0.0283 | 0.208 |  | ${ }^{0.237}$ | 4.03 | ${ }_{6}^{6.6}$ | 29.9 | 0.076 | 1470. | ${ }^{0.005}$ | ${ }^{89}$ | 1020. | 1. | ${ }^{3}$ | 0.002 | 0.47 | 0.5 | 0.012 | ${ }^{0.74}$ | 201 | 0.1 | 7.05 | 1110. | 15.8 |
| P10 | ss |  | $2022 \cdot 10 \cdot 14$ |  | 0.0089 | 0.38 |  | ${ }^{0.389}$ | 6.52 | 52 | 41.2 | 0.2 | 14900. | -0.025 | 150. | 1000. | 1. | 15. | 0.01 | 1. | 25 | 0.01 | 1. | 4.76 | 0.5 | 7.4 | 1459. | 15.5 |
| P10 | ss |  | 2022-1.201 | clears, sighly yelow | 0.007 | 0.024 |  | 0.031 | 6.67 | ${ }^{222}$ | 41.7 | 0.06 | 15700. | - 0.01 | 156. | 1060. | 1. | ${ }^{6}$ | 0.0049 | 0.4 | 1. | 0.0166 | 0.9 | 205 | 02 | 7.02 | 1414. | ${ }^{13}$ |
| 910 | ss |  | 202302.24 | vens sighty yelow, no wutbidy | 0.005 | 0.02 |  | 0.02 | ${ }^{3} 38$ | ${ }_{9} 9.7$ | 27.9 | 0.064 | 14700. | ${ }_{-}^{0.005}$ | ${ }_{84} 8$ | 957. | 1. | 3. | 0.002 | 02 | 0.5 | 0.007 | 0.77 | 1.11 | 0.1 | 7.9 | 986. | ${ }^{14.8}$ |
| P11 | ss |  | 20220.001 | Moderatey sily, ven gey | 0.119 | 19.7 | b | 19.8 | 0.786 | ${ }^{11.2}$ | 1.6 | 0.114 | ${ }^{11200 .}$ | ${ }^{0.005}$ | ${ }^{17.2}$ | 1160. | 350. | 117. | ${ }^{0.0038}$ | 02 | 0.5 | 0.78 | ${ }^{0.95}$ | ${ }^{0.54}$ | 0.1 | 7.39 | ${ }^{906}$ | 16.9 |
| P11 | ss |  | $2022 \cdot 1 \cdot 2$ | Claeand colouriess | 0.0565 | 322 | - | 323 - | 1.73 | ${ }_{6} 6$ | 3.85 | 0.69 | 6520. | - 0.005 | ${ }^{13,8}$ | 339. | 190. | ${ }_{58,4}$ | 0.0072 | 0.2 | 0.5 | 0.292 | 0.72 | 4.38 | 0.1 | ${ }_{6}^{68}$ | ${ }_{645}$ | ${ }^{11.8}$ |
| 811 | ss |  | 20230308 | clarand colouriess | 0.056 | ${ }^{30.3}$ | ${ }^{-}$ | 30.4 b | 0.927 | 6. | 22 | 0.74 | 6140. | -0.005 | 14.9 | ${ }_{380}$ | 170. | 61.2 | 0.009 | -0.2 | - 0.5 | 0.473 | 0.68 | 385 | - 0.1 | 7.06 | 716. | 10.2 |

${ }_{b}$ Above cs5 schedu 3.2 2WW Standard


| $(1)$ |
| :--- |
| (1) |
| (1) |

(4)

${ }^{(5)}$
${ }^{(5)}$

(7)

${ }_{(9)}^{(9)}$







Figure 5-4. Groundwater Quality North of Phase 1

### 5.5.3 Monitoring Site 40

Monitoring well GW-40-1-1 is not a Boundary Compliance Well and was installed between the upper and lower leachate lagoons, approximately 70 m from the property boundary. Water quality impacts associated with leachate have been investigated at this location in the past and were subject of a previous report (AECOM 2009a). In 2022/23, all parameters met the CSR AW and DW standards on all sampling dates, and conductivity values remained below $510 \mu \mathrm{~S} / \mathrm{cm}$. Conductivity in GW-40-1-1 exhibited a pattern that was consistent with the north purge wells, with the most elevated conductivity observed in September 2022. Ammonia and nitrate concentrations were higher than those observed in 2021/22, while chloride concentrations were slightly lower.

Long-term concentrations of leachate indicator parameters (i.e., conductivity, ammonia, and chloride) suggest groundwater quality continued to improve at GW-40-1-1. The improvements in groundwater quality are likely due to reduced landfill leachate generation and ongoing lowering of the Phase 1 leachate mound following closure of Phase 2 Cell 1 in 2011. Further, the operation of GW-80-1-0-P8 and GW-81-1-0-P9 as purge wells since 2016 has also likely improved leachate collection in this area. This hypothesis is supported by the observation of more concentrated leachate collected by the North Purge wells in 2022/23. Statistical trend analysis revealed a decreasing trend in chloride concentrations over the past five years. However, sulphate concentrations were elevated slightly throughout 2022/23, ranging from 32 to $56 \mathrm{mg} / \mathrm{L}$, and nitrate concentrations were elevated, ranging from 1.41 to $6.06 \mathrm{mg} / \mathrm{L}$. Ammonia concentrations were elevated throughout the monitoring year, ranging from 0.52 to $1.60 \mathrm{mg} / \mathrm{L}$. These water quality trends may reflect aggregate runoff impacts.

Water quality at GW-40-1-1 should continue to be closely monitored to verify the effectiveness of leachate collection north of the Phase 1 landfill and to identify any potential water quality impacts associated with aggregate stockpiling at Hartland. The ongoing purge well performance evaluation program should be expanded to include GW-81-1-0-P9, with the value of each well assessed based on the rate of contaminant mass removal.

### 5.5.4 Monitoring Sites 20 and 21

Monitoring locations 20 and 21 are considered Boundary Compliance Wells and are located directly north of the Phase 1 landfill and the lower leachate lagoon and between 5 and 15 m from the landfill property boundary. These monitors are in the most probable path for any potential subsurface leachate migration below the unlined lower lagoon. The concentrations of leachate indicator parameters at locations 20, 21, 40, 52, and 58 are plotted against time in Figure 5-4.

In 2022/23, groundwater quality at locations 20 and 21 met both CSR AW and DW standards, except for one copper concentration ( $41.5 \mu \mathrm{~g} / \mathrm{L}$ ) in GW-21-1-1 in May 2022, which exceeded the CSR AW standard. However, the recorded exceedance may not accurately represent in-situ water quality due to the high discrepancy between parent and duplicate samples. Copper concentrations observed in GW-21-1-1 throughout the rest of the monitoring year were low, ranging from 0.227 to $0.285 \mu \mathrm{~g} / \mathrm{L}$.

Groundwater quality in wells GW-21-1-2 and GW-21-2-1 have generally exhibited higher conductivity and ammonia/chloride concentrations because they were screened at shallower depths than adjacent wells. In 2022/23, conductivity concentrations in wells GW-21-1-2 and GW-21-2-1 ranged from 316 to $489 \mu \mathrm{~S} / \mathrm{cm}$, slightly lower than those observed in 2021/22. Chloride concentrations at location 21 have remained well-below the CSR DW standard ( $250 \mathrm{mg} / \mathrm{L}$ ) since 1997 and below $60 \mathrm{mg} / \mathrm{L}$ in $2022 / 23$. At location 20, conductivity values stayed below $220 \mu \mathrm{~S} / \mathrm{cm}$, and ammonia and chloride concentrations were consistent with background values. Sulphate and ammonia remained within their historical ranges in all four wells.

Over the past five years, groundwater quality data have exhibited statistically significant decreasing trends in ammonia in wells GW-21-2-1 and GW-20-1-2. Furthermore, chloride concentrations have exhibited a statistically significant decreasing trends in wells GW-20-1-2 and GW-20-1-1. These decreasing trends in leachate indicator parameters shows improved water quality at these locations. Overall, groundwater at monitoring stations 20 and 21 was not impacted by landfill leachate or aggregate stockpiling.

### 5.5.5 Monitoring Site 31

Monitoring wells at location 31 are considered Boundary Compliance Wells and they are located along the landfill north property line, south of Willis Point Road and 160 m northeast of the lower leachate lagoon. They are downgradient of the landfill and have the lowest groundwater elevations measured at the site.

In 2022/23, groundwater quality at location 31 met both CSR AW and DW standards. While the concentrations of conductivity, hardness, ammonia, sulphate, and nitrate were lower than their peak levels observed in Q3 and Q4 of 2021, which were
associated with the wastewater residual spill event in October 2020, these parameters remained elevated compared to historical ranges. Conductivity and sulphate concentrations remain elevated, ranging from 436 to $589 \mu \mathrm{~S} / \mathrm{cm}$ and 85.0 to 270 $\mathrm{mg} / \mathrm{L}$, respectively. Moreover, sulphate and conductivity in groundwater samples collected from GW-31-1-1 and GW-31-1-2 have exhibited statistically significant increasing trends for the past five years. In 2022/23, average nitrate concentrations increased by a factor of 4.0 in GW-31-1-2 (average of $1.36 \mathrm{mg} / \mathrm{L}$ ) and by a factor of 1.5 in GW-31-1-1 (average of $0.65 \mathrm{mg} / \mathrm{L}$ ). Groundwater quality at this location is impacted by aggregate stockpiling on the Northeast Ridge and should continue to be monitored closely.

### 5.5.6 Monitoring Sites 29 and 30

Monitoring locations 29 and 30 are located north of Willis Point Road and are Boundary Compliance Wells. Long-term conductivity, ammonia, and chloride concentrations trends observed at these locations are displayed in Figure 5-5. In 2022/23, all parameters met applicable CSR standards at sites 29 and 30.

At monitoring locations 29 and 30, elevated conductivity and chloride concentrations observed since 2007 likely reflect intermittent road salting on Willis Point Road. The District of Saanich has confirmed the use of de-icing salt (sodium chloride) on an as-needed basis on this road for several years and that no records of application dates are kept. As shown on Figure 55 , conductivity and chloride concentrations measured in the shallow monitors at locations 29 and 30 have exhibited seasonal fluctuations for a very long time, with maximum concentrations typically occurring in winter and early spring months. In $2022 / 23$, chloride concentrations at locations 29 and 30 remained below $100 \mathrm{mg} / \mathrm{L}$, and ammonia concentrations remained very low ( $<0.02 \mathrm{mg} / \mathrm{L}$ ). Low conductivity, chloride, and ammonia concentrations indicate that landfill leachate is not impacting groundwater quality north of Willis Point Road. The occasional elevated ammonia/nitrate concentration may reflect organic processes that occur in lowland environments during the winter season, which is supported by simultaneous observation of increased iron and manganese concentrations.

In 2022/23 nitrate concentrations observed at sites 29 and 30 continued to increase, with the average nitrate concentration observed at GW-29-1-1 increasing to $1.19 \mathrm{mg} / \mathrm{L}$, and average nitrate concentrations observed at Site GW-30-1-2 increasing to $1.54 \mathrm{mg} / \mathrm{L}$. Over the past five years, groundwater quality data collected from wells at locations 29 and 30 has exhibited statistically significant decreasing trends in chloride and nitrate concentrations in GW-30-1-1, however increasing trends in conductivity have been observed at GW-29-1-2 and GW-30-1-1. Additionally, increasing trends in sulphate concentrations have been observed at GW-29-1-1, GW-29-1-2, and GW-30-1-2. Sulphate concentrations at these stations were $<51.5 \mathrm{mg} / \mathrm{L}$, but the occasionally elevated nitrate concentrations are likely associated with surface water runoff from the Northwest and Northeast aggregate stockpiles.

### 5.5.7 Monitoring Sites 28 and 39

Groundwater monitoring wells 28 and 39 are located between the upper leachate lagoon and Willis Point Road and are considered Boundary Compliance Wells. In 2022/23, groundwater quality at sites 28 and 39 met CSR standards.

Similar to previous years, ammonia concentrations observed in well GW-28-1-0 were below the detection limit, and chloride concentrations were low, indicating groundwater quality was not impacted by landfill leachate. Sulphate concentrations were low, and over the past five years, groundwater quality in well GW-28-1-0 has exhibited statistically significant increasing trends in conductivity and nitrate concentrations. However, nitrate concentrations were less than $0.514 \mathrm{mg} / \mathrm{L}$ on all sampling dates.

At groundwater monitoring station 39, conductivity values were below $400 \mu \mathrm{~S} / \mathrm{cm}$, and nitrate concentrations decreased on most sampling dates. However, nitrate concentration was elevated at $7.49 \mathrm{mg} / \mathrm{L}$ in GW-39-1-1 and $1.94 \mathrm{mg} / \mathrm{L}$ in GW-39-2-1 in March 2023. Groundwater quality in both wells has exhibited statistically significant increasing trends in sulphate and nitrate concentrations, and a decreasing trend in ammonia concentration has been observed at GW-39-1-1. On all sampling dates, sulphate concentrations in both wells were below $25 \mathrm{mg} / \mathrm{L}$. Overall, the slight increase in nitrate levels observed in groundwater at sites 28 and 39 may suggest minor impacts related to aggregate production, stockpiling and road construction activities along the northern property boundary. Groundwater quality at these locations should be monitored closely to evaluate aggregate runoff impacts.


Figure 5-5. Groundwater Quality North of Willis Point Road

### 5.6 Monitors West and North of the Phase 2 Landfill

Background groundwater quality has been represented historically by monitoring location 63 (west of Phase 2) and five monitoring locations (77, 78, 79, 87, and 88) north of the Phase 2 landfill. Throughout the winter and spring of 2016/17, a portion of the slope between Phase 2 and the Hartland North Pad was cleared of vegetation to allow for construction of an aggregate storage area (The Northwest Stockpile). Since clearing, overburden soils have been stripped to expose the bedrock surface that consists of several closed depressions that may impact groundwater recharge. This work was conducted in the vicinity of monitoring locations $77,78,79,87$ and 88 . Groundwater monitoring wells GW-79-1-1 and GW-79-2-1 were decommissioned in May 2018 to accommodate aggregate stockpiling. In 2019/20, excavation and blasting were conducted at the North Pad area and east of the Toutle Valley Road. Given the recent land clearing, excavation, and quarrying activities, water quality data from locations $77,78,87$ and 88 was not considered representative of background groundwater quality since 2020, and groundwater at these wells should be closely monitored for landfill, aggregate stockpiling and construction related impacts.

Throughout July and September 2022, seven new monitoring wells (GW-104-1-1, GW-105-1-1, GW-106-1-1, GW-107-1-1, GW-108-1-1, GW-109-1-1, and GW-110-1-1) were installed North of Phase 2, near the Contractor's Shed, to characterize hydrogeological conditions near the Northwest Sedimentation Pond (NWSP) and monitor impacts associated with landfill operations and construction, aggregate production, and storage. The site investigation results are summarized in AECOM (2023). Quarterly groundwater sampling at these locations commenced in September 2022.

Overall, groundwater quality was monitored at two background locations (63 and 94), six locations (25, 27, 36, 37, 38 and 53) north of the Phase 2 landfill near the upper leachate lagoon, eight locations (41, 42, 43, 44, 55, 56,57 and 62) near the Hartland North Pad, and ten locations (95, 96, 97, 98, 103, 104, 105, 106, 107, and 108) near the perimeter of the Phase 2 Basin, the NWSP, and Heal Creek.

### 5.6.1 Background Groundwater Quality

Monitoring well 63 is located at the western edge of the property, upgradient of the landfill, where groundwater quality is considered representative of background conditions. In 2022/23, groundwater quality at Site 63 was consistent with the previous monitoring year, and leachate/aggregate indicator parameters including conductivity, chloride, ammonia, nitrate, and sulphate concentrations were low and consistent with historical values. Groundwater quality at Site 63 shows no signs of landfill impacts.

In previous years, groundwater quality at monitoring station 94 was considered representative of background conditions. In 2022/23, groundwater quality at station 94 was characterized by low but variable conductivity ( $190-400 \mu \mathrm{~S} / \mathrm{cm}$ ), and low sulphate ( $<49 \mathrm{mg} / \mathrm{L}$ ), chloride ( $<10 \mathrm{mg} / \mathrm{L}$ ), and ammonia ( $<0.015 \mathrm{mg} / \mathrm{L}$ ) concentrations. Statistically significant decreasing trends in chloride and sulphate have been observed in well GW-94-1-1 since it was installed in 2019. However, nitrate concentrations increased from the previous monitoring year. In 2021/22, nitrate concentrations were below detection on all sampling dates, and in 2022/23, nitrate exceeded the detection limit on three of four sampling dates, reaching a maximum concentration of $2.58 \mathrm{mg} / \mathrm{L}$ in March 2022, and an average of $0.68 \mathrm{mg} / \mathrm{L}$ throughout the monitoring year. Although sulphate concentrations were still below threshold of $75 \mathrm{mg} / \mathrm{L}$ in 2022/23, the elevated nitrate concentrations likely reflect aggregate stockpiling on the North Ridge, so groundwater quality at this site can no longer be considered representative of background conditions.

### 5.6.2 Wells North of the Phase 2 Landfill

In 2022, AECOM performed a water quality assessment to identify contaminant sources and pathways conveying dilute leachate from Phase 2 and nitrogen-rich runoff from the Northwest Aggregate Stockpile toward the NWSP and Heal Creek (AECOM 2023). A total of 17 wells were drilled around the northern perimeter of Phase 2, the NWSP, and Heal Creek to assess the groundwater quality on a local scale. These new wells were sampled quarterly, and the data has been integrated into 2022/23 annual monitoring report. Figures 5-6, 5-7, and 5-8 illustrate groundwater quality with selected landfill and leachate indicator parameters.

### 5.6.2.1 Monitoring Site 36

Monitoring location 36 is not a Boundary Compliance Well and is located 20 m northeast and downgradient of the Phase 2 basin. In 2022/23, well GW-36-3-1 was monitored and sampled on a quarterly basis. Well GW-36-2-1 has not been monitored since 2016/17 due to well construction related impacts. Well GW-36-3-1 is well-suited to assess whether leachate is migrating beneath the clay liner along the north side of Phase 2 because the elevation of the bottom of the well screen is 112 m asl, and
the bottom of the Phase 2 basin is at 113 m asl. The bottom of the clay liner along the north side of the Phase 2 basin is at 114 m asl.

In 2022/23, groundwater quality in GW-36-3-1 was consistent with the previous monitoring year, exhibiting elevated conductivity ( 805 to $1,348 \mu \mathrm{~S} / \mathrm{cm}$ ), nitrate ( 0.055 to $0.583 \mathrm{mg} / \mathrm{L}$ ) and sulphate ( 140 to $240 \mathrm{mg} / \mathrm{L}$ ) concentrations. Conductivity slightly decreased following the installation of a leachate sump, and chloride concentrations remained stable and marginally below $20 \mathrm{mg} / \mathrm{L}$. Ammonia concentrations were generally low, with one spike observed in May 2022. Over the past five years, statistically significant increasing trends in conductivity and chloride concentration have been observed, as well as a decreasing trend in nitrate concentrations. The elevated conductivity, sulphate, and nitrate concentrations likely reflect groundwater quality impacts associated with aggregate stockpiling on the North Ridge, and the low levels of leachate indicator parameters confirm that leachate has not migrated beneath the clay liner at the north edge of Phase 2.

Groundwater quality at this location should continue to be monitored closely to evaluate the severity and evolution of aggregate runoff impacts in this area of the landfill.

### 5.6.2.2 Monitoring Site 37

Groundwater monitoring station 37 is not a Boundary Compliance Well, and it is located 25 m north of the Phase 2 basin. In 2022/23, well GW-37-3-1 was routinely monitored, and well GW-37-1-1 has been excluded from the monitoring program due to the limited access. Sampling of GW-37-2-1 was discontinued in 2017 due to its low concentrations of leachate indicator parameters.

In 2022/23, conductivity and sulphate concentrations observed at GW-37-3-1 ranged from 590 to $838 \mu \mathrm{~S} / \mathrm{cm}$, and 110 to 240 $\mathrm{mg} / \mathrm{L}$, respectively, showing substantial increases from the previous monitoring year. The average conductivity observed at GW-37-3-1 increased by about 48\% from the previous monitoring year, whereas sulphate concentrations increased by about $28 \%$. Chloride concentrations remained low (i.e., $<15 \mathrm{mg} / \mathrm{L}$ ), but the average chloride concentration doubled from the previous monitoring year. Ammonia and nitrate concentrations increased slightly, peaking at levels of $0.51 \mathrm{mg} / \mathrm{L}$ and $0.689 \mathrm{mg} / \mathrm{L}$, respectively. Due to the low chloride and ammonia concentrations, water quality degradation is interpreted to be the result of onsite aggregate stockpiling and not leachate.

Throughout the 2021/22 and 2022/23 monitoring periods, groundwater quality at monitoring station 37 degraded slightly, showing evidence of impacts associated with aggregate stockpiling. Groundwater quality at this location should be monitored closely to confirm the cause(s) of groundwater quality changes in this area. Consideration should be given to reinitiating sampling of GW-37-2-1.


Figure 5-6. Groundwater Quality North of Phase 2 - Landfill Leachate Impacts

### 5.6.2.3 Monitoring Site 38

Groundwater monitoring station 38 is not a Boundary Compliance Well and is located roughly 80 m north of the Phase 2 basin immediately north of the upper leachate lagoon. Water quality at this location met CSR standards for all parameters in 2022/23, and conductivity continued to increase from previous monitoring years. Nitrate concentrations increased from an average of $0.28 \mathrm{mg} / \mathrm{L}$ in $2021 / 22$ to $0.35 \mathrm{mg} / \mathrm{L}$ in $2022 / 23$. Ammonia and chloride concentrations were low and within historical ranges. Sulphate concentrations were below $50 \mathrm{mg} / \mathrm{L}$ and slightly lower than the previous monitoring year. Over the past five years, no statistically significant trends in groundwater quality have been observed, indicating that the groundwater quality is stable.

### 5.6.2.4 Monitoring Sites 25 and 53

Wells GW-25-1-1, GW-25-1-2 (not Boundary Compliance Wells) and GW-53-1-1 (Boundary Compliance Well) are located on an east-west trending ridge 100 m north of the Phase 2 basin near the property boundary. These wells have exhibited good water quality in the past and remain unaffected by landfill leachate. However, evidence of aggregate runoff impacts were apparent in 2022/23, where water quality at all three wells met CSR AW and DW standards except for nitrate ( $12.9 \mathrm{mg} / \mathrm{L}$ ) at GW-25-1-1 in May 2022.

In 2022/23, water quality in GW-25-1-1 and GW-25-1-2 degraded slightly, exhibiting higher average conductivity and nitrate concentrations than the previous monitoring year. Nitrate concentrations were highest in the deep well, GW-25-1-1, ranging from 3.54 to $12.9 \mathrm{mg} / \mathrm{L}$. In GW-25-1-2, nitrate concentrations ranged from 0.212 to $0.655 \mathrm{mg} / \mathrm{L}$. On all sampling dates, sulphate concentrations were moderate to elevated, ranging from 52 to $190 \mathrm{mg} / \mathrm{L}$, and chloride and ammonia concentrations were low. Statistically significant decreasing trends in ammonia and chloride concentrations were observed in GW-25-1-1, along with increasing trends in sulphate and conductivity. In well GW-25-1-2, statistically significant increasing trends in sulphate and conductivity were observed.

Overall, groundwater quality at monitoring station 25 continued to show no evidence of landfill leachate contamination, but water quality impacts associated with aggregate stockpiling were prevalent, especially in well GW-25-1-1. This indicates that nitrogen-rich runoff from the Northwest Stockpile has infiltrated relatively deep bedrock around the NWSP. Groundwater quality at this location should continue to be monitored closely to monitor aggregate runoff impacts.

In 2022/23, water quality in GW-53-1-1 was relatively stable, exhibiting background conductivity values, and low chloride, ammonia, sulphate, and nitrate concentrations. Over the past five years, statistically significant increasing trends in conductivity and sulphate concentrations have been detected, and chloride concentrations have trended downward. Still, solute concentrations are low, providing no indication of groundwater quality impacts associated with aggregate stockpiling or landfill leachate.

### 5.6.2.5 Monitoring Site 27

Groundwater monitoring station 27 is located northwest of Phase 2 and adjacent to aggregate stockpiles, where shallow groundwater has been historically impacted by blasting residues. Similar to previous years, groundwater quality in well GW-27-1-1 met all applicable CSR standards. Well GW-27-1-2 was destroyed during construction in 2022 and could not be sampled.

In 2022/23, Groundwater quality at well GW-27-1-1 showed no evidence of impacts associated with landfill leachate or aggregate runoff; nitrate concentrations were below the detection limit, and ammonia, chloride, and sulphate concentrations were low. Conductivity values increased from the previous monitoring year, exhibiting an average of $221 \mu \mathrm{~S} / \mathrm{cm}$. In the past five years, statistically significant increasing trends in conductivity and sulphate concentrations have been observed, but all solute concentrations reflected background values.

Ultimately, hydraulic conditions and contaminant concentrations should be carefully monitored at this location because groundwater in this area reports to the surface water collection and conveyance system. Ongoing quarry development may impact well integrity and any wells in this area should be properly decommissioned in advance of construction. It is recommended that the CRD commence early planning for shallow groundwater, surface water, and leachate management in this area to maximize diversion of clean water out of the leachate collection system.

### 5.6.2.6 Monitoring Sites Along the Northern Edge of the Phase 2 Basin

Five monitoring wells (GW-103-1-1, GW-104-1-1, GW-105-1-1, GW-106-1-1, and GW-107-1-1/GW-107-2-1) were installed just outside the northwestern edge of the Phase 2 Basin and monitored in $2022 / 23$. These wells were sampled quarterly, beginning in September 2022. The groundwater quality observed in these wells varies, with some wells (GW-105-1-1 and GW-106-1-1) showing distinct evidence of landfill leachate impacts, and other wells (GW-104-1-1, GW-105-1-1, and GW-107-1-1) exhibiting impacts related to aggregate stockpiling on the North Ridge. Well GW-103-1-1 was installed at the toe of the North Ridge, west of the Contractor Shed, and showed no evidence of groundwater quality impacts.

In 2022/23, monitoring well GW-104-1-1 showed evidence of groundwater quality impacts associated with aggregate stockpiling on the North Ridge and road-salt. Conductivity ranged from 160 to $1,000 \mu \mathrm{~S} / \mathrm{cm}$, and increased throughout the monitoring year. Chloride concentrations showed a similar trend, increasing from $4.4 \mathrm{mg} / \mathrm{L}$ in September to $120 \mathrm{mg} / \mathrm{L}$ in March. Elevated sodium ( $36.6 \mathrm{mg} / \mathrm{L}$ ) and calcium ( $130 \mathrm{mg} / \mathrm{L}$ ) concentrations were observed on the same sampling date, which may indicate that the groundwater was affected by road-salt. Ammonia concentrations were below the detection limits, so it is unlikely that the elevated chloride concentration was caused by landfill leachate. Nitrate concentrations increased throughout the monitoring year from $0.502 \mathrm{mg} / \mathrm{L}$ to $21.5 \mathrm{mg} / \mathrm{L}$, exceeding the CSR DW standard on two of three sampling dates. Like conductivity, nitrate, and chloride concentrations, sulphate concentrations generally increased throughout the monitoring year, ranging from 16 to $140 \mathrm{mg} / \mathrm{L}$. Overall, groundwater quality at monitoring station 104 showed evidence of aggregate runoff impacts, especially near the end of the wet season. Given the undetectable levels of ammonia concentrations, it is unlikely that the groundwater was affected by landfill leachate.

In 2022/23, monitoring well GW-105-1-1 showed similar trends as GW-104-1-1, with the highest conductivity ( $1,300 \mu \mathrm{~S} / \mathrm{cm}$ ), chloride ( $87 \mathrm{mg} / \mathrm{L}$ ), sulphate ( $290 \mathrm{mg} / \mathrm{L}$ ) and nitrate ( $32 \mathrm{mg} / \mathrm{L}$ ) concentrations observed in March 2023. Zinc and strontium concentrations were also elevated during the March 2023 sampling event. All solute concentrations observed at GW-105-1-1 met CSR standards, except for nitrate on all sampling dates. Ammonia concentrations were slightly elevated, ranging from 0.05 to $0.073 \mathrm{mg} / \mathrm{L}$. Given the low sodium levels, the source of the elevated chloride is unlikely to be road salt. The concurrent rise in conductivity, chloride, and metal levels, coupled with a mild increase in ammonia concentrations, could suggest the presence of dilute landfill leachate. However, the elevated nitrate and sulphate concentrations also suggest that effects of aggregate stockpiling on the North Ridge are more pronounced.

In September 2022, groundwater in well GW-106-1-1 exhibited exceptionally high conductivity ( $2,160 \mu \mathrm{~S} / \mathrm{cm}$ ), ammonia ( 91 $\mathrm{mg} / \mathrm{L}$ ), and chloride ( $140 \mathrm{mg} / \mathrm{L}$ ) concentrations, characteristic of dilute landfill leachate. Ammonia concentrations exceeded CSR AW standards, and nitrite ( $1.92 \mathrm{mg} / \mathrm{L}$ ) and nitrate ( $10.9 \mathrm{mg} / \mathrm{L}$ ) concentrations exceeded CSR DW standards. Sulphate concentrations were relatively high throughout the monitoring year, ranging from 170 to $260 \mu \mathrm{~S} / \mathrm{cm}$, and groundwater quality was impacted by aggregate stockpiling on the North Ridge. The leachate impacts observed at monitoring station 106 reflect a leachate impact beyond the northern boundary of the Phase 2 Leachate Collection System, which impacted groundwater quality around the NWSP and downstream in Heel Creek in 2021/22 (AECOM 2023).

In early 2022, groundwater in well GW-107-1-1 exhibited elevated conductivity ( 627 to $1,100 \mu \mathrm{~S} / \mathrm{cm}$ ), nitrate ( 9.07 to 23.9 $\mathrm{mg} / \mathrm{L}$ ) and sulphate ( 190 to $270 \mathrm{mg} / \mathrm{L}$ ) concentrations, with nitrate concentrations exceeding the CSR DW standard. Chloride concentrations were slightly elevated, ranging from 15 to $32 \mathrm{mg} / \mathrm{L}$, and ammonia concentrations were generally near or at the detection limit. The elevated conductivity, sulphate, and nitrate concentrations in the absence of elevated chloride and ammonia concentrations signify groundwater quality impacts related to aggregate stockpiling on the North Ridge.

Overall, groundwater quality in this area is primarily impacted by runoff from aggregate stockpiling, with some wells (GW-106-1-1 and GW-105-1-1) showing evidence of dilute landfill leachate impacts. The influence of landfill leachate on groundwater is limited to an area less than 20 meters northwest of the Phase 2 Basin. However, the impact of aggregate runoff is more extensive, with most wells showing elevated nitrate and sulphate concentrations. Well GW-103-1-1 was installed at the toe of the North Ridge, west of the Contractor Shed, and showed no evidence of groundwater quality impacts. The use of magnesium lignosulfonate for road dust control has also resulted in increased sulphate concentration and conductivity at some locations.

### 5.6.2.7 Monitoring Sites near the Northwest Sedimentation Pond and Heal Creek

In 2022, during a Site Investigation, five monitoring wells (GW-95-1-1, GW-96-1-1, GW-97-1-1, GW-98-1-1, and GW-108-1-1) were installed around the NWSP and Heal Creek to assess the groundwater quality (AECOM 2023). These wells were sampled quarterly beginning in September 2022. Most of these wells have exhibited evidence of groundwater quality impacts associated with aggregate stockpiling.

Wells GW-95-1-1 and GW-96-1-1 were installed metres apart, along the north edge of Heel Creek, downstream of the NWSP. Groundwater quality in both wells was consistent with the CSR AW and DW guidelines, except for the groundwater sample collected from GW-95-1-1 in September 2022. On that sampling date, exceedances were observed in aluminum, chromium, cobalt, copper, and vanadium concentrations, and silicon and iron concentrations were exceptionally high. CRD staff noted that turbidity was very high during sampling, which likely explains the anomalous trace metal concentrations, as well as the elevated silicon and iron concentrations. At these wells, conductivity reflected background values, ranging from 392 to 490 $\mu \mathrm{S} / \mathrm{cm}$, and chloride and ammonia concentrations were low. In December 2022, nitrate concentrations were elevated in both wells, ranging from 2.65 to $8.68 \mathrm{mg} / \mathrm{L}$, and sulphate concentrations were elevated, ranging from 100 to $120 \mathrm{mg} / \mathrm{L}$. Overall,
groundwater quality at monitoring stations 95 and 96 show no evidence of landfill leachate impacts, but elevated nitrate and sulphate concentrations suggest impacts by runoff emanating from the Northwest Stockpile.

Wells GW-97-1-1 and GW-98-1-1 were installed on either side of Heal Creek, between Heal Creek and the NWSP. Well GW-97-1-1 was installed south of Heel Creek, and well GW-98-1-1 was installed north of Heal Creek. Nitrate concentrations in groundwater collected from GW-98-1-1 exceeded the CSR DW in April 2022. At GW-97-1-1, groundwater quality met all CSR standards on all sampling dates. However, conductivity, nitrate and sulphate concentrations increased substantively in December 2022, when sulphate and nitrate concentrations reached $170 \mathrm{mg} / \mathrm{L}$ and $8.56 \mathrm{mg} / \mathrm{L}$, respectively. Conductivity increased from $290 \mu \mathrm{~S} / \mathrm{cm}$ in September to $472 \mu \mathrm{~S} / \mathrm{cm}$. Conversely, ammonia concentration was elevated in September 2022 ( $0.91 \mathrm{mg} / \mathrm{L}$ ) but decreased in December ( $0.016 \mathrm{mg} / \mathrm{L}$ ). At GW-98-1-1, conductivity, nitrate, ammonia, and sulphate concentrations were elevated in April and September 2022, but concentrations generally subsided after the installation of a leachate sump. Chloride concentrations were low in both wells (i.e., $<5.0 \mathrm{mg} / \mathrm{L}$ ), so there is no evidence of landfill leachate impacts in this area. However, groundwater in these two wells was impacted by aggregate runoff.

Well GW-108-1-1 was installed a few metres south of the NWSP and screened in relatively deep bedrock ( 14.48 m BGS; 116.2 m ASL ). On all sampling dates, groundwater quality at GW-108-1-1 met all CSR standards, and conductivity was elevated, ranging from 519 to $640 \mu \mathrm{~S} / \mathrm{cm}$. Chloride, ammonia, and nitrate concentrations were low, and sulphate concentrations were elevated, ranging from 200 to $230 \mu \mathrm{~S} / \mathrm{cm}$. The elevated sulphate concentration may be due to aggregate stockpiling or natural sulphide minerals present in the bedrock, which has been observed in several monitoring wells around the landfill footprint in the past (AECOM 2023). Groundwater quality at Site 108 was good, with no evidence of impacts related to landfill leachate or aggregate runoff.

Overall, the aggregate stockpiles located in the Toutle Valley and atop the bedrock ridge north of the Phase 2 landfill are interpreted to be the primary source of nitrate and sulphate observed in groundwater wells. This is mainly evident in the shallow aquifer. However, the deeper groundwater in the bedrock aquifer, as seen in well GW-25-1-1, is also influenced by the aggregate stockpiles. This suggests that runoff from aggregate stockpiles has entered the bedrock before moving towards Heal Creek. However, there is no indication of landfill leachate impacts on groundwater in this area.

### 5.6.3 Wells near Hartland North Pad (Residual Treatment Facility)

The Hartland North Pad has had a variety of uses including yard waste composting (1994-2004), aggregate stockpiling (2006 to 2018), and currently, it accommodates the north trailer and new north scale residue. ). The aggregate is only stored on the ridge south of the North Pad. Additionally, the area was a construction site for the RTF associated with CRD's McLoughlin Point Wastewater Treatment Plant project. The construction of the RTF buildings began in early 2019 and was completed in September 2020. The RTF is now operational and permitted under ENV Operational Certificate \#109471.

Groundwater monitoring stations $41,42,55,56$ and 57 are considered Boundary Compliance Wells around the Hartland North Pad. Monitoring locations 43,44 and 62 were installed adjacent to the Hartland North Pad and are not Boundary Compliance Wells. Monitoring locations 91, 92, 93 and 94 were established in November 2019 and located south of the RTF. 93-1-1 was destroyed in 2022 due to construction and no sample was collected in 2021/22.

In 2022/23, groundwater quality in all monitoring wells on the North Pad met all applicable CSR standards.

### 5.6.3.1 Monitoring Sites 44 and 62

Groundwater sampling at well GW-44-1-1 commenced in 2016/17 to track potential construction related impacts at the Hartland North Pad area. Well GW-44-1-1 is located southwest of monitoring location 43, and the groundwater quality is generally consistent with background conditions. In 2022/23, groundwater quality in GW-44-1-1 was consistent with background values, exhibiting moderate to low conductivity ( $<450 \mu \mathrm{~S} / \mathrm{cm}$ ), and low sulphate ( $<42 \mathrm{mg} / \mathrm{L}$ ), chloride ( $<6.0 \mathrm{mg} / \mathrm{L}$ ) and nitrate ( $<0.024 \mathrm{mg} / \mathrm{L}$ ) concentrations. Ammonia concentrations were below detection. Over the past five years, a statistically significant increasing trend in nitrate concentrations has been observed. Overall, water quality at Site 44 is generally consistent with background conditions.

Monitoring location 62 is located southwest of monitoring location 44 . Similar to previous years, groundwater quality at GW-62-1-1 (23.7 m BGS) and GW-62-2-1 (18.9 m BGS) reflected background conditions throughout 2022/23, except for slightly elevated nitrate concentrations in GW-62-1-1 that increased from $0.278 \mathrm{mg} / \mathrm{L}$ in September 2022 to $1.8 \mathrm{mg} / \mathrm{L}$ in December. Despite the elevated nitrate concentrations, sulphate concentrations were minimal. Over the past five years, a statistically significant increasing trend in nitrate concentrations and decreasing trends in chloride and conductivity have been observed at GW-62-1-1. In the shallow well, GW-62-2-1, nitrate concentrations reached a maximum of $0.473 \mathrm{mg} / \mathrm{L}$, ammonia concentrations were below detection, sulphate concentrations were below $26 \mathrm{mg} / \mathrm{L}$, and chloride concentrations were below
$2.9 \mathrm{mg} / \mathrm{L}$. Ultimately, groundwater quality at Site 62 shows no evidence of landfill leachate impacts, and the elevated nitrate concentrations in GW-62-1-1 likely do not reflect aggregate runoff because sulphate concentrations were low.

### 5.6.3.2 Monitoring Sites 41, 42, 43, 55, 56 and 57

Figure 5-8 displays leachate indicator parameters for monitoring locations 41, 42, 43, 55 and 56 . In 2022/23, water quality at all of these locations was consistent with historical data. Conductivity in all wells increased during the wet season, reaching a max of $640 \mu \mathrm{~S} / \mathrm{cm}$ at Site 43 . Sulphate concentrations were relatively high at monitoring stations 42 and 43 , ranging from 69 to $86 \mathrm{mg} / \mathrm{L}$.

Over the past five years, statistically significant increasing trends in conductivity have been observed at groundwater monitoring stations $42,43,56$ and 57 . An increasing trend in sulphate concentration was observed at monitoring station 43 , and an increasing trend in chloride concentration was detected at monitoring station 41 . Decreasing trends in chloride concentrations were identified at monitoring stations 41 and 43.

Groundwater quality in this area of the landfill generally reflects background conditions, with slightly elevated sulphate concentration. Groundwater quality should continue to be monitored for any impacts associated with landfill operations, including quarrying, aggregate stockpiling, and construction.

### 5.6.3.3 Monitoring Sites 77, 78, 87 and 88

In 2022/23, concentrations of all parameters measured at monitoring stations 77, 78, 87 and 88 (not Boundary Compliance Wells) were below applicable CSR standards.

Groundwater quality at monitoring station 77 exhibited low leachate indicator parameters and nitrate concentrations. Groundwater quality at station 78 exhibited low leachate and aggregate runoff indicator parameters, but nitrate concentrations were elevated in the shallow well (GW-78-2-1), ranging from 1.66 to $2.35 \mathrm{mg} / \mathrm{L}$. Sulphate concentrations observed in well GW-$78-2-1$ were low to moderate, ranging from 40 to $46 \mathrm{mg} / \mathrm{L}$. Statistically significant decreasing trends in ammonia and chloride concentrations were observed over the past five years in well GW-78-2-1, and increasing trends in sulphate concentration have been identified in wells GW-77-2-1 and GW-78-1-1. The elevated nitrate and low to moderate sulphate concentrations observed at monitoring station 78 likely reflect aggregate runoff impacts, given the position of well downgradient of the Northwest Stockpile. Well GW-77-1-1 may have been screened at a depth that is above the flow path conveying aggregate runoff to monitoring station 78.

Groundwater monitoring stations 87 and 88 are located on either side of the Highland Fault, immediately south of the Northwest Stockpile. Groundwater collected from location 87 exhibited low to moderate nitrate concentrations, ranging from below the detection limit to $0.402 \mathrm{mg} / \mathrm{L}$ in well GW-87-2-1. Chloride concentrations were low (i.e., $<5 \mathrm{mg} / \mathrm{L}$ ), ammonia concentrations were below the detection limit, and sulphate concentrations were $<25 \mathrm{mg} / \mathrm{L}$. At location 88 , nitrate concentrations were appreciably higher, ranging from 6.20 to $6.72 \mathrm{mg} / \mathrm{L}$ in well GW-88-1-1 and from 8.13 to $7.87 \mathrm{mg} / \mathrm{L}$ in well GW-88-2-1. Additionally, sulphate concentrations were elevated, ranging from 47 to $54 \mathrm{mg} / \mathrm{L}$ in well GW-88-1-1 and from 99 to $250 \mathrm{mg} / \mathrm{L}$ in well GW-88-2-1. Ammonia concentrations were below the detection limit, and chloride concentrations were below $3.6 \mathrm{mg} / \mathrm{L}$. Over the past five years, a statistically significant increasing trend in sulphate concentration has been observed at well GW-87-1-1, and an increasing trend in conductivity and nitrate concentration has been observed at well GW-87-2-1. At location 88, increasing trends in conductivity, nitrate, and sulphate concentrations have been identified in GW-88-1-1, as well as decreasing trends in ammonia and chloride concentrations in GW-88-2-1. Groundwater quality in this area of the landfill shows no evidence of landfill leachate impacts. However, elevated nitrate and sulphate concentrations observed in the shallow well at location 88 clearly reflect aggregate impacts associated with runoff from the Northwest Stockpile.

### 5.6.3.4 Monitoring Sites 91 and 92, and 93

Figure 5-8 shows plots of leachate indicator parameters for monitoring locations 91, 92, and 93. Wells GW-91-1-1, GW-92-1-1, and GW-93-1-1 are located near the Hartland North Pad and are not Boundary Compliance Wells. Groundwater sampling at these locations commenced in March 2020. A groundwater sample was not collected from well GW-93-1-1 because it is no longer active.

In 2022/23, conductivity was generally higher at groundwater monitoring station 91 than previous years, increasing to maximum of $560 \mu \mathrm{~S} / \mathrm{cm}$ in March 2023. Hardness ( $\mathrm{as}_{\mathrm{CaCO}}^{3}$ ) was relatively high, ranging from 247 to $302 \mathrm{mg} / \mathrm{L}$. Chloride concentrations were below $5.2 \mathrm{mg} / \mathrm{L}$, and ammonia concentrations were below the detection limit. Nitrate concentrations were low, reaching $0.078 \mathrm{mg} / \mathrm{L}$ in December 2022, and sulphate concentrations were moderate, ranging from 63 to $78 \mathrm{mg} / \mathrm{L}$. Over the past five years, statistically significant increasing trends in conductivity and sulphate concentrations have been observed in well GW-91-1-1.

At groundwater monitoring station 92, conductivity was elevated and ranged from 406 to $580 \mu \mathrm{~S} / \mathrm{cm}$. Chloride and ammonia concentrations were low or below the detection limit. Nitrate concentrations were low, ranging from $0.067 \mathrm{mg} / \mathrm{L}$ in December 2022 to $0.204 \mathrm{mg} / \mathrm{L}$ in September 2022, and sulphate concentrations were elevated, ranging from 110 to $120 \mathrm{mg} / \mathrm{L}$. Over the past five years, statistically significant increasing trends in conductivity and sulphate concentrations have been observed in well GW-92-1-1.

At groundwater monitoring stations 91 and 92 , elevated conductivity and sulphate concentrations did not correlate with elevated nitrate concentrations, so it is unlikely that aggregate runoff impacted groundwater quality in this area of the landfill. Similarly, there are no signs of leachate impacts.


Figure 5-7. Groundwater Quality North of Phase 2 - Aggregate Impacts (2021 to Present)

### 5.7 Monitors South of the Phase 1 Landfill

Monitoring well locations $4,7,19,60,71,72,73$ and 85 are all located south of the Phase 1 landfill. Only wells at locations 4, 71,72 and 73 are considered Boundary Compliance Wells. Well GW-71-1-1 is located off site along Hartland Avenue. Wells at locations $4,19,60,71,72$ and 73 are multi-level nested monitoring wells, and the well at location GW-7-1-1 is a 37 m deep, open borehole that was used until 1989 for domestic water supply at the landfill. Wells at locations 71,72 and 73 were installed in 2003. Well GW-85-1-1 was installed in 2009 to replace well GW-3-2-1, which was decommissioned to permit construction of the bin facility.

Figure 5-9 shows plots of leachate indicator parameters in wells located south and downgradient of the landfill, and Figure 510 plots leachate indicator parameters for wells located southeast of the landfill. Like previous years, in 2022/23, all analytes in all groundwater sampled from wells south of the Phase 1 landfill met the applicable CSR standards.

Groundwater quality in these wells is sensitive to the performance of the south purge well system. When the purge well system is functioning properly, concentrations of leachate indicator parameters have been shown to decline, and seasonal fluctuations in concentrations of leachate indicator parameters are dampened. The reported concentrations indicate that the south purge well system successfully mitigated southward migration of leachate in 2022/23.

### 5.7.1 South Purge Wells (P1, P2, P3, P4 and P10)

Groundwater quality data collected from the south purge wells (P1, P2, P3 and P10) exhibited statistically significant increasing trends for multiple leachate indicator parameters over the past five years, indicating water quality degradation and therefore improved leachate capture. In 2022/23, the average conductivity of all groundwater samples collected from the south purge wells was $1,809 \mu \mathrm{~S} / \mathrm{cm}$, which was slightly higher than in the $1,784 \mu \mathrm{~S} / \mathrm{cm}$ observed in 2021/22.

In 2022/23, ammonia concentrations in the south purge wells exceeded the CSR AW standard throughout the entire monitoring year. Ammonia concentrations ranged from 23 to $91 \mathrm{mg} / \mathrm{L}$, and sulphate concentrations ranged from <1.0 to 4.1 $\mathrm{mg} / \mathrm{L}$. Chloride concentrations were relatively high, ranging from 39.0 to $220 \mathrm{mg} / \mathrm{L}$.

### 5.7.2 Monitoring Site 85

In 2022/23, at groundwater monitoring station 85 (not a Boundary Compliance Well), chloride concentrations remained elevated, ranging from 130 to $160 \mathrm{mg} / \mathrm{L}$. Conductivity increased throughout the monitoring year, reaching $770 \mu \mathrm{~S} / \mathrm{cm}$. Nitrate concentrations were generally consistent with historical ranges, fluctuating between approximately $1.21 \mathrm{and} 2.69 \mathrm{mg} / \mathrm{L}$. Sulphate concentrations were moderate and ranged from 34 to $47 \mathrm{mg} / \mathrm{L}$. Ammonia concentrations were typically just above the detection limit. Given the low ammonia concentrations, groundwater at this location is unlikely to be impacted by leachate. The slightly increase in nitrate, sulphate and conductivity may be related to runoff from aggregate placed during nearby construction works. The $\mathrm{Cl} / \mathrm{Na}$ molar ratio in 85-1-1 was greater than 1.5 , indicating an additional source of chloride other than road salt and leachate. In 2022/23, the average CI/Na molar ratios in south purge wells and Hartland Valve Chamber were calculated to be 0.67 and 0.52 , respectively. Over the past five years, a statistically significant decreasing trend in ammonia concentration has been observed at this location, indicating the south purge wells have remained effective.

### 5.7.3 Monitoring Site 60

Three monitoring wells are present at location 60: GW-60-1-1 (23 m BGS), GW-60-2-1 ( 16 m BGS), and GW-60-3-1 (7 m BGS). These wells are not considered Boundary Compliance Wells. In 2022/23, solute concentrations met applicable CSR standards in all three wells, but elevated chloride concentrations were observed in each well, and elevated nitrate concentrations were observed in GW-60-3-1. Chloride concentrations ranged from 75 to $160 \mathrm{mg} / \mathrm{L}$. Nitrate concentrations were below the detection limit in wells GW-60-1-1 and GW-60-2-1, but concentrations were higher near the surface, ranging from 1.85 to $2.6 \mathrm{mg} / \mathrm{L}$ in well GW-60-3-1. Conductivity values were elevated but slightly lower than in 2021/22, ranging from 465 to $850 \mu \mathrm{~S} / \mathrm{cm}$. Ammonia concentrations remained below the detection limit except for one sample collected from GW-60-3-1 ( $0.023 \mathrm{mg} / \mathrm{L}$ ). Sulphate concentrations were low to moderate, and increased toward ground surface, with concentrations ranging from 38 to $40 \mathrm{mg} / \mathrm{L}$ in GW-60-1-1, 56 to $60 \mathrm{mg} / \mathrm{L}$ in GW-60-2-1, and 59 to $71 \mathrm{mg} / \mathrm{L}$ in GW-60-3-1.

Over the past five years, statistically significant increasing trends in chloride concentrations and conductivity have been detected in all three wells at monitoring station 60. Moreover, an increasing trend in sulphate concentrations has been observed in well GW-60-3-1. Decreasing trends in ammonia and sulphate concentrations were observed in GW-60-1-1, and a decreasing trend in nitrate concentrations was observed in well GW-60-2-1. The CI/Na molar ratios observed in wells GW-60-1-1, GW-60-2-1, and GW-60-3-1 were well above 1 (i.e., $>2.0$ ), indicating an additional source of chloride other than road salt
and leachate. The highest $\mathrm{Cl} / \mathrm{Na}$ ratios ( $\sim 5$ ) were observed in well GW-60-2-1, where ammonia and nitrate concentrations were below detection. Overall, groundwater quality at monitoring station 60 degraded slightly in 2022/23, exhibiting relatively high conductivity and chloride concentrations, as well as elevated nitrate and sulphate concentrations near ground surface. Overall, landfill leachate likely did not impact groundwater at monitoring station 60 because ammonia concentrations were low, but the elevated conductivity, nitrate, and sulphate concentrations in GW-60-3-1 were consistent with aggregate runoff.

### 5.7.4 Monitoring Site 07

In 2022/23, two groundwater samples were collected from GW-07-1-0 (a Boundary Compliance Well). All solute concentrations were below applicable CSR standards. Sulphate and nitrate concentrations were low, and chloride concentration and conductivity were elevated. Chloride concentrations were $140 \mathrm{mg} / \mathrm{L}$ in May 2022 and $120 \mathrm{mg} / \mathrm{L}$ in October, whereas conductivity was $728 \mu \mathrm{~S} / \mathrm{cm}$ in May and $411 \mu \mathrm{~S} / \mathrm{cm}$ in October. The elevated conductivity readings occurred along with elevated nickel and cobalt concentrations, which ranged from 5.96 to $12.6 \mu \mathrm{~g} / \mathrm{L}$ and 3.05 to $6.35 \mu \mathrm{~g} / \mathrm{L}$, respectively. Additionally, iron and manganese concentrations were elevated, ranging from 1.87 to $6.30 \mathrm{mg} / \mathrm{L}$ and 2.06 to $2.66 \mathrm{mg} / \mathrm{L}$, respectively. Ammonia concentrations were relatively low, ranging from 0.016 to $0.056 \mathrm{mg} / \mathrm{L}$.

Over the past five years, statistically significant increasing trends in chloride concentration and conductivity have been observed at this location, as well as a decreasing trend in nitrate concentration. High $\mathrm{Cl} / \mathrm{Na}$ molar ratios (i.e., >4.4) indicate an additional source of chloride other than road salt. It is unlikely that landfill leachate impacted this area of the landfill because ammonia concentrations were low and nitrate concentrations were below detection limits. Similarly, sulphate concentrations were low, so it is not likely that aggregate stockpiling impacted groundwater at this location.

### 5.7.5 Monitoring Sites 71, 72, 73

Wells GW-71-1-1, GW-72-1-1, and GW-73-1-1 are located at or near the eastern landfill property boundary, south of Phase 1, and are considered Boundary Compliance Wells. Concentrations of leachate indicator parameters observed at monitoring stations 71, 72, and 73 are plotted in Figures 5-9 and 5-10.

In 2022/23, all solute concentrations observed at monitoring station 71 were below applicable CSR standards. Groundwater quality at this location was generally consistent with historical results, with conductivity values ranging from 216 to $340 \mu \mathrm{~S} / \mathrm{cm}$, and chloride concentrations below $10 \mathrm{mg} / \mathrm{L}$. Ammonia concentrations were typically near or below the detection limit, except for in one sample collected from GW-71-1-1 in January 2023 ( $1.24 \mathrm{mg} / \mathrm{L}$ ). Nitrate concentrations were below detection limits, except for well GW-71-3-1, where nitrate concentrations ranged from 0.82 to $1.34 \mathrm{mg} / \mathrm{L}$. Sulphate concentrations were generally below $35 \mathrm{mg} / \mathrm{L}$ in all three wells. Over the past five years, a statistically significant increasing trend in conductivity has been observed in all three wells, and an increasing trend in nitrate concentrations is present for well GW-71-3-1. Overall, groundwater quality at this location showed no evidence of landfill leachate impacts, but shallow groundwater may be mildly impacted by dilute aggregate runoff.

In 2022/23, all solute concentrations observed at location 72 were below applicable CSR standards. The groundwater exhibited slightly elevated conductivity values, ranging from 426 to $610 \mu \mathrm{~S} / \mathrm{cm}$, and ammonia and nitrate concentrations were low. Chloride and sulphate concentrations were moderate, ranging from 39.0 to $74.0 \mathrm{mg} / \mathrm{L}$ and 44 to $68 \mathrm{mg} / \mathrm{L}$, respectively. Over the past five years, a statistically significant increasing trend in conductivity has been observed in GW-72-1-1, and ammonia has shown a decreasing trend. Additionally, in well GW-72-3-1, conductivity and chloride concentrations have exhibited decreasing trends. Overall, groundwater quality at location 72 is good, with no evidence of landfill leachate or aggregate runoff impacts.

In 2022/23, all solute concentrations observed at location 73 were below applicable CSR standards. Conductivity was low to moderate, ranging from 269 to $525 \mu \mathrm{~S} / \mathrm{cm}$, and ammonia concentrations were generally at or below detection, except for in one groundwater sample collected from GW-73-3-1 in September 2022 ( $0.099 \mathrm{mg} / \mathrm{L}$ ). Average nitrate concentrations were $0.083 \mathrm{mg} / \mathrm{L}$ in GW-73-1-1 and $0.708 \mathrm{mg} / \mathrm{L}$ in GW-73-3-1, with the highest concentrations nearest ground surface. In March 2023, a maximum nitrate concentration of $1.18 \mathrm{mg} / \mathrm{L}$ was observed at GW-73-3-1, but sulphate concentrations remained low ( $30 \mathrm{mg} / \mathrm{L}$ ). Over the past five years, statistically significant increasing trends in conductivity and nitrate concentration have been observed in wells GW-73-1-1 and GW-73-2-1, and well GW-73-1-1 has exhibited an increasing trend in sulphate concentrations. Additionally, well GW-73-3-1 has exhibited and increasing trend in nitrate, and well GW-73-2-1 has exhibited a decreasing trend in ammonia. Groundwater quality at this location shows no evidence of landfill leachate impacts. Although nitrate concentrations were occasionally elevated, the low and stable sulphate concentrations indicate that groundwater quality was not impacted by aggregate runoff.


Figure 5-8. Groundwater Quality North of Hartland North Pad


Figure 5-9. Groundwater Quality South of Landfill

### 5.7.6 Monitoring Site 04

Location 04 is the southernmost groundwater monitoring location at the landfill and is considered a Boundary Compliance Well. Groundwater sampling in the deepest monitoring well at Site 04 was discontinued in 2016 due to extremely slow recharge and resultant challenges in collecting representative samples.

In 2022/23, all concentrations observed in GW-04-3-1 and GW-04-4-1 met the applicable CSR standards. Groundwater quality in GW-04-4-1 continued to report slightly elevated concentrations of chloride in 2022/23, with a maximum concentration of $43 \mathrm{mg} / \mathrm{L}$ during March 2023 sampling event. Conductivity and ammonia concentrations in both wells were generally low and consistent with previous years, indicating water quality was not impacted by leachate.

Nitrate concentrations were elevated in the shallow well, with an average concentration of $1.13 \mathrm{mg} / \mathrm{L}$, and a maximum of 1.84 $\mathrm{mg} / \mathrm{L}$. The maximum nitrate concentration observed in the deep well was $0.046 \mathrm{mg} / \mathrm{L}$. Sulphate concentrations were slightly higher in the deep well, ranging from 38 to $39 \mathrm{mg} / \mathrm{L}$ as compared to 21 to $32 \mathrm{mg} / \mathrm{L}$ in the shallow well. Over the past five years, statistically significant increasing trends in conductivity, chloride, and sulphate concentration have been observed in GW-04-31, and nitrate concentrations have trended downward. An increasing trend in chloride concentrations was also detected in GW-04-4-1. Due to the low ammonia and conductivity concentrations, groundwater quality in at monitoring station 04 was not impacted by landfill leachate. The source of the chloride remains uncertain but is likely linked to the same cause of water quality impacts at stations 07,85 and 60 . Although nitrate concentrations were occasionally elevated, the low and stable sulphate concentrations indicate that groundwater quality was not impacted by aggregate runoff.

### 5.7.7 Monitoring Site 19

Four monitoring wells were installed at monitoring station19: GW-19-1-1 (38 m BGS), GW-19-1-2 (28 m BGS), GW-19-2-1 ( 17 m BGS), and GW-19-2-2 ( 9 m BGS). These wells are not considered Boundary Compliance Wells. In 2022/23, all samples collected at station 19 met applicable CSR AW and DW standards.

Unlike previous years, conductivity in one well (GW-19-1-1) surpassed $500 \mu \mathrm{~S} / \mathrm{cm}$, reaching a maximum conductivity of 740 $\mu \mathrm{S} / \mathrm{cm}$ in March 2023. Ammonia and nitrate concentrations were generally close to detection limits, except for slightly elevated ammonia concentrations that ranged from 0.031 to $0.27 \mathrm{mg} / \mathrm{L}$ in wells $\mathrm{GW}-19-1-1$ and GW-19-1-2. Chloride and sulphate concentrations were generally low to moderate, ranging from 15.0 to $33.0 \mathrm{mg} / \mathrm{L}$ and $<1.0$ to $49 \mathrm{mg} / \mathrm{L}$, respectively. Over the past five years, statistically significant increasing trends in ammonia concentration and conductivity have been observed in GW-19-1-1, and a decreasing trend in ammonia concentrations has been observed in GW-19-1-2. Increasing trends in conductivity and chloride have been observed in well GW-19-2-1, and an increasing trend in chloride concentration has also been observed in GW-19-2-2. Similar to previous years, groundwater quality at monitoring station 19 was not impacted by leachate or aggregate placement.

### 5.8 Monitors East of the Phase 1 Landfill

Groundwater monitoring stations $16,17,18,50,54$ and 76 are situated along the east boundary of the Phase 1 landfill, north of Hartland Avenue. Groundwater quality is no longer monitored stations 50, 54, and 76 due to continued demonstration of groundwater flow toward the landfill, and the presence of water quality reflective of background conditions.

### 5.8.1 Monitoring Sites 17 and 18

Figure 5-11 displays concentrations of leachate indicator parameters observed at groundwater monitoring stations 17 and 18. In 2022/23, solute concentrations in all wells at both sites met applicable CSR standards. At both sites, the groundwater exhibited relatively low conductivity (average of $380 \mathrm{mg} / \mathrm{L}$ ), except for one sample collected from GW-18-1-1 in May 2022 $(1,646 \mu \mathrm{~S} / \mathrm{cm})$. The anomalous conductivity value does not correspond to a noticeable increase in any major ion concentrations or alkalinity, so it was likely a transcription error. Throughout the monitoring year, ammonia and chloride concentrations were low, and nitrate concentrations were generally low to moderate, ranging from below detection to 1.23 $\mathrm{mg} / \mathrm{L}$, with an average of approximately $0.26 \mathrm{mg} / \mathrm{L}$. Sulphate concentrations were low to moderate, ranging from 24 to 57 $\mathrm{mg} / \mathrm{L}$. Over the past five years, statistically significant increasing trends in conductivity, chloride, sulphate, and nitrate concentrations have been observed at GW-18-2-1, and an increasing trend in sulphate concentrations has been observed in GW-18-1-1 and GW-18-2-2. At monitoring station 17, decreasing trends in chloride concentration were observed in GW-17-1-1 and GW-17-1-2.

Groundwater quality at monitoring stations 17 and 18 did not show any evidence of landfill leachate impacts. Aggregate impacts were not observed at these locations, as nitrate and sulphate concentrations were relatively low and decreased slightly from the previous monitoring year.

### 5.8.2 Monitoring Site 16

Site 16 is located northeast of Phase 1 and is not considered a Boundary Compliance Station. In 2022/23, only two groundwater samples were collected at Site 16. All leachate indicator parameters were indicative of background conditions and met all applicable CSR standards. Average conductivity was relatively low at approximately $358 \mu \mathrm{~S} / \mathrm{cm}$. However, in March 2023, highly elevated nitrate and moderate sulphate concentrations were observed in all wells at monitoring station 16, when the average nitrate concentration was $14.5 \mathrm{mg} / \mathrm{L}$. Sulphate concentrations showed similar increases over this period, reaching $71.0 \mathrm{mg} / \mathrm{L}$ in GW-16-1-2. Over the past five years, a statistically significant increasing trend in chloride concentrations has been observed in GW-16-1-2, GW-16-2-1, and GW-16-2-2. An increasing trend in conductivity has been observed at GW-16-2-2, as well as a decreasing trend in sulphate concentration at GW-16-2-1.

In 2022/23, groundwater quality at Site 16 showed no evidence of landfill leachate impacts because chloride and ammonia concentrations were low. However, during the March 2023 sampling event, groundwater quality at Site 16 appeared to be influenced by aggregate, as indicated by significantly elevated nitrate levels and moderate sulphate concentrations. Groundwater quality at this location should continue to be closely monitored to evaluate the severity and evolution of groundwater quality impacts linked to the Northeast Stockpile.

### 5.9 Summary

Groundwater quality results from 2022/23 indicate that leachate-impacted groundwater was contained within the landfill property boundary. At the north end of the landfill, leachate impacts extend just north of the unlined Lower Leachate Lagoon and through the middle of the lined Upper Leachate Lagoon but did not extend off-site. Similarly, south of the landfill, leachateimpacted groundwater did not extend off-site. Leachate-related exceedances were confined to the landfill footprint on the east side of Phase 1 and are inferred to extend to the west side of the Phase 2 landfill.

In 2022/23, multiple wells exhibited groundwater quality impacts related directly to aggregate production, use, and stockpiling at Hartland. Compared to the 2021/22, groundwater quality impacts associated with aggregate production and stockpiling have become more widespread. Nested wells impacted by aggregate stockpiling indicate that nitrate concentrations in shallower wells were more evidently elevated and more frequently exceeded the threshold of $0.1 \mathrm{mg} / \mathrm{L}$, in comparison to sulphate levels.

Our review of the 2022/23 groundwater quality data revealed the following:

- Boundary Compliance Wells and off-site monitoring wells met CSR AW and DW standards, except for an anomalous copper exceedance at location 21 in May 2022. However, this recorded exceedance may not be an accurate representation of water quality as there was a high relative percent difference between parent and duplicate samples.
- CSR exceedances in groundwater were observed in on-site monitoring wells near the north and south purge wells and known leachate sources as follows:
- Similar to previous years, water quality in well GW-58-1-0 (within the landfill waste footprint) exceeded CSR AW standards for ammonia and cobalt on all four sampling dates. Additionally, chloride, cobalt, nickel, strontium, and vanadium concentrations exceeded CSR DW on one or more sampling events. Sodium concentrations were above the CSR DW of $200 \mathrm{mg} / \mathrm{L}$ but well below the regional background concentration of $1,700 \mathrm{mg} / \mathrm{L}$, indicating sodium concentrations were compliant with the CSR at this location.
- Groundwater quality in well GW-52-1-1 (near the north purge wells) exceeded CSR AW standards for ammonia and CSR DW standards for strontium on all four sampling dates.
- Groundwater quality in the north (GW-52-4-0-P7 and GW-80-1-0-P8) and south purge wells (P1, P2, P3, P4 and P10) exceeded CSR AW standards for ammonia on one or more sampling dates. Chloride, chromium, arsenic, barium, and nitrate concentrations occasionally exceeded CSR DW standards in some wells. Lithium and sodium concentrations in all wells were below regional background values, and therefore elevated concentrations were not considered CSR exceedances.
- Operation of the Phase 1 North Purge Well System continues to mitigate leachate impacts north of the landfill, as indicated by long-term stable or decreasing concentrations of leachate indicator parameters at stations 40, 20 and 21.

Groundwater quality in wells GW-20-1-1, GW-21-1-1, and GW-40-1-1 were generally stable in 2022/23, but nitrate concentrations at monitoring station 40 increased considerably from the previous monitoring year.

- Groundwater quality in some shallow wells located along the North Ridge (e.g., monitoring stations 77, 78, 87, 88), near the Northwest Stockpile exhibited elevated nitrate and sulphate concentrations. Groundwater quality was primarily impacted by aggregate stockpiling on the bedrock ridge north of the Phase 2 landfill.
- Along the northern edge of the Phase 2 basin, groundwater quality is primarily impacted by runoff from aggregate stockpiling, with some wells (GW-106-1-1 and GW-105-1-1) showing evidence of dilute landfill leachate impacts. The influence of landfill leachate on groundwater is limited to an area less than 20 meters northwest of the Phase 2 Basin and additional mitigation measures are being evaluated and implemented by CRD.
- Groundwater quality between the Lower Leachate Lagoon and Willis Point Road continued to show no indication of leachate impacts. However, the elevated conductivity, sulphate, and nitrate levels observed at Boundary Compliance Station 31 are likely associated with local aggregate use and stockpiling to the northeast of the landfill.
- Water quality along the southern boundary of the Phase 1 landfill showed no evidence of landfill leachate impacts. Since 2020, chloride concentrations have been occasionally elevated at some monitoring stations (e.g., stations 85 and 60) in the area, but the elevated chloride concentrations have not correlated with elevated ammonia concentrations. High $\mathrm{Cl} / \mathrm{Na}$ molar ratios $(>1)$ suggest there is additional source of chloride other than road salt, but the source of chloride is currently unknown. In addition, elevated nitrate concentrations were observed in some shallow wells, accompanied by moderate concentrations of sulphate. The elevated nitrate may be sourced from the Phase 1 Stockpile or dust originating from other stockpiles.
- Water quality along the east boundary of the Phase 1 landfill was similar to previous years and was not impacted by leachate. Ammonia concentrations at monitoring stations 16,17 and 18 were below the detection limit on all sampling events, and groundwater quality at met all applicable CSR standards. However, groundwater quality at station 16 likely showed signs of aggregate impacts during the March 2023 sampling event.


Figure 5-10. Groundwater Quality Southeast of Landfill


Figure 5-11. Groundwater Quality East of Landfill

## 6. Groundwater Quality in Domestic Wells

### 6.1 Monitoring Locations

This section of the report summarizes our interpretation of water quality data collected from domestic wells around Hartland Landfill in June 2023. Routine groundwater samples were collected June 2023, from sixteen (16) selected domestic wells located within a 2 km radius of the landfill.

Since the 1980s, the CRD has performed routine sampling and analysis of domestic wells near the landfill that are used as a primary source of drinking water. The number of wells included in the program have been reduced gradually as municipal water became available and residents chose to connect to the municipal supply system. Most of the domestic wells near Hartland Landfill are situated southeast of the landfill, as shown on Figure 6-1. The wells are primarily 0.15 m in diameter and penetrate between 30 m and 120 m of bedrock. Three of the wells are shallow dug wells completed in overburden. Well yields are generally low and substantial drawdown occurs during pumping, particularly during the dry summer months.

Routine groundwater samples collected in 2023 were analyzed for general water quality parameters and total metals. Tabulated results are presented in Appendix B.2. Results were compared to the British Columbia Approved Source Drinking Water Quality Guidelines (SDWQGs) where available and Guidelines for Canadian Drinking Water Quality (CDWQ). SDWQGs and CDWQ were updated in September 2020 and June 2019, respectively, with several parameters (e.g., sulphate, antimony, lead, cobalt, chromium, copper, manganese, nickel, etc.) added to the drinking water guidelines or updated.

### 6.2 Domestic Well Quality

Groundwater quality in the domestic wells was consistent with groundwater quality results reported since 2000. Although concentrations of some parameters have varied since the sampling program began, and pH in some of the wells was less than circumneutral, the groundwater is considered representative of natural conditions.

At all locations, the groundwater quality met applicable guidelines in all sampled wells, with some exceptions. Exceedances of the CDWQ guideline (aesthetic objectives) were noted at select locations as described below:

- Field pH values were below the CDWQ guideline of 7.0 , at monitoring stations $3,4,7,8,9,10,12,13$, and 14 .
- The manganese concentration observed in domestic wells 37 and 38 (stations 12 and 13 , respectively) exceeded the total manganese SWWQG ( $0.02 \mathrm{mg} / \mathrm{L}$ ). The SDWQG for total manganese in drinking water is an aesthetic objective to protect against staining and unpleasant taste but such concentrations are not considered toxic.

The five domestic wells located northwest of the landfill met CDWQ and SDWQG guidelines for leachate indicator parameters, indicating these wells were not impacted by leachate from Hartland Landfill.

In summary, domestic well water quality results are consistent with historical data and show no evidence of landfill-related impacts.


## 7. Surface Water Quality near the Landfill

### 7.1 Compliance Monitoring Locations

Five surface water compliance monitoring stations (i.e., Boundary Compliance stations) surround Hartland Landfill. These stations are concentrated along the southern and northern property boundaries, downgradient of areas that could potentially be impacted by leachate or runoff from the landfill. The following Boundary Compliance stations are monitored to assess landfill compliance with the landfill operating permit:

South of the Landfill

- Sw-S-04

North of Phases 1 and 2

- $\mathrm{Sw}-\mathrm{N}-05$
- $\quad \mathrm{Sw}-\mathrm{N}-16$


## North of the Hartland North Pad

- $\quad \mathrm{Sw}-\mathrm{N}-41 \mathrm{~s} 1$
- $\quad \mathrm{Sw}-\mathrm{N}-42 \mathrm{~s} 1$

In 2022/23, a total of 34 surface water stations were sampled, including nine newly established stations. Surface water quality analytical results are presented in Appendix B.3. Between December 2022 and February 2023, a total of nine (9) surface water stations (SW-N-57, SW-N-58, SW-N-59, SW-N-60, SW-N-61, SW-N-62, SW-N-63, SW-N-64, and SW-N-65) were established north of the landfill. These stations aim to monitor the effects on surface water quality due to various landfill activities, such as road construction, blasting, and aggregate stockpiling, and to assess their impact on the downstream environment.

### 7.1.1 Regulatory Comments

As discussed in Section 2.5.2, the results were compared to the BCWQG for the protection of freshwater aquatic wildlife (AW). Exceedances of BCWQGs are noted in Table 7-1. Some parameters have variable guidelines, as noted below:

- For ammonia, there is no single value for the protection of freshwater AW. The toxicity of ammonia is related to the temperature and pH of the water, and the BCWQG includes values for acute and chronic effects. The appropriateness of the chronic (allowable 30-day average) concentration for the assessment of ongoing operations is currently being evaluated. Using surface water quality data, CRD staff calculated the allowable LTC concentration of ammonia for the protection of freshwater AW and the short term acute (STA) based on the pH and temperature of each sample (as discussed in the footnotes of Table 7-1), with exceedances highlighted based on BCWQG-STAs.
- The BCWQG-STA for sulphate was calculated based on the hardness of each sample. A detailed description is presented in the footnotes of Appendix B.3.
- The BCWQG-STA and LTC values for total suspended solids (TSS) reference a "change from the background value" and the flow conditions (i.e., clear/turbid waters). Due to influences from aggregate stockpiling and ongoing construction, the water quality at stations Sw-N-CSs2 and Sw-N-14 cannot be considered representative of background conditions north of the landfill. Sw-S-52 is the only background station at the landfill and was used to assess TSS criteria. In 2022/23, background TSS values observed at Sw-S-52 were less than $1.0 \mathrm{mg} / \mathrm{L}$.
- BCWQGs for cadmium, lead, nickel, manganese, sulphate, and silver are hardness dependent. Total metal concentrations were compared to guidelines based on the hardness of each sample.
- The dissolved copper BCWQG varies with hardness, pH , dissolved organic carbon (DOC) and temperature, and is calculated using the Biotic Ligand Model (BLM). The calculated BLM results were obtained from CRD. However, copper exceedances should be interpreted with caution due to anomalously high DOC concentrations resulting from the use of compromised laboratory-supplied preservatives.
- The total aluminum table was updated in the approved BCWQG version dated August 2023. The total aluminum guideline is dependent on pH , hardness, and DOC and is calculated using the BC Water Quality Guideline calculator for Aluminum in Freshwater Ecosystems. Due to anomalously high DOC concentrations resulting from the use of compromised laboratory-supplied preservatives, the aluminum guideline is based on the 2021/22 average DOC concentration of $4.5 \mathrm{mg} / \mathrm{L}$ across the landfill.
- The total arsenic WQGs for freshwater and marine aquatic life were corrected in the approved BCWQG version from August 2023.
- The dissolved zinc table was updated in the approved BCWQG version from August 2023. The dissolved guideline is dependent on pH , hardness, and DOC. Due to anomalously high DOC concentrations resulting from the use of compromised laboratory-supplied preservatives, the aluminum guideline is based on the 2021/22 average DOC concentration of $4.5 \mathrm{mg} / \mathrm{L}$ across the landfill.






















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| sw.w05 | ${ }_{\text {fr1 }}$ | $r$ | ${ }^{20220504}$ |  | 0.0782 | 0.768 | 122 | ${ }^{43} 4$. | ${ }^{378}$ | 0.005 | 159. | 296 | ${ }_{526}$ | 0.587 | 0.005 | ${ }_{688}$ | 1400. |
| sw..05 | ${ }_{\text {fR2 }}$ | $\checkmark$ | ${ }_{202020504}$ |  | 0.082 | 0.77 | 125 | ${ }^{423}$ | ${ }^{27}$ | 0.0086 | 159. | 297 | 5.15 | 0.575 | 0.005 | 2.16 | 1700. |
| sw..065 | ss | $r$ | ${ }^{2022} 21.107$ |  | 0.0214 | 0.589 | 296 | ${ }^{33}$. | ${ }^{3} 1$ | 0.0072 | 7.92 | 0.74 | 1.47 | 2.11 | 0.005 | 4.16 | 13. |
| sw.nos | ${ }_{\text {fr1 }}$ | r | 202301.05 | Claar. coloumess | 0.023 | 0.132 | 481 | 617. | ${ }_{428}$ | 0.006 | ${ }^{38}$ | 282 | 243 | 1.65 | 0.005 | ${ }_{667}$ | 8.6 |
| sw.wos | ${ }_{\text {fr2 }}$ | $\checkmark$ | 202301.05 | Clara colouness. | 0.037 | 0.128 | 481 | ${ }^{62}$ | ${ }_{3}{ }^{3}$ | 0.051 | ${ }_{364}$ | ${ }^{283}$ | 238 | 1.73 | 0.005 | ${ }_{6.54}$ | 190. |
| sw..nos | ${ }_{\text {fR1 }}$ | $\checkmark$ | ${ }_{20330207}$ | Modeasat four, cear 8 colourtess. | 0.0414 | 0.134 | ${ }_{587}$ | ${ }^{230}$. | ${ }^{3}$. | ${ }_{0}^{0.0071}$ | ${ }_{276}$ | ${ }_{282}$ | ${ }_{1.56}$ | ${ }_{0} .324$ | 0.005 | ${ }^{277}$ | ${ }^{30}$ |
| sw.nos | ${ }_{\text {fr2 }}$ | r | 20230207 | Modeasteray, cearar cooumeses. | 0.181 | 0.14 | 609 | 282 | 3.8 | 0.096 | 295 | 294 | 1.65 | 0.306 | 0.005 | 3.49 | 94 |
| sw.N14 | ss |  | $2022 \cdot 11-10$ |  | 0.005 | 0.093 | 1.38 | 181. | 59 | 0.005 | 122 | ${ }_{0} 0375$ | 0.36 | 0.51 | .005 | 122 | 1500 |
| sw.N.14 | ss |  | ${ }_{202320213}$ | Modease foum, ciarer colouness. | 0.005 | 0.104 | 134 | ${ }^{176}$ | 7. | 0.005 | 122 | ${ }_{0} .386$ | 0.389 | 0.11 | 0.005 | 128 | 32. |
| sw.w. 4 | ss |  | ${ }_{20332221}$ |  | 0.05 | 0.066 | 0.965 | 174. | 64 | 0.005 | 0.927 | 0.386 | 0235 | 0.077 | 0.05 | 0.95 | 34 |
| sw.w 14 | ss |  | 20230320 |  | 0.005 | 0.736 | ${ }_{0} 985$ | 160. | 4.5 | 0.005 | 0266 | 0.368 | 0231 | 0.064 | 0.005 | 0.9 | 32 |
| sw.N.15 | ss |  | 20230221 |  | 0.05 | 0.011 | 0.625 | 782 | 1. | 0.005 | 0.188 | 0.28 | 0.09 | 0.45 | 0.005 | 0.16 | 1.3 |
| sw.N.15 | ss |  | ${ }^{20230320}$ | Hogh fou, cearana colouness. | 0.005 | 0.0108 | ${ }_{0} 223$ | ${ }^{124}$ | 14 | 0.005 | 0.101 | 0.24 | 0.022 | 0.04 | 0.005 | ${ }_{0} 020$ | ${ }_{1} 13$ |
| sw.N.16 | ${ }_{\text {fR1 }}$ | r | 202211.09 |  | 0.0257 | 0222 | 568 | 210. | 40.1 | 0.0018 | ${ }_{36} 8$ | 1.02 | ${ }^{128}$ | 0.152 | 0.005 | ${ }_{16} 8$ | 1000 |
| sw.N.16 | ${ }_{\text {fr2 }}$ | r | 2022.1109 | Sutumied tob bas as 20220032200 | 0.024 | 023 | 595 | 210 | ${ }_{39} 9$ | 0.008 | 37 | 103 | ${ }^{123}$ | 0.168 | 0.005 | 188 | 98. |
| sw.w.16 | ${ }_{\text {fr1 }}$ | r | 2023.10 .05 | Clara colounes. | 0.188 | 0.181 | 622 | 179. | 485 | 00073 | 36.5 | 088 | 1.1 | 0.199 | 0.005 | 109 | 7 |
| sw.N.16 | ${ }_{\text {fr2 }}$ | r | 20230105 | Clar, olouness. | 0.017 | 0.124 | 6.19 | 177. | ${ }_{502}$ | 0.075 | ${ }_{36} 6$ | 0.808 | 0.938 | 0.1 | 0.005 | ${ }^{11}$ | 7.1 |
| sw.w.16 | ${ }_{\text {fr1 }}$ | r | 20230207 |  | 0.0241 | 0295 | 807 | 12. | ${ }_{546}$ | 0.065 | ${ }_{38} 2$ | 0.721 | 143 | 0.115 | 0.005 | 164 | ${ }^{20}$ |
| sw.w.16 | ${ }_{\text {fr2 }}$ | $r$ | ${ }_{20230207}$ | Ver nigntou, modesatey wutad, boum. | 0.036 | 0298 | 823 | 175. | ${ }_{59} 9$ | 0.062 | 382 | 0.76 | 1.42 | 0.13 | 0.005 | 162 | 710. |
| sw.N17 | ss |  | $2022.11-10$ |  | 0.013 | 027 | 0.958 | ${ }^{288}$ | ${ }_{4}{ }^{4}$ | 0.0087 | 10.5 | 0.58 | 0.62 | 0.801 | 0.005 | ${ }_{5} .13$ | 100. |
| sw.w.17 | ss |  | 20230213 | Modeates Pou, carar coloumes. | 0.005 | 0.093 | 0.82 | 29. | 4. | 0.005 | 1.67 | 0.508 | 0222 | 0.151 | 0.005 | 0.91 | ${ }_{3} 3$ |
| sw. N W | ss |  | 20230221 |  | 0.05 | 0.085 | 0.757 | 219. | 58 | 0.005 | 32 | 0.515 | ${ }^{0237}$ | 0.108 | 0.005 | ${ }^{1.18}$ | ${ }^{3}$ |
|  | ss |  | 20230320 |  | 0.05 | 0095 | 0.98 | 197. | 6.1 | 0.0108 | 265 | 0.49 | 029 | 0.097 | 0.005 | 1.09 | 26 |
| sw.N.18 | ss |  | 20220.504 |  | 0.038 | 825 | 1.5 | 255. | ${ }_{66} 6$ | 0.154 | ${ }_{63}$ | ${ }_{131}$ | ${ }^{18,6}$ | 0.32 | 0.008 | ${ }_{16} 16$ | 1800. |
| sw.N.18 | ss |  | ${ }^{2022} 21.107$ |  | 0.085 | 0.984 | 104 | ${ }_{427}$ | ${ }_{42}$ | 0.098 | ${ }_{50} 5$ | 3. | ${ }_{283}$ | 3. | 0.005 | 1.02 | 99. |
| sw.w 18 | ss |  | 2023.9 .05 | Clas, columeses. | 0.0059 | 0289 | 0.9 | 189. | 193 | 0.099 | ${ }^{5} 5$ | 1.71 | 1.02 | 0.664 | 0.005 | ${ }^{378}$ | 0.91 |
| sw.w 18 | ${ }_{\text {fr1 }}$ |  | 20230207 |  | 129 | 0.788 | ${ }^{27}$ | ${ }^{32}$. | 171 | 0.085 | ${ }_{402}$ | 1.68 | 271 | 1.01 | 0.005 | 4.3 | 20. |
| sw.N.18 | ${ }_{\text {fr2 }}$ |  | ${ }_{20332027}$ |  | 0.004 | 0.63 | 112 | ${ }_{32} 8$. | 40.5 | 0.026 | ${ }_{38} 2$ | ${ }_{1.72}$ | 154 | 1.01 | 0.05 | 1.7 | 23. |
| sw.w 19 | ss |  | 20220.504 |  | 0.0141 | 0.41 | 156 | 181. | ${ }^{254}$ | 0.0062 | ${ }^{18}$ | 1.02 | 282 | 0.125 | 0.005 | 1.9 | 1500 |
| sw.*.19 | ss |  | 2027.100 |  | 0.0084 | 0.378 | 124 | ${ }^{226}$ | 506 | 0.0091 | 16.2 | 12 | 1.5 | 0228 | 0.05 | 2.19 | ${ }_{60} 6$. |
| sw.N.19 | ss |  | ${ }_{2023} 20.105$ | Clara coomenes. | 0.134 | 0.191 | 12 | 20. | 21.8 | 0.0098 | 4.92 | 1.15 | 1.83 | 0.137 | 0.005 | 3.4 | 82 |
| sw.N.4181 | ${ }_{\text {fr1 }}$ | r | 202301.05 | Clasar.olouness. | -0.05 | 0.046 | 0263 | 168. | ${ }^{261}$ | 0.005 | ${ }_{33} 5$ | 026 | 0.083 | 0.134 | 0.005 | 0.54 | 2. |
| SWWN4151 | ${ }_{\text {fr2 }}$ | r | 20230905 | Claer.olouness. | 0.05 | 0.045 | ${ }_{0} 287$ | 165. | ${ }^{27.7}$ | 0.05 | ${ }_{33} 2$ | 0.259 | ${ }_{0} 0.87$ | 0.124 | 0.005 | 0.54 | 1.9 |
| sw.wnes1 | ${ }_{\text {fr1 }}$ | r | 20230208 | Lontou, dears colounes. | -005 | 0.0541 | 0.172 | 19. | 27. | - 0.005 | 44.6 | 0.306 | 0.09 | 0.095 | 0.005 | ${ }^{1.13}$ | 180. |
| Sw.wnis | ${ }_{\text {fr2 }}$ | $r$ | ${ }_{202320208}$ | Loutou, | 0.005 | 0.553 | 0.82 | 181 | 27.5 | 0.058 | 44.5 | 0.323 | 0.1 | 0.087 | 0.005 | 1.4 | 8. |
| Sw-W2s1 | ${ }_{\text {fR1 }}$ | r | ${ }^{20220.5094}$ |  | 0.005 | 0.0511 | ${ }_{0}, 386$ | 186. | 16.4 | 0.005 | 193 | 0.508 | 0.108 | 0.04 | 0.005 | 0.4 | 1900 |
| SWWN/281 | ${ }^{\text {fr2 }}$ | $\checkmark$ | 20220504 |  | 0.005 | 0.054 | ${ }_{0} 933$ | 188 | ${ }_{132}$ | -0.005 | 191 | 0.527 | 0.113 | 0.04 | 0.005 | 0.44 | 220. |
| sw.wnes | ${ }_{\text {fr1 }}$ | $\checkmark$ | 2022.1109 |  | 0.0057 | 00427 | 0.421 | ${ }^{251}$. | 94 | 0.015 | 708 | 0.628 | 0.159 | 0.079 | 0.005 | ${ }_{1} 13$ | 30. |
| sw.wnest | ${ }_{\text {fr2 }}$ | $\checkmark$ | 2022.100 | Sumplead tabe sas 202.2032799 | -0.05 | 0.045 | 0.48 | 250 | ${ }^{85}$ | 0.001 | 6.74 | 0.654 | 0.164 | 0.089 | 0.005 | 1.65 | 1700. |
| sw..4281 | ${ }_{\text {fr1 }}$ | $\checkmark$ | 20230.05 | Clar, columes. | 0.05 | 0.097 | 0.508 | 188. | 83 | 0.062 | ${ }_{6.75}$ | 0.489 | 0.11 | 0.072 | 0.005 | 0.61 | 24. |
| SWWN(22s1 | ${ }_{\text {fr2 }}$ |  | 202301.05 | Clas, colouness. | 0.005 | 0.981 | 0.599 | 188. | 82 | 0.064 | 679 | 0.492 | 0.133 | 0.072 | 0.005 | 0.57 | 25. |
| sw.N45 | ss |  | 2022.11 .10 |  | 0.014 | 0.408 | ${ }_{148}$ | ${ }^{39}$ | 25 | 0.012 | 6.88 | 0.885 | 0.702 | 1.15 | 0.005 | 1.82 | 1700. |
| sw.N45 | ss |  | 20230273 | Modeateat ouv, cear cocouness. | O.0089 | 0.11 | ${ }_{178}$ | 24. | ${ }_{34}$ | 0.005 | 5.11 | 0.979 | 0.479 | 0.316 | 0.005 | 089 | 42 |
| sw.N45 | ss |  | 20230221 |  | 0.0069 | 0.108 | 1.44 | 23. | 6.5 | 0.007 | 92 | 0.975 | 0.438 | 0283 | 0.005 | 1.09 | 32 |
| sw.N.45 | ss |  | 20230320 | Lownom, coearatac colueless | 0.0074 | 0.1 | 1.16 | 205. | 4. | 0.065 | 3.1 | 0873 | 0324 | 021 | 0.005 | 081 | 25 |
| sw.n.s. | ss |  | 202320207 |  | 0.078 | ${ }^{128}$ | 413 | ${ }^{128}$ | 8800 | 245 | ${ }_{337}$ | 142 | 9,12 | 0.19 | 0.0157 | 416 | 34. |
| sw.Ns ${ }^{\text {d }}$ | ss |  | 20220.504 |  | 0.005 | 0213 | 13. | 169. | 46. | 0.05 | ${ }_{38} 6$ | 0.45 | 1.31 | 0.186 | 0.005 | 0.83 | 1500. |
| sw.Ns4 | ss |  | $2022 \cdot 11-10$ |  | 0.014 | 1.03 | 753 | 27. | 22. | 0.0165 | 361 | 0.83 | 249 | 0254 | 0.005 | 0.92 | 1800. |
| sw.N.s4 | ss |  | 202301.05 | Clasa, colounses. | 0.0545 | 577 | 20.1 | 220. | ${ }_{53}{ }^{3}$ | 0.054 | 3500 | 1.45 | ${ }_{907}$ | 0.372 | 0.007 | 645 | 61. |
| sw.N64 | ss |  | 20230207 | Hegh foew Moderatey mutidstandy yey. | 0.0206 | 0.958 | ${ }_{6} 3$ | ${ }^{225}$. | 50. | 0.029 | 477. | 0.786 | ${ }^{287}$ | ${ }_{0} 221$ | 0.005 | 191 | 1100 |
| sw.Ns4 | ss |  | 20230207 |  | 0.0206 | 0.958 | ${ }_{6} 64$ | ${ }^{225}$. | 50. | 0.029 | 477. | 0.736 | ${ }_{287}$ | 022 | 0.005 | 191 | 1100. |
| sw.w.s7 | ${ }_{\text {fr1 }}$ |  | 20230207 | Fied chasaesisists soteseoveded | 0.006 | 0.64 | ${ }_{1.12}$ | 29. | 17.2 | 0.005 | ${ }^{294}$ | 1.74 | ${ }_{123}$ | 1.08 | 0.005 | ${ }_{128}$ | 28. |
| sw.w.s7 | ${ }_{\text {fr2 }}$ |  | ${ }_{20332027}$ |  | 0.0057 | ${ }_{0.652}$ | 1.04 | 292 | 16.2 | 0.061 | 304 | 1.78 | 1.31 | 1.04 | 0.005 | 1.1 | ${ }^{3} 3$. |
| sw.ent | frm |  | ${ }_{20230207}$ | Menor taupleases | 0.0055 | 0.46 | 1.08 | 29. | 16.7 | 0.0055 | ${ }^{29} 9$ | ${ }_{1.765}$ | ${ }_{127}$ | 1.05 | 0.005 | ${ }_{1.19}$ | ${ }^{35} 5$ |
| sw.N62 | ss |  | 20230207 |  | 0.0054 | 0.598 | 0.932 | 20. | ${ }_{5} 5$ | - 0005 | 24. | 1.69 | ${ }_{1.16}$ | 1.07 | 0.005 | 0.54 | ${ }^{69}$ |
| sw.N62 | ss |  | 20230207 |  | 0.0054 | 0.59 | 0.932 | 20. | 5.5 | - 0005 | 24. | 1.69 | 1.16 | 1.07 | 0.005 | 0.54 | 69. |
| swwwes | ss |  | ${ }_{202320213}$ | NEN. Moteasal fou, ceara coouruss | -0.05 | 0.0984 | 1.02 | ${ }^{872}$ | ${ }^{16}$ | 0.025 | 13.1 | 0228 | ${ }^{0.337}$ | 0.07 | 0.005 | ${ }_{0} 0.3$ |  |
| sw.wes | ss |  | 20230221 | Low four, lear and coloutess. | 0.05 | 0.118 | 0.98 | ${ }^{89} 2$ | ${ }_{392}$ | 0.027 | 18.5 | 0278 | ${ }^{0.37}$ | 0.082 | 0.005 | 0.45 | 32 |
| sw.w63 | ss |  | 20230320 |  | 0.005 | 0.138 | 0887 | ${ }_{8} 87$ | ${ }_{98}{ }^{4}$ | 0.026 | 25.5 | 0.27 | 0378 | 0.071 | 0.005 | 0.73 | 3. |
| sw.N64 | ss |  | ${ }_{202329213}$ | NeN. Modeasel foum.dears colouress. | 0.05 | 0.11 | 122 | ${ }_{50} 9$ | 186. | 0.088 | 11.1 | 0.239 | 0.58 | 0.046 | 0.005 | 1.7 | ${ }^{62}$ |
| sw.N64 | ss |  | 20230320 |  | 0.005 | 0.154 | 1.14 | ${ }_{54} 4$ | 20. | 0.087 | 24. | 0.238 | 0.623 | 0.92 | 0.005 | 1.43 | ${ }_{6} 6$ |
| sw.N65 | ss |  | 20230213 |  | 0.05 | 0.102 | ${ }^{1.16}$ | ${ }^{627}$ | ${ }^{143}$. | 0.069 | ${ }^{11.4}$ | 0.251 | 0.487 | 0.088 | 0.005 | 1.35 | ${ }_{5} 5$ |
| sw.w.6s | ss |  | 20230320 |  | 0.005 | 0.154 | 0.98 | ${ }_{65} 5$ | 187. | 0.056 | ${ }_{25} 9$ | 0.24 | 0.53 | 0.05 | 0.005 | ${ }_{1} 13$ | ${ }_{53}$ |
| Sw.w.cs2 | ss |  | 20230105 | Claer, columeses. | 0.005 | 0.018 | 0.388 | 57. | 1. | 0.005 | -004 | 0.091 | 0.031 | 0.082 | 0.005 | 0.9 | 1.8 |
| Sw.w.cs2 | ss |  | ${ }_{202302088}$ | Lonfou, clear 8 coomueses. | 0.05 | 0.0136 | 0.16 | ${ }^{85} 4$ | 1. | 0.005 | - 0.05 | 0.073 | 0.025 | 0.084 | 0.005 | 025 | 30. |
| sw.so3 | ss |  | 202200505 |  | 0.0125 | 0.358 | 5.17 | ${ }^{22} 1$ | 30. | 0.0315 | 124. | ${ }_{0} 028$ | 1.04 | 0.063 | 0.005 | ${ }_{73} 3$ | 50. |
| sw.s.03 | 5 |  | ${ }_{2022,1.108}$ |  | 0.157 | 0284 | 608 | 180. | 469 | 0.064 | ${ }_{23} 3$ | 0.36 | ${ }^{1.19}$ | 0.178 | 0.01 | 32 | 1800. |
| sw.s.03 | ss |  | 20230104 | Clear, coluruss. | 0.0098 | 0.478 | 507 | 110. | ${ }_{52}$. | 00138 | 199. | 0.144 | 12 | 0.07 | 0.005 | ${ }^{3} 4$ | 49 |
| sw, 03 | ss |  | ${ }_{202302008}$ | Hoghow vey wubd. ver ger | 0.014 | 0.357 | 408 | 79.5 | 147. | 0.028 | 123. | 0.21 | 0.098 | 0.078 | 0.005 | 892 | 36. |
| sw. 904 | ${ }_{\text {fr } 11}$ | r | ${ }_{202020505}$ |  | 0.005 | 0.979 | 278 | ${ }_{76} 7$ | ${ }^{2} 0$ | 0.949 | 215 | ${ }_{0} 024$ | 0.52 | 0.94 | 0.005 | 255 | 1700. |
| sw.s.04 | fr2 | r | 202200505 |  | 0.0052 | 0.0964 | 275 | ${ }^{664}$ | 51. | 0.057 | 2.17 | 0272 | 0.43 | 0.044 | 0.005 | 248 | 1900 |
| sw. $0^{4} 4$ | fr1 |  | 040012023 | Clear, coumeses. | 0.005 | 0.504 | 1.81 | ${ }^{897}$ | 133 | 0.005 | 0.186 | 0.127 | 0.3 | 0.07 | 0.005 | 326 | ${ }_{37}$ |
| sw, 04 | ${ }_{\text {erm }}$ | $r$ | ${ }^{2023} 20208$ | Menor fotureatas | 0.0102 | 0.758 | 1.715 | 1225 | 10865 | 0.098 | 0.74 | 03005 | 0.335 | 0.035 | 0.005 | 179 | ${ }^{8865}$ |
| sw.s.12 | ss |  | ${ }_{20220505}$ |  | 0.0213 | 0.86 | ${ }^{7} 78$ | 106. | 1820. | 0.033 | ${ }^{356}$ | 0217 | 1.94 | 0.078 | 0.005 | ${ }_{567}$ | 74. |
| sw.s.12 | ss |  | 2027.108 |  | 0.042 | 0238 | 195 | 208. | ${ }^{58} 8$ | 0.052 | 152 | 0294 | 4.86 | 0274 | 0.005 | 405 | 1300. |
| sw.s.12 | ss |  | 20230104 | Clearas stanty gey. | 0.0151 | 0.98 | ${ }_{6} 5$ - | 122 | 1380. | 0.015 | 40. | 0.087 | ${ }_{188}$ | 0.101 | 0.005 | ${ }_{39}$ | 59 |
| sw.s.12 | ss |  | 20230208 | Heghow ver wubd. ver ger | 0.0202 | 0.74 | 6.08 | ${ }_{19} 1$ | 1830. | 0021 | ${ }_{36} 8$. | 0.158 | 201 | 0.081 | 0.005 | 604 | 7. |
| Sw, 20 | 5 |  | 20230104 | claer, counues. | 0.005 | 0.111 | 1.04 | ${ }_{48} 8$ | ${ }^{88} 1$ | 0.013 | ${ }^{238}$ | 0.088 | 0.165 | 0.089 | 0.05 | 681 | 42 |
| sw. s2 $^{\text {2 }}$ | ss |  | 2030104 | Clear.coumess. | 0.005 | 0.326 | 0.989 | 40. | 10.8 | 0.005 | 0.095 | 0.098 | 0.15 | 0.93 | .005 | 7.8 | 4. |
| sw. 2 24 | ss |  | 20230104 | clear coloueses. | 0.006 | 0.073 | 247 | 122. | 18.6 | 0.0077 | 1.54 | 0.168 | 0.44 | 0.092 | 0.005 | 3.1 | ${ }_{38}$ |
| sw, ${ }^{\text {27 }}$ | ss |  | 20220.505 |  | 0.07 | 0.75 | 0.83 | ${ }^{977}$ | 554 | 0.037 | ${ }_{56}$ | 0.185 | 0.42 | 0.04 | - 0005 | 244 | то. |

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### 7.3 Assessment of Surface Water Quality Impacts

The primary causes of any surface water quality degradation at the site include landfill leachate, road salt, and aggregate production, stockpiling, and use. Magnesium lignosulphonate has been used for dust suppression on gravel roads since 2014, so that may contribute some nitrate, nitrite, or ammonia to the surface water system. Background conductivity in surface water is typically below $200 \mu \mathrm{~S} / \mathrm{cm}$ but has been observed at higher concentrations at some locations during periods of low flow, or after periods of prolonged dry weather. Background ammonia concentrations are typically below $0.01 \mathrm{mg} / \mathrm{L}$, but occasionally exceed $0.1 \mathrm{mg} / \mathrm{L}$ downgradient of wetland areas. Background chloride concentrations are typically below $10 \mathrm{mg} / \mathrm{L}$. Background sulphate concentrations are typically below $10 \mathrm{mg} / \mathrm{L}$ but often increase in the downstream direction as streams receive groundwater discharge.

Surface water is considered impacted by leachate when electrical conductivity surpasses $500 \mu \mathrm{~S} / \mathrm{cm}$, ammonia concentrations exceed $0.5 \mathrm{mg} / \mathrm{L}$ and chloride concentrations exceed $20 \mathrm{mg} / \mathrm{L}$. Peak concentrations at surface water monitoring stations are typically observed during the dry summer and early fall months, when flows are low and there is limited dilution by precipitation.

Surface water is considered impacted by aggregate production, stockpiling or site construction works, when sulphate concentrations exceed $75 \mathrm{mg} / \mathrm{L}$, and ammonia or nitrate (from blasting residuals) are present at concentrations above background levels ( $0.1 \mathrm{mg} / \mathrm{L}$ ). Peak concentrations are typically observed during the first sampling event following the onset of wet weather in the fall months, and in areas of active quarry development, aggregate stockpiling or aggregate placement.

Surface water is considered impacted by road salt when electrical conductivity and chloride concentrations exceed background levels ( $300 \mu \mathrm{~S} / \mathrm{cm}$ and $20 \mathrm{mg} / \mathrm{L}$, respectively), and concentrations of ammonia and its degradation products (primarily nitrate) are below background levels ( $0.1 \mathrm{mg} / \mathrm{L}$ ). Sites that are judged to be impacted by road salt must also be located downstream (or downgradient) of surfaces and roadways where road salt was applied. Electrical conductivity and chloride concentrations typically peak following cold weather periods when de-icing salt is applied to roadways.

Surface water quality in natural and anthropogenically modified systems can be highly variable due to temporal variations in flow and chemical inputs, and professional judgement is used to determine the nature and degree of any impacts due to leachate, road salt and aggregate stockpiles. The authors are hydrogeologists and geochemists with considerable experience evaluating leachate and water quality impacts at other landfills and industrial facilities in coastal regions of British Columbia. If any surface water concentrations exceed BCWQG for the protection of freshwater at the property boundary, standard CSR protocols for notification of affected property owners should be followed.

### 7.4 Data

Hartland landfill is located within the Tod Creek watershed on a drainage divide between the Heal Creek drainage basin to the north and the Killarney Creek basin to the south. Surface water from both the Heal Creek drainage basin and Killarney Creek basin flow into Tod Creek, ultimately discharging to Tod Inlet. Surface water sampling stations are shown on Figure 7-1. Surface water sampling stations have been established on the landfill property to monitor compliance at the property boundary and identify changes in surface water quality that could be related to landfill operations. Surface water quality monitoring stations located off-site are used to monitor surface water quality around the landfill.

Surface water quality data was collected at eight (8) locations south and west of the landfill, 21 locations north of the Phase 1 and Phase 2 landfill areas, and five (5) locations northwest of the landfill (Hartland North Pad). Surface water samples are collected four times per year from property Boundary Compliance and other key stations and twice per year from all other monitoring stations. Samples were not collected when watercourses were dry.

In December 2022, a total of five (6) surface water stations (SW-N-57, SW-N-58, SW-N-59, SW-N-60, SW-N-61 and SW-N-62) were established within the footprint of the Phase 2 landfill and quarry. The purpose of these stations is to monitor surface water quality impacts related to landfill activities, including road construction, blasting, and aggregate stockpiling. Many of these stations were either dry or only yielded a single set of data.

In February 2023, SW-N-63, SW-N-64, and SW-N-65 were established near the confluence of Durrance Creek and Heal Creek to assess and delineate any water quality impacts related to aggregate in the receiving environment.

Surface water sampling points utilized in the 2022/23 monitoring program are summarized below:

## South and West of the Landfill

- Sw-S-03
- Sw-S-04 - compliance
- Sw-S-12
- Sw-S-20
- Sw-S-21
- Sw-S-24
- Sw-S-27
- Sw-S-52

North of the Hartland North Pad

- Sw-N-41s1 - compliance
- $\quad \mathrm{Sw}-\mathrm{N}-41 \mathrm{~s} 3$
- $\quad \mathrm{Sw}-\mathrm{N}-41 \mathrm{~s} 4$
- Sw-N-42s1-compliance
- Sw-N-CS2

North of Phase 1 and Phase 2

- Sw-N-05-compliance
- $\mathrm{Sw}-\mathrm{N}-14$
- $\mathrm{Sw}-\mathrm{N}-15$
- Sw-N-16 - compliance
- $\mathrm{Sw}-\mathrm{N}-17$
- $\quad \mathrm{Sw}-\mathrm{N}-18$
- $\quad \mathrm{Sw}-\mathrm{N}-19$
- $\mathrm{Sw}-\mathrm{N}-45$
- $\mathrm{Sw}-\mathrm{N}-50$
- $\mathrm{Sw}-\mathrm{N}-51$
- $\mathrm{Sw}-\mathrm{N}-53$
- $\mathrm{Sw}-\mathrm{N}-54$
- $\mathrm{Sw}-\mathrm{N}-57$
- $\mathrm{Sw}-\mathrm{N}-58$
- $\mathrm{Sw}-\mathrm{N}-59$

North of Phase 1 and Phase 2

- $\mathrm{Sw}-\mathrm{N}-60$
- $\mathrm{Sw}-\mathrm{N}-61$
- $\mathrm{Sw}-\mathrm{N}-62$
- Sw-N-63
- $\mathrm{Sw}-\mathrm{N}-64$
- $\mathrm{Sw}-\mathrm{N}-65$


### 7.5 Overview of Surface Water Exceedances

Surface water results were compared to the approved and working BCWQG AW criteria in Appendix B.3. A summary of surface water exceedances is presented in Table 7-1. In 2022/23, surface water quality data met the BCWQG-STA for all parameters except for TSS, total and dissolved iron, nitrite, and nitrate. Iron exceedances mostly reflect low flow and/or turbulent conditions, whereas nutrient exceedances are interpreted to reflect impacts due to runoff from the Northeast Stockpile and the Northwest Stockpile.

In 2022/23, surface water at the compliance location (SW-N-05) continued to show elevated and non-compliant nutrient concentrations at SW-N-05. On two out of the four sampling events, nitrate concentrations surpassed the BCWQG AW threshold of $32.8 \mathrm{mg} / \mathrm{L}$, and reached a peak concentration of $50.5 \mathrm{mg} / \mathrm{L}$. Runoff from the aggregate stockpiles situated in the Toutle Valley and on the bedrock ridge north of the Phase 2 landfill are believed to be the primary contributors of the nitrate and sulphate detected at SW-N-05 (AECOM 2023).

### 7.6 Surface Water Quality North of Phase 1

Clean runoff from the Phase 1 closure and the eastern perimeter of the landfill is directed to the NWSP and then into the wetland located north of the lower leachate lagoon. The wetland discharges northward to Heal Creek. Heal Creek also receives drainage from the area north of Phase 2 (High Level Road ditch), and from a small stream draining a small wetland below the east end of the Hartland North Pad. Aggregate and piping were added to select sections of the High Level Road ditch in June 2018 to protect the ditch from nearby construction and mitigate ponding, which may contribute to elevated concentrations of sulphate in surface drainage over time.

Heal Creek flows north-easterly from the NWSP at the north end of the Phase 2 Basin to the confluence with Durrance Creek, as shown in Figure 4-1. Durrance Creek discharges to Tod Creek, which in turn discharges to Tod Inlet, about 3 km north of the landfill. Heal Creek is a small creek with a watershed area of 128 ha. Heal Creek is mainly steep and rocky although the creek passes through a few small wetlands near the upper end. The creek dries up during the summer months except in the lower reaches where groundwater discharge maintains flow year-round.

Figures 7-2 and 7-3 present solute concentration and conductivity trends observed at surface water monitoring stations in the north area of the landfill and beyond the northern property boundary. In Figure 7-2, the distance of each station from the landfill boundary is shown in brackets in the legend. Sw-N-05 and Sw-N-16 are compliance monitoring stations. Elevated and noncompliant nitrate concentrations were observed at Sw-N-05. In addition, nitrate concentrations exceeded the BCWQG-LTC at Sw-N-05 and Sw-N-16 during all sampling events. Nitrite and ammonia concentrations exceeded BCWQG-LTC at Sw-N-05 on two out of four sampling dates. The water quality impacts observed at $\mathrm{Sw}-\mathrm{N}-05$ and $\mathrm{Sw}-\mathrm{N}-16$ are linked to runoff from the Northwest Stockpile. Similar water quality impacts were observed at Sw-N-17 and Sw-N-14. Like previous years, these sampling locations were dry periodically in 2022/23 due to arid summer conditions.


### 7.6.1 Monitoring Site Sw-N-16

A small wetland is located just north of the lower leachate lagoon. In this wetland, surface water flows north and through a weir at Sw-N-16 (Boundary Compliance Station) before discharging into a culvert under Willis Point Road and then to Heal Creek. Sw-N-16 is located on the landfill property and is the compliance point used to monitor the quality of the surface water leaving the landfill through this route.

In 2022/23, all parameters met BCWQG-S guidelines at Sw-N-16 except for total iron concentration during one sampling event in February 2023. Similar to 2021/22, nitrate and sulphate concentrations were elevated, and nitrate concentrations exceeded the BCWQG-LTC on all sampling dates. Nitrate concentrations ranged from 6.80 to $8.28 \mathrm{mg} / \mathrm{L}$ from December 2022 through February 2023. Sulphate concentrations ranged from 36.0 to $110 \mathrm{mg} / \mathrm{L}$. Ammonia and chloride concentrations were low, and conductivity was moderate, ranging from 237 to $386.5 \mu \mathrm{~S} / \mathrm{cm}$. Given the low ammonia and chloride concentrations, it is unlikely that surface water was impacted by leachate. However, the elevated nitrate and sulphate concentrations are characteristic of aggregate impacts, which may be related to nearby road building and hauling of aggregate to the Northeast Stockpile. Surface water quality at this location should continue to be monitored to assess the severity and evolution of these impacts.

The iron exceedance observed at Sw-N-16 was likely due to the disturbance of sediment during sampling, as indicated by the reported high TSS concentrations. Alternatively, elevated iron concentrations may be derived from the wetland located upgradient of $\mathrm{Sw}-\mathrm{N}-16$ or related to sediment accumulation at sampling locations. Dissolved copper concentrations were above the BCWQG-LTC guideline on two of four sampling events, corresponding with relatively lower pH (i.e., <7.0).

### 7.6.2 Monitoring Site Sw-N-05

Another route for surface water to leave the property to the north is through the main channel of Heal Creek, located just north of Phase 2 at Sw-N-05 (Boundary Compliance Station). During a hydrogeochemical site investigation in 2022, surface water quality was evaluated at location Sw-N-05, and AECOM determined that elevated metal and nutrient concentrations observed at Sw-N-05 were influenced by aggregate production/stockpiling and dilute leachate from the northwest corner of Phase 2 . In 2022/23, surface water samples were collected on all four sampling dates at Sw-N-05.

Throughout the monitoring year, surface water quality at $\mathrm{Sw}-\mathrm{N}-05$ continued to degrade, with elevated conductivity, nutrient and sulphate concentrations. All parameters other than nitrate ( $50.5 \mathrm{mg} / \mathrm{L}$ ) were compliant with the BCWQG-STA. Nitrate and dissolved copper concentrations exceeded the BCWQG-LTC on all sampling dates, ammonia and nitrite concentrations exceeded the BCWQG-LTC on two sampling dates (May and November 2022). TSS, selenium, and sulphate concentrations exceeded the BCWQG-LTC on one sampling date.

Nitrate concentrations were highly elevated, ranging from $13.6 \mathrm{mg} / \mathrm{L}$ to $50.5 \mathrm{mg} / \mathrm{L}$, and were accompanied by sulphate concentrations ranging from 150 to $470 \mathrm{mg} / \mathrm{L}$. Conductivity in the surface water was high, ranging from 460 to $860 \mu \mathrm{~S} / \mathrm{cm}$. Ammonia concentrations were elevated on the first two sampling dates, with an average of $4.7 \mathrm{mg} / \mathrm{L}$. At the same time, nitrite concentrations were elevated, ranging from 0.16 to $0.39 \mathrm{mg} / \mathrm{L}$. During the February 2023 sampling event, the surface water quality improved, but nitrate remained elevated, and chloride increased, reaching a maximum concentration of $29.5 \mathrm{mg} / \mathrm{L}$. Over the past five years, statistically significant increasing trends in chloride, sulphate, and nitrate concentrations have been detected at Sw-N-05.

Similar to last year, surface water at Sw-N-05 displayed water quality impacts related directly to aggregate production and stockpiling north of the landfill. In general, solute concentrations tend to increase in the wet season, likely due to flushing of dust and ANFO residue from roads and the aggregate stockpiles. Surface water quality at this location should continue to be monitored closely to assess the severity and evolution of these impacts.

### 7.6.3 Monitoring Sites Sw-N-14

This sampling location is off site along Heal Creek north of Willis Point Road, within the Heals Rifle Range. Sw-N-14 is located on Heal Creek, upstream of the confluence with Durrance Creek. In 2022/23, surface water samples were collected at Sw-N14 on four sampling dates.

Surface water quality at Site 14 met the BCWQG-STA, but nitrate exceeded the BCWQG-LTC on all four sampling dates. Additionally, dissolved copper and TSS exceeded the BCWQG-LTC during one sampling event. Nitrate concentrations ranged from 5.35 to $7.46 \mathrm{mg} / \mathrm{L}$, decreasing throughout the monitoring year. Similarly, sulphate concentrations were elevated at the beginning of the monitoring year ( $88 \mathrm{mg} / \mathrm{L}$ ), decreasing to $63 \mathrm{mg} / \mathrm{L}$ by the end of the monitoring year. These trends reflect
impacts from aggregate production and stockpiling. Surface water quality at this location should continue to be closely monitored to assess the severity and evolution of these impacts.

### 7.6.4 Monitoring Site Sw-N-19

Sw-N-19 is located within the landfill property below the northeast freshwater retention pond, approximately 80 m east of the lower leachate lagoon. It is not a compliance monitoring station. It receives runoff primarily from the bedrock/refuse interface adjacent to Phase 1 and discharges to the wetland upstream of Sw-N-16. Surface water samples were collected from Sw-N-19 in May and November of 2022, and in May 2023.

In 2022/23, all parameters were below BCWQG-STA guidelines on all sampling dates, but nitrite, nitrate and dissolved copper concentrations exceeded the BCWQG-LTC on at least one sampling date. Nitrate concentrations ranged from 3.05 to 11.7 $\mathrm{mg} / \mathrm{L}$, exceeding the BCWQG-LTC on all sampling dates. Copper concentrations ranged from 7.24 to $15.6 \mu \mathrm{~g} / \mathrm{L}$. Sulphate concentrations reflected historical values, ranging from 40 to $120 \mathrm{mg} / \mathrm{L}$, and ammonia and chloride concentrations were low. Over the past five years, a statistically significant decreasing trend in conductivity has been observed at Sw-N-19. Ultimately, the elevated nitrate and sulphate concentrations suggest aggregate stockpiling at Hartland impacted surface water quality at this location.

### 7.6.5 Monitoring Sites Sw-N-45 and Sw-N-17

Sw-N-45 is located north of Phase 2, and outside the landfill property boundary on the north side of Willis Point Road. In 2022/23, water quality at Sw-N-45 met BCWQG-STA, except for nitrate ( $44.2 \mathrm{mg} / \mathrm{L}$ ) in November 2022. Nitrate concentrations exceeded the BCWQG-LTC on all sampling dates, and dissolved copper concentration exceeded the BCWQG-LTC on three of four sampling dates. Copper concentrations only ranged from 1.16 to $1.78 \mu \mathrm{~g} / \mathrm{L}$. Sulphate concentrations were elevated and decreased throughout the monitoring year, ranging from 70 to $160 \mathrm{mg} / \mathrm{L}$. Ammonia concentration was below the detection limit on all sampling dates. Chloride concentrations were generally low and increased to $54 \mathrm{mg} / \mathrm{L}$ in March 2023. Over the past five years, no statistical trends in solute concentrations have been identified at Sw-N-45.

Sw-N-17 is located north of Willis Point Road and is downstream of Sw-N-45 and Sw-N-05. Water quality at Sw-N-17 met BCWQG-STA guidelines on all sampling dates, but concentrations exceeded the BCWQG-LTC for nitrate on all sampling dates, and for dissolved copper in March 2023. Sulphate concentrations were elevated and decreased throughout the monitoring year, ranging from 86.0 to $130 \mathrm{mg} / \mathrm{L}$. Chloride concentrations were also elevated slightly and decreased throughout the monitoring year from 13.0 to $39.0 \mathrm{mg} / \mathrm{L}$. Ammonia concentrations were below the detection limit on all sampling dates. Conductivity was relatively high at the beginning of the monitoring year ( $690 \mu \mathrm{~S} / \mathrm{cm}$ ) and later decreased to values ranging from 331 to $351 \mu \mathrm{~S} / \mathrm{cm}$. Over the past five years, a statistically significant increasing trend in chloride concentrations has been observed at $\mathrm{Sw}-\mathrm{N}-17$, which may reflect road salting.

In 2022/23, surface water quality at monitoring locations 17 and 45 was impacted by aggregate production and stockpiling, as indicated by elevated conductivity, nitrate and sulphate concentrations. Surface water quality at these locations should continue to be monitored closely to assess the evolution of these water quality impacts downstream of the landfill.




Figure 7-2. Surface Water Quality North of Phase 1

### 7.7 Surface Water Quality North of Phase 2

### 7.7.1 Surface Water Quality in the Hartland North Pad Area

The Hartland North Pad is located northwest of the landfill as shown on Figure 7-1. Recent geologic mapping of the ridge to the south of the North Pad (AECOM 2018) revealed an undulating bedrock surface with extensive deformation, fracturing, and mineralization. This area continued to undergo significant changes with the construction of the Residuals Treatment Facility (RTF). Development in the vicinity of the RTF began in 2017, and the construction was completed in late 2020. The treatment process at the RTF anaerobically digests residual solids from the McLoughlin Point Wastewater Treatment Plant into Class A biosolids.

The west side of the Hartland North Pad drains northward through an ephemeral channel that originates at the northwest corner of the Hartland North Pad. The water is carried through a culvert under Willis Point Road and into a drainage channel that eventually discharges into Durrance Lake approximately 450 m to the north. Flow through the channel only occurs during wet weather periods. During dry periods, several wetlands persist along the drainage course, but they are not connected by surface flows. In the downstream portions of the creek, flows increase due to groundwater discharge to the stream. Further downstream, a second creek of similar size joins the original creek. The "combined" creek has a well-defined channel in the area where it discharges to Durrance Lake.

### 7.7.1.1 Monitoring Sites Sw-N-41, Sw-N-42, and Sw-N-CSs2

Figure 7-5 presents conductivity, ammonia, chloride, and sulphate concentrations at sampling locations along the ephemeral channel at the northwest corner of the Hartland North Pad and the drainage channel discharging into Durrance Lake at stations Sw-N-41s1, Sw-N-41s3, Sw-N-41s4, Sw-N-42s1, and Sw-N-CSs2, which are located off-site. The sampling location Sw-N-CSs2 was dry in 2022/23, and surface water samples were not collected at that location.

At surface water station Sw-N-41s1, all parameters met the BCWQG-STA, but TSS exceeded the BCWQG-LTC on two sampling dates. Chloride and ammonia concentrations were low, and nitrate and sulphate concentrations were moderate, ranging from 0.17 to $0.736 \mathrm{mg} / \mathrm{L}$ and 48 to $68.5 \mathrm{mg} / \mathrm{L}$, respectively. Conductivity was elevated, ranging from 229 to $360 \mathrm{mg} / \mathrm{L}$. Over the past five years, statistically significant decreasing trends in ammonia, sulphate, and nitrate concentrations have been observed at this location. The slightly elevated sulphate, nitrate and conductivity exhibited continued minor impacts from aggregate stockpiling and placement on or near the Hartland North Pad.

At surface water station Sw-N-41s3, all parameters met applicable BCWQGs on all sampling dates. Conductivity was low and ranged from 122 to $320 \mathrm{mg} / \mathrm{L}$. Chloride ( $<8.8 \mathrm{mg} / \mathrm{L}$ ), ammonia ( BDL ), and sulphate ( $<25 \mathrm{mg} / \mathrm{L}$ ) concentrations were low, and nitrate concentrations were low until February 2023, when they increased to $1.56 \mathrm{mg} / \mathrm{L}$. On that sampling date, conductivity $(147 \mu \mathrm{~S} / \mathrm{cm})$ and sulphate ( $17.0 \mathrm{mg} / \mathrm{L}$ ) concentrations were low. The slightly elevated nitrate and sulphate concentrations may reflect dilute aggregate runoff.

At surface water station Sw-N-42s1, all parameters met applicable BCWQGs, except for TSS on the first two sampling dates. Conductivity was elevated, ranging from 276 to $371.5 \mu \mathrm{~S} / \mathrm{cm}$. Ammonia and chloride concentrations were low, and nitrate concentrations were moderate, reaching up to $0.39 \mathrm{mg} / \mathrm{L}$ in January 2023. Sulphate concentrations were elevated and ranged from 67.5 to $120 \mathrm{mg} / \mathrm{L}$. Water quality at this location was impacted by aggregate stockpiling.

Sw-N-41s4 is located further downstream of Sw-N-41s3. In 2022/23, all parameters met applicable BCWQGs on all sampling dates. Conductivity was low, as well as sulphate, nitrate, ammonia, and chloride concentrations. No water quality impacts were observed at this location.

Overall, surface water quality at the Hartland North Pad and in the downstream receiving environment was generally consistent with historical results and exhibited continued minor impacts from aggregate stockpiling and placement. Further downstream of $\mathrm{Sw}-\mathrm{N}-41 \mathrm{~s} 3$, surface water is not impacted by aggregate stockpiling or any other landfill activities.

### 7.7.2 Surface Water Quality in the Phase 2 Area

### 7.7.2.1 Monitoring Sites Sw-N-51 and Sw-N-50

Stations Sw-N-50 and Sw-N-51 are located near the northern boundary of the newly constructed Phase 2 Cell 3 and near the northwest diversion ditch near the Toutle Valley rock quarry and aggregate processing area. They are not Boundary Compliance stations. In 2022/23, stations Sw-N-50 and Sw-N-51 were destroyed due the construction of Cells 4,5 and 6 and could not be sampled.

### 7.7.2.2 Monitoring Sites Sw-N-54 and Sw-N-18

Sw-N-18 and Sw-N-54 (not Boundary Compliance stations) are located adjacent to the Upper Leachate Lagoon. Station Sw-N18 is located near the headwaters of Heal Creek and monitors the combined discharge from the Sw-N-50 and Sw-N-51 catchment areas. Sw-N-54 monitors discharge from the northwest freshwater retention pond to Phase 2 Cell 1 and the NWSP.

Surface water at Sw-N-18 exhibited highly elevated conductivity, ammonia, nitrate, nitrite, aluminum, and metal concentrations. TSS, total and dissolved iron, ammonia, nitrite, and nitrate concentrations exceeded the BCWQG-STA, and TSS, dissolved copper, dissolved zinc, dissolved cadmium, total cobalt, and total aluminum concentrations exceeded the BCWQG-LTC. Conductivity values ranged from 440 to $1,240 \mu \mathrm{~S} / \mathrm{cm}$. Ammonia concentrations ranged from 6.3 to $82 \mathrm{mg} / \mathrm{L}$, fluctuating throughout the monitoring year. Similarly, nitrate and nitrite concentrations fluctuated throughout the monitoring period, ranging from 2.02 to $58.4 \mathrm{mg} / \mathrm{L}$ and 0.306 to $0.681 \mathrm{mg} / \mathrm{L}$, respectively. Chloride concentrations ranged from 3.6 to 84.0 $\mathrm{mg} / \mathrm{L}$, and sulphate concentrations ranged from 81.0 to $180 \mathrm{mg} / \mathrm{L}$. Overall, surface water quality at location Sw- $\mathrm{N}-18$ was impacted by dilute landfill leachate and aggregate runoff.

In November 2022, the plug and diversion measures at surface water station SW-N-18 were removed, directing the discharge to the Upper Leachate Lagoon, in response to water quality exceedances. Recognizing the need to handle potentially contaminated water more effectively, work is currently underway to install a diversion pipe at station SW-N-18. This new infrastructure aims to channel the contaminated water directly into the leachate system.

Similar to Sw-N-18, surface water at location Sw-N-54 exhibited elevated conductivity, ammonia, nitrate, iron, and cobalt concentrations. TSS, dissolved iron, and total iron and manganese concentrations exceeded the BCWQG-STA on at least one sampling date. Additionally, TSS, total aluminum, dissolved copper, and cobalt concentrations exceeded the BCWQG-LTC on at least one sampling date. Conductivity ranged from 331.0 to $570 \mathrm{mg} / \mathrm{L}$, ammonia concentrations ranged from 0.016 to 1.2 $\mathrm{mg} / \mathrm{L}$, and nitrate concentrations ranged from 0.69 to $2.88 \mathrm{mg} / \mathrm{L}$. Chloride concentrations ranged from 2.2 to $19.0 \mathrm{mg} / \mathrm{L}$, and sulphate concentrations ranged from 36 to $140 \mathrm{mg} / \mathrm{L}$. Overall, surface water continued to be impacted by aggregate stockpiling. Given the concurrent increase in ammonia, chloride and conductivity concentrations in January and February 2023, water quality is interpreted to have been impacted by dilute leachate over that period of time.

Due to the dilute leachate impacts and ongoing quarrying activities, CRD has implemented mitigation measures to manage water quality in the Toutle Valley and the NWSP. Temporary tarps were installed on a large portion of the Phase 2 landfill, and aggregate storage in the Toutle Valley was carefully managed. In 2021, aggregate was removed from Toutle Valley and placed in the cleared area south of the Hartland North Pad. A blasting program was initiated for Cell 4 preparation in October 2021. Given the sensitivity of surface water quality to blasting, quarrying and runoff from aggregate stockpiles at these locations, water quality at these locations should continue to be monitored closely. Careful surface water management planning is required for this area as the landfill develops to minimize impacts on groundwater and surface water that is not captured by the leachate collection system. Additional sediment control measures and efforts to reduce the quantity of blasting products may help reduce impairment to water quality as quarry development becomes increasingly close to the northern boundary of the landfill and these water quality monitoring stations.

### 7.7.2.3 Monitoring Site Sw-N-53

Sw-N-53 discharges into the NWSP and conveys surface water from a shotcrete-lined ditch south of the Hartland North Pad. Historically, few surface water samples have been collected from Sw-N-53 due to dry conditions. In 2022/23, only one sample was collected under very turbid conditions (February 2023). On that date, TSS, manganese, cobalt, total and dissolved iron, and zinc concentrations exceeded the BCWQG-STA, and TSS, dissolved copper, total zinc, chloride, aluminum and manganese concentrations exceeded the BCWQG-LTC. The chloride concentration was $170 \mathrm{mg} / \mathrm{L}$, but ammonia was only $0.19 \mathrm{mg} / \mathrm{L}$. Sulphate and nitrate concentrations were low.

At the time of sampling, turbidity had notably impacted the surface water quality at $\mathrm{Sw}-\mathrm{N}-53$, evident from the significantly elevated concentrations of iron and aluminum. While historically, water quality at Sw-N-53 was impacted by runoff from aggregate stockpiles, the one 2022/23 sample which was collected during the wet season, did not clearly indicate impacts by runoff from nearby aggregate stockpiles.

### 7.8 Surface Water Quality Further North of the Landfill

Surface water was collected at three locations (SW-N-63, SW-N-64, and SW-N-65) approximately 900 m northeast of Hartland, at confluence of Durrance Creek and Tod Creek. These stations are not Boundary Compliance stations. Water quality at these locations met all applicable BCWQG guidelines, except for low pH readings at Sw-N-63 and Sw-N-64, which were below the BCWQG-STA. Dissolved copper concentrations exceeded the BCWQG-LTC at all three locations on at least
one sample date, likely due to lower pH and hardness values. Overall, surface water quality at these locations was good, with low sulphate, ammonia, and chloride concentrations, and near background conductivity ( $<210 \mu \mathrm{~S} / \mathrm{cm}$ ). At station Sw-N-63, nitrate and sulphate concentrations were slightly elevated, ranging from 0.958 to $1.28 \mathrm{mg} / \mathrm{L}$, and 16 to $20 \mathrm{mg} / \mathrm{L}$, respectively. Water quality at $\mathrm{Sw}-\mathrm{N}-63$ may be marginally affected by aggregate runoff or the application of fertilizers to neighboring agricultural areas. Monitoring at SW-N-63, SW-N-64, and SW-N-65 should continue to be monitored to assess the impact of aggregate runoff and to characterize the background signatures of nitrate/sulphate downstream of the landfill.

### 7.9 Surface Water Quality South of the Landfill

An ephemeral stream drains the area to the south of the landfill and flows southward towards Killarney Lake, which subsequently drains to Prospect Lake. Surface water flow south of the landfill occurs mainly during periods of wet weather, and groundwater seepage has been observed in the Killarney Creek channel during dry periods. Clean surface water runoff from the south slope of Phase 1 runs westward in a ditch to a culvert that discharges into a small wetland at Sw-S-03 and then into the ephemeral stream that flows south to Killarney Lake.

There are several surface water sampling stations located south of the landfill, listed from upstream to downstream, as follows:

- Sw-S-52 diversion ditch rerouted from north of the landfill, upstream of wheel wash facility.
- Sw-S-20 flow monitoring weir along diversion ditch at south end of Phase 1.
- Sw-S-12 flow monitoring weir upstream of the public weigh scale.
- Sw-S-03 culvert emerging from southeast corner of recycling area immediately upstream of a small natural wetland.
- Sw-S-27 Killarney Creek, north tributary.
- Sw-S-24 Killarney Creek, downstream of confluence of north and west tributaries.
- Sw-S-21 drainage ditch along road south of diversion ditch at south end of Phase 1.
- Sw-S-04 Killarney Creek, on property boundary, 270 m south of the landfill.


### 7.9.1 Upgradient Surface Water Quality

### 7.9.1.1 Monitoring Site Sw-S-52

Sw-S-52 is a background monitoring station. In 2022/23, surface water samples were collected at Sw-S-52 on all sampling dates. Concentrations of all parameters were below the BCWQG-STA. Like previous years, sulphate concentrations at Sw-S-52 were relatively stable and remained below $10 \mathrm{mg} / \mathrm{L}$ throughout the monitoring year. Nitrate concentrations were below $0.039 \mathrm{mg} / \mathrm{L}$. No statistically significant trends have been observed at this sampling location over the past five years.

### 7.9.1.2 Monitoring Site Sw-S-20

Surface water samples were collected from Sw-S-20 on three sampling dates. Concentrations of all parameters were below the BCWQG-STA, and dissolved copper concentrations exceeded the BCWQG-LTC. Slightly elevated metal concentrations during the wet season may be partially due to elevated TSS. Ammonia concentrations were elevated ( $1.0 \mathrm{mg} / \mathrm{L}$ ) in February 2023, but the chloride concentration ( $7.7 \mathrm{mg} / \mathrm{L}$ ) was low. No statistically significant trends have been observed at this sampling location over the past five years.

### 7.9.2 Surface Water Quality Near and South of the Recycling Area

Figure 7-6 presents water quality data for surface water sampling stations located south of the landfill, including Sw-S-03, Sw-S-04, and Sw-S-12. The distance from each station to the landfill boundary is shown in brackets in the legend. CRD's recycling area went into operation in January 2001 and is located near Sw-S-12.

### 7.9.2.1 Monitoring Site Sw-S-12

In 2022/23, surface water samples were collected at station Sw-S-12 on all four sampling events. BCWQG-STA exceedances were reported for TSS (two sampling events), dissolved iron (three sampling events), and total iron (three sampling events). All three parameters exceeded BCWQG-STA on February 2023, which may reflect turbid flow conditions and high TSS. In November 2022, ammonia concentrations were relatively low at $0.017 \mathrm{mg} / \mathrm{L}$. However, for all subsequent sampling dates, concentrations increased to more than $2.6 \mathrm{mg} / \mathrm{L}$. In comparison, other parameters such as chloride, conductivity, nitrate, and sulphate exhibited a distinct pattern. Their highest concentrations were recorded in November 2022, which then decreased over the next two sampling events. The concurrent increase in conductivity, chloride, nitrate, and sulphate levels during the

November 2022 sampling suggests possible impacts from aggregate runoff or runoff from the paved areas around the bin facility. The elevated ammonia levels seen during the wet season could have resulted from denitrification, which may be due to decomposing organic matter. Chloride concentrations peaked at $33 \mathrm{mg} / \mathrm{L}$ in November 2022 but remained below $20 \mathrm{mg} / \mathrm{L}$ on other sampling dates. This occasional elevated chloride could be attributed to road salt application and unlikely to be influenced by leachate.

### 7.9.2.2 Monitoring Site Sw-S-03

Sw-S-03 is located on the landfill property in the main channel of Killarney Creek where the culvert discharges into a small wetland area. $\mathrm{Sw}-\mathrm{S}-03$ is not a compliance location. Historically, water quality at $\mathrm{Sw}-\mathrm{S}-03$ was affected by contaminated runoff from the south face of the landfill and the former truck wash area, until the truck wash facility was relocated in the fall of 1997. Due to the proximity to the public drop-off and storage area, water quality in Sw-S-03 may be affected by runoff from the bin facility, heavy traffic, and industrial activities.

In 2022/23, surface water quality at Sw-S-03 was poor. Surface water samples were collected on all four sampling events, and BCWQG-STA exceedances were reported for TSS (three sampling events), chloride (one sampling event), total iron (two sampling events), dissolved iron (one sampling event). Furthermore, exceedances of BCWQG-LTC criteria were reported for total aluminum (three sampling events), total cobalt (two sampling events), total lead (one sampling event), nitrate (one sampling event), and dissolved copper concentrations (three sampling events). Except for November 2022, ammonia concentrations were elevated, and ranged from 1.3 to $1.5 \mathrm{mg} / \mathrm{L}$. In contrast, nitrate, conductivity, chloride, and sulphate peaked in November 2022 but reverted to historical values during subsequent sampling events. The notable conductivity spike to $2,370 \mu \mathrm{~S} / \mathrm{cm}$ can be attributed to an exceptionally high chloride level of $950 \mathrm{mg} / \mathrm{L}$. Leachate is not likely the source of the chloride and conductivity impacts, given the low ammonia reading of $0.045 \mathrm{mg} / \mathrm{L}$ in November and modest chloride concentrations ( 39 to $220 \mathrm{mg} / \mathrm{L}$ ) in the south purge wells.

Overall, surface water quality at Sw-S-03 was temporarily impacted by road salting, aggregate runoff from the stockpiles on Phase 1 and/or from the paved area at the landfill entrance. Although surface water quality deteriorated in November 2022, it showed improvement during subsequent 2023 sampling events. Water quality should continue to be closely monitored for aggregate stockpiling and leachate impacts in the area downgradient of the bin facility and the south purge wells.

### 7.9.2.3 Monitoring Site Sw-S-04

Sw-S-04 is the southernmost Boundary Compliance Station at Hartland. In 2022/23, surface water quality at station Sw-S-04 met all BCWQG-STA criteria, except for one total iron concentration observed in May 2022. Furthermore, TSS, total aluminum, zinc, and dissolved copper concentrations exceeded the BCWQG-LTC on one sampling date. Nitrate concentrations were relatively high, ranging from 1.02 to $2.33 \mathrm{mg} / \mathrm{L}$, and sulphate concentrations were moderate, ranging from 16 to $60 \mathrm{mg} / \mathrm{L}$. Ammonia concentrations were generally at or below the detection limit, and conductivity was moderate to elevated, ranging from 157 to $253 \mu \mathrm{~S} / \mathrm{cm}$. Overall, the surface water quality at Sw-S-04 showed evidence of dilute aggregate runoff impacts, but low ammonia and chloride concentrations indicate that water quality was not impacted by leachate.

### 7.9.2.4 Monitoring Sites Sw-S-24 and Sw-S-27

Stations Sw-S-24 and Sw-S-27 (not Boundary Compliance stations) are located on the landfill property downgradient of the Phase 1 landfill, the landfill administration area, and mountain bike trails. In 2022/23, surface water samples were collected from Sw-S-27 on one date. The surface water quality met all BCWQG-STA criteria, except for TSS on one date, and all leachate and aggregate runoff indicator parameters were low. Surface water quality at Sw-S-24 met all BCWQG-STA criteria, except for TSS and total iron concentration in May 2022. Additionally, dissolved copper, and total aluminum, iron, and zinc concentrations exceeded the BCWQG-LTC during one sampling event. Nitrate concentrations were elevated throughout the monitoring period, ranging from 1.13 to $2.86 \mathrm{mg} / \mathrm{L}$, and sulphate concentrations ranged from 15 to $72 \mathrm{mg} / \mathrm{L}$. Similar to previous years, conductivity was elevated, with an average of $252 \mu \mathrm{~S} / \mathrm{cm}$. Overall, surface water quality at $\mathrm{Sw}-\mathrm{S}-27$ showed evidence of aggregate runoff impacts, as indicated by elevated conductivity, sulphate, and nitrate concentrations. Low chloride concentrations suggest water quality was not impacted by leachate.






Figure 7-3. Surface Water Quality North of Phase 1 - Aggregate Impacts






$$
\begin{array}{ccccccc}
\text { Jan-2018 Jan-2019 Jan-2021 Jan-2022 } & \text { Jan-2020 } & \text { Jan-2023 } & \text { Jan-2024 }
\end{array}
$$

Figure 7-4. Surface Water Quality North of Phase 2






Figure 7-5. Surface Water Quality Downstream of the Hartland North Pad





Figure 7-6. Surface Water Quality South of Landfill

### 7.10 Summary

The surface water quality observed at and around the landfill indicates that nearby surface water bodies, Tod Creek, Durrance Lake, Durrance Creek, and Killarney Lake were not impacted by landfill leachate in 2022/23. However, surface water quality monitoring stations at the landfill continued to show signs of water quality degradation, especially in the area northwest of Phase 2. Surface water quality in the Phase 2 area exhibited impacts related to runoff from the aggregate stockpiles located in the Phase 2 quarry, on the bedrock ridge north of the Phase 2 landfill, and other sources of ammonia during wet weather. Based on historic data and the 2022/23 surface water quality data, AECOM made the following interpretations:

- Surface water at Boundary Compliance location SW-N-05 continued to exhibit elevated nutrient concentrations in 2022/23, resulting in non-compliant water quality at this location. Nitrate concentrations at Sw-N-05 exceeded BCWQG-STA during the May 2022 and November 2022 sampling events. The elevated nitrate and sulphate concentrations suggest an impact on surface water from aggregate production and stockpiling. However, the absence of paired ammonia and chloride concentrations suggest water quality was not affected by leachate. The occasionally elevated ammonia concentrations may be associated with nitrate reduction via denitrification under reducing conditions.
- Surface water quality at Boundary Compliance station Sw-N-16 met BCWQG-STA, except for total iron during the February 2023 sampling event. The iron exceedance observed at $\mathrm{Sw}-\mathrm{N}-16$ was likely due to the disturbance of sediment during sampling, as indicated by the reported high TSS concentrations. Surface water quality at Sw-N-16 was not impacted by leachate, but continued to exhibit minor influence from nearby construction activities involving blasting, aggregate placement, aggregate hauling and excavation of organic soils. Similar aggregate impacts were observed at downstream at stations Sw-N-17 and Sw-N-45.
- Surface water at Sw-N-18 reflected dilute landfill leachate and may also have been impacted by aggregate runoff. In November 2022, the plug and diversion measures at surface water station SW-N-18 were removed, directing the discharge to the NWSP. Work is currently underway to install a diversion pipe at station SW-N-18 to direct the contaminated water directly into the leachate collection system.
- In the Hartland North Pad area, surface water quality at Boundary Compliance stations (Sw-N-41s1 and Sw-N-42s1) met BCWQG-STA in 2022/232, except for TSS. Leachate indicator parameters remained low in 2022/23, indicating surface water was not impacted by landfill leachate or construction activities. The slightly elevated sulphate, nitrate and conductivity exhibited continued minor impacts from aggregate stockpiling and placement.
- Historically, surface water stations Sw-N-14 and Sw-N-CS2 were used to monitor background conditions north of the landfill, but elevated conductivity, nitrate, and sulphate concentrations observed in 2022/23 indicate that surface water quality at these stations has been impacted by aggregate runoff. Sw-S-52 consistently showed no signs of impacts related to the landfill, confirming that the water quality remains representative of the background conditions south of the landfill.
- Surface water quality downgradient of the area north of the landfill (Sw-N-41s3) exhibited slightly elevated nitrate concentrations ( 1.28 to $1.57 \mathrm{mg} / \mathrm{L}$ ) corresponding to low to moderate sulphate concentrations ( 17.0 to $20.0 \mathrm{mg} / \mathrm{L}$ ). In the absence of elevated sulphate concentrations, it is difficult to interpret whether the elevated nitrate concentrations reflect a background process, agricultural impacts, or dilute aggregate runoff originating from Hartland landfill. Further downstream of $\mathrm{Sw}-\mathrm{N}-41 \mathrm{~s} 3$, water quality at $\mathrm{Sw}-\mathrm{N}-41 \mathrm{~s} 4$ was consistent with background conditions and showed no signs of aggregate or leachate impacts.
- Further downstream to the north of the landfill, at the confluence of Durrance Creek and Tod Creek (Sw-N-64 and Sw-$\mathrm{N}-65$ ), surface water quality showed no impacts from landfill leachate or aggregate runoff. The slightly elevated nitrate ( $<1.5 \mathrm{mg} / \mathrm{L}$ ) and sulphate ( $<20 \mathrm{mg} / \mathrm{L}$ ) concentrations at $\mathrm{Sw}-\mathrm{N}-63$ may have originated from the on-site aggregate runoff or were associated with the application of fertilizers to the surrounding agricultural lands.
- Water quality at the Boundary Compliance location (Sw-S-04) met the BCWQG-STA values for all analytes in all samples collected during 2022/23, except for one total iron concentration observed in May 2022. Surface water quality along the south boundary was not impacted by leachate but exhibited impacts from dilute aggregate runoff.
- Water quality at station Sw-S-52 (not a Boundary Compliance location) is representative of background water quality. In 2021/22, concentrations of all parameters were below the BCWQG-STA. Concentrations of leachate indicator parameters were consistent with previously reported values.
- Surface water quality south of the recycling area (Sw-S-03, Sw-S-12) exhibited several BCWQG-STA exceedances, including TSS, dissolved and total iron, and chloride during one or more sampling date. Elevated ammonia, nitrate, conductivity and sulphate concentrations at these stations may be related to aggregate dust from the south face of Phase 1 and runoff from paved areas surrounding the bin facility, heavy traffic, and industrial activities.


## 8. Leachate

### 8.1 Compliance Monitoring Locations

Discharge from the leachate pipeline is subject to the CRD Regional Source Control Program (RSCP) Waste Discharge Authorization (Waste Discharge Authorization Number SC97.001). The compliance monitoring location for leachate at Hartland Landfill is the Hartland Valve Chamber (flow detection chamber) at the start of the leachate pipeline. Leachate compliance data is reported to the CRD RSCP on a quarterly basis.

### 8.2 Data

Our interpretation of the leachate chemistry data was based on samples collected at the following locations by CRD staff:

- Hartland Valve Chamber (leachate pipeline flow detection chamber and compliance point)
- Phase 1 North Purge Well System (combined discharge from 52-4-0-P7, 80-1-0-P8 and 81-1-0-P9)
- Phase 1 South Purge Well System (combined discharge from P1, P2, P3, P4 and P10)
- Controlled Waste Drainage
- West Face Drainage
- Cell 3 Pipe Outlet Drainage

These locations were sampled and analyzed for conventional parameters, organic compounds, and metals on a monthly basis in 2022/23. Additionally, Hartland Valve Chamber samples were analyzed quarterly for trace organic compounds including polycyclic aromatic hydrocarbons (PAHs), phthalate esters, ketones, aromatics, phenols, ethers, nitrosamines, alkanes, alkenes, and other select organic compounds.

In 2022/23, eight (8) leachate samples were collected at Cell 3 Pipe Outlet, and sixteen (16) leachate samples were collected from the Hartland Valve Chamber. Ten (10) leachate samples were collected from the Controlled Waste Drainage, and ten (10) leachate samples were collected from the South Purge Wells. Nine (9) leachate samples were collected from the West Face Drainage, where intermittent drainage patterns are attributed to precipitation, refuse settlement, and the increasing depth of waste cover over the toe drain. In 2022/23, no leachate samples were collected from the Phase 2 Cleanout because of ongoing site maintenance.

### 8.3 Leachate Generation and Discharge

Leachate collected from the Phase 1 and Phase 2 landfill is discharged to the lower leachate lagoon. During wet winter months, leachate is pumped into the lined upper leachate lagoon to minimize head build-up in the unlined lower lagoon. Leachate from the lagoons is discharged from the site through an 8.6 km long pipeline that discharges to the Saanich sanitary sewer and ultimately to the new McLoughlin Point Wastewater Treatment Plant. The CRL was activated on December 30, 2020, and the formal switch from the leachate pipeline to the CRL occurred in March 2021, with flow rates ranging between 60 and $80 \mathrm{~L} / \mathrm{s}$.

Total monthly leachate flows discharged to sewer are provided in Appendix D. Average monthly leachate flow in 2022/23 was 16.5 L/s and slightly lower than in 2021/22 (17.9 L/s). The highest monthly leachate flow was observed in January ( $90,297 \mathrm{~m}^{3}$ ) in response to intense winter precipitation events. Changes in leachate discharge rates throughout the year may be related to the increased capacity of the Centrate Return Line (CRL) from the RTF, variability in precipitation, biofouling within the line, or operational improvements aimed at minimizing leachate generation.

### 8.4 Leachate Quality

Sampling and testing of leachate quality have been carried out since the early 1970s. Since 2000, leachate samples have been collected primarily from the Hartland Valve Chamber, which represents the point of discharge for compliance with the RSCP Waste Discharge Authorization.

The analytical results of the routine monthly leachate discharge samples are provided in Appendix B.4. Analysis of trace organics in the leachate discharge was conducted quarterly and is provided in Appendix B.5. The analytical results for samples collected from the leachate collection and conveyance network are presented in Appendices B.6., B.7., B.8., B.9., B.10, and B.11. Analysis of emerging contaminants at the Hartland Valve Chamber was momentarily paused while AECOM is reviewing data from 2018 to 2022. An updated list of emerging contaminants will be integrated into the 23/24 monitoring program. The results of a single sample taken in May 2022 for analysis of emerging contaminants is provided in Appendix B.12. In addition to the Sewer Use Bylaw Criteria, leachate quality results for trace organic compounds were screened against CSR standards for the protection of drinking water and aquatic life to support operational decisions regarding leachate containment and management.

### 8.4.1 Routine Monthly Leachate Analyses and Sewer Use Bylaw Comparison

The Hartland Valve Chamber is the compliance point for the Waste Discharge Authorization. In 2022/23, all leachate quality samples met RSCP Waste Discharge Authorization criteria, except for COD on multiple sampling dates. Since COD concentrations measured in 2022/23 are unreliable and were not representative of leachate quality, COD values are not discussed below.

Although the Waste Discharge Authorization criteria only apply to the combined discharge at the Hartland Valve Chamber, comparison of other leachate monitoring station results to these criteria allows for evaluation of individual leachate contributions to the combined leachate discharge. Concentrations above RSCP criteria at locations other than the Hartland Valve Chamber are not considered to be non-compliant, and the criteria are used for reference purposes only.

Figure 8-1 and Figure 8-2 present time series plots for selected parameters in leachate at the compliance point (Hartland Valve Chamber) and in Phases 1 and 2. Evaluation of this data allows for comparison of leachate from each landfill area. At all sampling locations, concentrations of inorganic parameters such as conductivity and chloride show a seasonal dilution effect whereby greater precipitation in the fall and winter months results in lower concentrations during the wet winter months. Higher concentrations occur during drier periods from May to October. Leachate discharge from the Phase 2 Basin is significantly more concentrated than leachate generated in Phase 1, which was closed in 1996. The mixing of leachate from Phases 1 and 2 in the Hartland Valve Chamber results in a leachate that exhibits chemistry which is intermediate between the two sources.

Similar to previous years, leachate concentrations at the Hartland Valve Chamber were well above background concentrations observed in surface water and groundwater, with conductivity concentrations ranging from $2,303 \mu \mathrm{~S} / \mathrm{cm}$ to $5,846 \mu \mathrm{~S} / \mathrm{cm}$, ammonia concentrations ranging from $160 \mathrm{mg} / \mathrm{L}$ to $400 \mathrm{mg} / \mathrm{L}$, and chloride concentrations ranging from $200 \mathrm{mg} / \mathrm{L}$ to $490 \mathrm{mg} / \mathrm{L}$. Overall, annual average leachate concentrations in 2022/23 were slightly higher than in 2021/22.

Figure 8-2 shows BOD concentrations at the Hartland Valve Chamber. BOD concentrations at Hartland remain relatively low for a large landfill and were typically below $50 \mathrm{mg} / \mathrm{L}$. Changes in BOD may be related to the elevated temperatures observed during summer months, or changes in leachate storage and management prior to sample collection.

Total sulphide, dissolved sulphide and PAH concentrations did not exceed the Waste Discharge Authorization criteria at Hartland Valve Chamber during any of the 2022/23 sampling events. Total and dissolved sulphide concentrations exceeded the Waste Discharge Authorization criteria in January 2019 but have since remained well below the standard of $1 \mathrm{mg} / \mathrm{L}$. Infiltration through aggregate placed on the Phase 2 landfill during the winter months is known to be an important source of sulphate to the Phase 2 Cleanout. Sulphate can be reduced to sulphide by bacteria under reducing conditions in the absence of other reducing agents such as oxygen and nitrate.

Overall, leachate quality at the Hartland Valve Chamber was consistent with previous years, with concentrations of many parameters several orders of magnitude below Waste Discharge Authorization criteria. Over the past five years, a statistically significant decreasing trend in ammonia concentrations and an increasing trend in nitrate concentrations have been observed at the Hartland Valve Chamber. The elevated nitrate concentrations likely reflect aggregate runoff collected by the Leachate Collection System. Average nitrate concentrations decreased to $11.7 \mathrm{mg} / \mathrm{L}$ from $53.1 \mathrm{mg} / \mathrm{L}$ in 2022, and sulphate concentrations decreased to $95.3 \mathrm{mg} / \mathrm{L}$ from $129.0 \mathrm{mg} / \mathrm{L}$ in 2022.

### 8.4.1.1 Phase 2 Cleanout

In 2022/23, leachate samples were not collected from the Phase 2 cleanout due to ongoing site maintenance activities.

### 8.4.1.2 North Purge Wells

In 2022/23, leachate samples collected from the North Purge Wells (52-4-0-P7, 80-1-0-P8 and 81-1-0-P9) met all Waste Discharge Authorization criteria, except for COD on multiple sampling dates. Total sulphide concentrations in the Phase 1 leachate were low ( $<0.19 \mathrm{mg} / \mathrm{L}$ ).

### 8.4.1.3 South Purge Wells

Leachate quality data from combined effluent samples collected from the South Purge Wells (P1, P2, P3, P4 and P10) is provided in Appendix B.9. Leachate sampling at the South Purge Wells began in November 2020 and is now part of the monthly leachate sampling program. During the 2022/23 monitoring year, leachate samples collected from the South Purge Wells met all Waste Discharge Authorization criteria. In 2022/23, total sulphide concentrations ranged from 0.0097 to $0.036 \mathrm{mg} / \mathrm{L}$ and was well below the Waste Discharge Authorization criteria. BOD ranged from 2.5 to $54 \mathrm{mg} / \mathrm{L}$, with a median of $5.85 \mathrm{mg} / \mathrm{L}$.

Leachate quality in the South Purge Wells is characterized by moderately elevated conductivity ( 1,059 to $1,944 \mu \mathrm{~S} / \mathrm{cm}$ ), ammonia ( 44 to $250 \mathrm{mg} / \mathrm{L}$ ) and chloride concentrations ( 78 to $490 \mathrm{mg} / \mathrm{L}$ ), and low concentrations of sulphate ( $<10 \mathrm{mg} / \mathrm{L}$ ). In 2022/23, all metal concentrations in the South Purge Wells met Waste Discharge Authorization criteria.

### 8.4.1.4 Controlled Waste Drainage

Leachate collected by the Controlled Waste Drainage is not as concentrated as leachate collected from Phase 2 Cleanout and the West Face Drainage. In 2022/23, all leachate parameters and metal concentrations in the Controlled Waste Drainage leachate met Waste Discharge Authorization criteria.

### 8.4.1.5 West Face Drainage

In 2022/23, leachate discharged from the West Face Drainage throughout the year and was sampled monthly. Like previous years, the West Face Closure Toe Drain generally had the most concentrated leachate, with the highest BOD, ammonia, conductivity, and chloride concentrations.

In 2022/23, the average chloride, sulphate, nitrate, and BOD concentrations were generally lower than those measured in $2021 / 22$, but the average conductivity and ammonia concentrations were slightly higher. Sulphide concentrations remained low and met Waste Discharge Authorization criteria on all sampling dates, and total phenols exceeded the Waste Discharge Authorization criteria on one sampling date (July 2022).

Leachate from the West Face Drainage is strongly reduced with abundant organic content. A weir and flow monitor have been installed at the West Face Drainage to evaluate seasonal trends in flows. Because leachate from the West Face Drainage is contributing a relatively small volume of leachate to the leachate collection system, it does not noticeably affect the quality of leachate at the Hartland Valve Chamber.

### 8.4.1.6 Cell 3 Pipe Outlet

Cell 3 includes new leachate containment and gravity flow conveyance infrastructure (i.e., the Toutle Drain), which discharges directly into the upper leachate lagoon. Starting in 2016, the Cell 3 Pipe Outlet began discharging leachate, and leachate from this location is considered representative of newly deposited refuse. In 2022/23, samples were collected on eight of 12 sampling dates. Sampling from this station is challenging due to the intermittent flows. In 2022/23, all parameters met Waste Discharge Authorization criteria.


Figure 8-1. Hartland Valve Chamber Leachate Chemistry (Conductivity, Ammonia and Chloride)




Figure 8-2. Hartland Valve Chamber Leachate Chemistry (Sulphide, BOD and COD)

### 8.4.2 Quarterly Trace Organic Analysis at Hartland Valve Chamber

Since 1998, trace volatile and semi-volatile organic analyses have been carried out quarterly on leachate samples collected from the Hartland Valve Chamber. Chlorinated phenol compound analytical results for combined leachate in 2022/23 are presented in Appendix B.5.

A total of four (4) volatile and semi-volatile organic compounds were reported at detectable concentrations across all sampling dates. The detected compounds were found at concentrations that are low compared to those commonly found in leachate at municipal solid waste landfills of similar size. None of the trace organics exceeded Waste Discharge Authorization criteria. Like previous years, concentrations of volatile organic compounds in Phase 2 leachate (i.e., the West Face Drainage and Phase 2 Cleanout) are exceedingly low and typically at concentrations on the order of $1 \%$ to $20 \%$ of Waste Discharge Authorization criteria. Regular sampling and analysis for VOC concentrations in leachate sources should continue, but it is not warranted in groundwater at compliance monitoring locations at this time.

In 2022/23, two phenolic compounds were above detection limits (2,4,6-trichlorophenol and 2,3,4,6-tetrachlorophenol). However, the phenol concentrations in leachate at Hartland are lower than or similar to those found at other large municipal solid waste landfills. Phenols are used in several manufacturing processes and occur naturally at low concentrations due to their presence in wood and other natural organic matter.

In 2022/23, all low-weight PAHs were detected in Hartland Landfill leachate. Total concentrations of low-weight PAHs ranged from 0.011 to $15 \mu \mathrm{~g} / \mathrm{L}$, which was consistent with historical concentrations. Acenaphthene exhibited a maximum concentration of $6.1 \mu \mathrm{~g} / \mathrm{L}$ in October 2022, slightly above the maximum value ( $5.2 \mu \mathrm{~g} / \mathrm{L}$ ) observed in $2021 / 22$. High-weight PAHs, including benzo(a)anthracene, fluoranthene, and pyrene continued to be detected with maximum concentrations of $0.12 \mu \mathrm{~g} / \mathrm{L}, 1.1 \mu \mathrm{~g} / \mathrm{L}$, and $0.79 \mu \mathrm{~g} / \mathrm{L}$, respectively. Chrysene and benzo(a)pyrene have been marginally above detection limits since 2016. Overall, high-weight PAH concentrations were considerably lower than in 2021/22, ranging from 0.005 to $1.1 \mu \mathrm{~g} / \mathrm{L}$.

The CRD Environmental Protection Division conducted high-resolution analyses of leachate quality between 2004 and 2007. The high-resolution analytes included polychlorinated biphenyls (PCBs), polybrominated diphenyl ethers (PBDEs), nonylphenols, and chlorobenzenes. Data collected revealed that nonylphenols and PBDEs are not present at levels of environmental concern, even in full strength leachate (Golder 2008). The aquatic risks of PCBs observed in leachate samples were also found to be negligible. Additionally, concentrations of nonylphenols, PCBs, DDT and select PDBEs varied in comparison to CRD wastewater. At the time of the study, Hartland leachate contributed only $2 \%$ of the total wastewater discharge from the Macaulay Point outfall. The evaluation concluded that despite differences in the chemical profile of leachate and Macaulay Point wastewater, leachate had no observable impact on the chemical profile at Macaulay Point. Based on these findings, the high-resolution analyses were discontinued. Co-ordination between the CRD's Marine Monitoring and Environmental programs is on-going.

### 8.5 Summary

The leachate flow and quality data collected in 2022/23 indicates the following:

- The average daily leachate flow rate was $16.5 \mathrm{~L} / \mathrm{s}$, with a maximum monthly discharge of $90,297 \mathrm{~m}^{3}$ in January 2023. The average flow was considerably higher than the long-term (1997 to 2022) average of approximately $12.2 \mathrm{~L} / \mathrm{s}$, which may be related to the expansion of the Phase 2, Cell 4/5/6 quarry.
- The exceedances of COD during multiple sampling events were determined to be due to compromised laboratorysupplied preservatives. Therefore, the analytical results are not considered representative of in-situ leachate.
- At the Hartland Valve Chamber, leachate quality was compliant with the Waste Discharge Authorization. Statistical trend analysis indicates that leachate quality has been relatively stable with minor improvements over the past five years.
- At the Hartland Valve Chamber, the highest concentrations of leachate parameters were observed in August and October 2022, when precipitation approached the annual minimum.
- Phase 1 leachate exhibited the lowest average values of BOD and highest sulphide concentrations compared to other leachate sampling locations.
- The West Face Closure Toe Drain exhibited the most concentrated leachate, with total phenols concentrations exceeding the Waste Discharge Authorization criteria on one sampling date.


## 9. Conclusions

Based on our review of historical data and interpretation of the groundwater, surface water, and leachate quality data collected between April 2022 and March 2023, the annual monitoring program allows for an effective assessment of landfill performance and compliance related to groundwater, surface water and leachate flow and quality. The following conclusions are drawn based on our interpretation of the 2022/23 data.

### 9.1 Leachate Flow

Based on review of historic data and leachate flow data collected in 2022/23, AECOM has drawn the following conclusions:

- Leachate elevation data collected in 2022/23 indicate that leachate mounding continued to persist in the Phase 1 landfill, as it has since it was closed in 1996. Leachate elevations in Phase 1 were generally stable and exhibited minor seasonal variations. The leachate mound in the upper portion of the refuse is interpreted as being 'perched' above the regional bedrock groundwater flow system, with relatively high water levels and strong downward hydraulic gradients.
- Like in 2021/22, leachate elevations in the Phase 2 Basin exceeded the elevation of the Lower Leachate Lagoon in November and remained above the elevation of the lagoon for the rest of the monitoring period. Historically, the leachate elevation in Phase 2 was approximately 1 to 2 m lower than the elevation of the Lower Leachate Lagoon. CRD confirmed that this trend was likely due to calibration/ instrument drift, and were not indicative of a change in operations. Nonetheless, the leachate elevation in Phase 2 was well-below the groundwater elevations observed at locations 36 and 37 , indicating that the hydraulic trap was preserved throughout the monitoring year.
- Leachate discharge rates in 2022/23 were lower than those observed in 2021/22. The total volume of leachate discharged in 2022/23 was $520,740 \mathrm{~m}^{3}$, approximately $7.5 \%$ lower than in 2021/22. The contrast in leachate discharge volumes likely reflects a lower volume of precipitation in 2022/23 compared to 2021/22. It is also possible the lower volume of leachate may be due to biofouling of the North Purge Wells and a resultant decrease in the volume of leachate extracted from these wells over the past year.
- In 2022/23. the highest leachate elevations ( 155 to 157 m asl) were typically observed in the east/southeast area of the Phase 1 (GW-46-2-1, VLGW-004D and VLGW-011S), an area with elevated topography and refuse heights. Leachate levels in deeper parts of the refuse respond to seasonal recharge events, indicating that the lower portions of the Phase 1 landfill are in hydraulic connection with the regional groundwater flow system in the bedrock.
- In 2022/23, a total of $30,580 \mathrm{~m}^{3}$ of leachate was collected from the South Purge Wells, approximately $13.3 \%$ less than in the previous monitoring year. Leachate discharge volumes and consistent groundwater levels observed in the South Purge Wells suggest the South Purge Well system functioned effectively in 2022/23.
- In 2022/23, a total of $14,277 \mathrm{~m}^{3}$ of leachate was collected from the North Purge Wells, approximately $21.3 \%$ less than in the previous monitoring year. Water levels in GW-40-1-1, GW-52-4-0-P7, GW-80-1-0-P8, and GW-81-1-0-P9 were generally consistent with historical ranges. Leachate discharge volumes and consistent groundwater levels observed in the North Purge Wells suggest the North Purge Well system functioned effectively in 2022/23, but there may be early signs of biofouling of the leachate collector wells. These wells have historically required rehabilitation every 5-10 years.


### 9.2 Groundwater Flow

In 2022/23, groundwater flow patterns observed at Hartland were consistent with historic interpretations, with some changes in the North Ridge area. Regional groundwater flows from Mount Work northeast to the north-south trending valley that underlies the northern portions of the Phase 1 and Phase 2 landfill footprint. Most of the northward groundwater flow in the bedrock below the landfill is captured by the Toutle Valley Underdrain, Phase 2 basin leachate collection system, springs discharging to the lower lagoon, and the north and south purge well systems. Groundwater monitors east of Phase 1 confirm flow from east to west toward the landfill, preventing off-site migration to the east.

Around the North Ridge and Hartland North Pad, located northwest of Phase 2, groundwater flows radially outward to the north, east and south from a topographic high situated north of Phase 2. Throughout 2022/23, continued blasting operations along the North Ridge resulted in reductions in both the topography and the groundwater potentiometric surface within the area contained by the Upper Level Road. Subsequently, this decline in groundwater surface led to diminished eastward
hydraulic gradients. It is suspected that the quarry cut through the Highland Fault, potentially facilitating the drainage of eastward moving groundwater that was previously impeded by the fault. This is consistent with the groundwater seepage observed at the base of quarry and generally lower groundwater elevations in the North Ridge area. Although groundwater elevations in the North Ridge area continued to exhibit seasonal fluctuations, the intensity of the fluctuations was less pronounced.

### 9.3 Groundwater Quality

The groundwater quality results from 2022/23 indicate that leachate-impacted groundwater was contained within the landfill property. At the north end of the landfill, leachate-affected groundwater extended just north of the unlined Lower Leachate Lagoon and through the middle of the lined Upper Leachate Lagoon but did not extend off-site. Leachate was identified in well GW-106-1-1, but impacts were limited to an area less than 20 meters northwest of the Phase 2 Basin. South of the landfill, leachate-affected groundwater did not extend off-site. Leachate related exceedances were confined to the landfill footprint on the east side of Phase 1 and are inferred to extend to the western extent of the waste footprint within the Phase 2 landfill. These results indicate that the leachate collection system continued to function as intended, minimizing surface water and groundwater quality impacts.

In 2022/23, Boundary Compliance wells and off-site monitoring wells met CSR AW and DW standards, except for an anomalous copper concentration exceedance at GW-21-1-1 in May 2022. However, this recorded exceedance may not accurately represent true conditions as indicated by the high RPD discrepancy between parent and duplicate samples. Dissolved copper in the parent sample was non-detected but highly elevated in the duplicate sample. Similar to previous years, most exceedances were present in groundwater wells near leachate purge wells and known leachate sources. However, nitrate concentrations in several groundwater wells located downgradient of aggregate stockpiles exceeded applicable CSR DW standards on one or more sampling event.

In 2022/23, groundwater in many areas of the landfill exhibited elevated conductivity, nitrate, and sulphate concentrations, reflecting the impacts of aggregate production, transport, stockpiling and use for construction at Hartland. Elevated concentrations of aggregate runoff parameters were observed around the Northwest Stockpile, North of Phases 1 and 2, around the Northeast Stockpile, south of Phase 1, and throughout the surface water system. AECOM is currently updating the Groundwater, Surface Water and Leachate Monitoring Plan to capture the cumulative impact of various activities, including aggregate stockpiling, placement, leachate discharge and on-going construction.

## North of the Landfill

- In 2022/23, annual average conductivity values in the North Purge Wells were generally higher than those in 2021/22, reflecting more concentrated leachate. Changes in leachate quality in the North Purge Wells may reflect more concentrated leachate due to lower precipitation in 2022/23 than 2021/22, or mixing of leachate from Phase 1 and the Lower Leachate Lagoon.
- Operation of the Phase 1 North Purge Well System continued to mitigate leachate impacts north of the landfill, as indicated by long-term stable or decreasing concentrations of leachate indicator parameters at locations 40, 20 and 21. However, nitrate concentrations at location 40 increased considerably from the previous monitoring year.
- Groundwater quality in proximity to the Phase 2 Basin confirms the hydraulic trap leachate collection system is effectively containing leachate north of Phase 2 . Groundwater quality 100 m north of Phase 2 continued to show low concentrations of leachate indicator parameters, indicating groundwater quality is not affected by landfill leachate. The increase in nitrate and sulphate concentrations in groundwater is interpreted to be due to runoff from aggregate stockpiles and roads constructed with aggregate.
- Along the northern edge of the Phase 2 Basin, groundwater quality is primarily impacted by runoff from aggregate stockpiles, with some wells (GW-106-1-1 and GW-105-1-1) showing evidence of dilute landfill leachate impacts. The influence of landfill leachate on groundwater is limited to an area less than 20 meters northwest of the Phase 2 Basin.
- Groundwater quality at Boundary Compliance Station 31 met all applicable CSR standards in 2022/23. However, highly elevated sulphate and nitrate concentrations observed at this location reflect aggregate runoff.


## Hartland North Pad

- Groundwater quality at the Hartland North Pad was slightly deteriorated, with elevated conductivity, nitrate, and sulphate concentrations observed at some monitoring stations (e.g., GW-44-1-1, GW-62-1-1, GW-77-1-1, GW-78-1-1,

GW-87-1-1, GW-88-1-1). The concurrent increase in conductivity, sulphate, and nitrate concentrations suggests widespread impacts of aggregate runoff on shallow groundwater quality.

- Groundwater quality in GW-91-1-1 and GW-92-1-1 was generally consistent with background conditions, except for elevated conductivity and sulphate concentrations. Nitrate concentrations remained low at both sites. The elevated sulphate may be associated with natural sulphide oxidation. Groundwater at location 94 was clearly impacted by aggregate runoff, indicating it cannot be considered a background station.


## South of the Landfill

- Groundwater quality south of the landfill met all applicable CSR standards. Although ammonia concentrations in some wells south of the landfill were slightly elevated, they were within historical ranges and well below the applicable CSR standards.
- Groundwater quality at several locations (e.g., GW-85-1-1, GW-60-1-1, and GW-71-1-1) showed no evidence of landfill leachate impacts. However, elevated nitrate and moderate sulphate concentrations reflect impacts from aggregate stockpiling and use, which may be related to runoff from the paved area and wind-blown or transported aggregate dust. Since 2020, chloride concentrations have been occasionally elevated at some monitoring stations (e.g., s GW-85-1-1 and GW-60-1-1), but the elevated chloride concentrations have not correlated with elevated ammonia concentrations. High $\mathrm{CI} / \mathrm{Na}$ molar ratios ( $>1$ ) suggest there is an additional source of chloride other than road salt, but the source of chloride is currently unknown.


## East of the Landfill

- Water quality along the east boundary of the Phase 1 landfill was consistent with previous years, and concentrations of all parameters were below applicable CSR standards. However, elevated sulphate and nitrate concentrations were observed at Site 16, reflecting the influence of aggregate runoff from the Northeast Stockpile.


### 9.4 Domestic Well Water Quality

As part of the CRD's groundwater quality monitoring program, sixteen (16) domestic wells within a 2 km radius of the landfill were sampled in 2022/23. The groundwater quality data was consistent with historic results, meeting all applicable federal and provincial drinking water quality guidelines (CDWQ and SDWQG). This indicates that offsite domestic water wells continue to remain unimpacted by landfill leachate.

### 9.5 Surface Water Quality

Surface water quality data collected in 2022/23 confirmed that nearby surface water bodies, including Tod Creek, Durrance Lake and Durrance Creek and Killarney Lake continued to be unimpacted by landfill leachate. However, surface water quality monitoring stations at the landfill continued to show signs of water quality degradation, especially in the area northwest of Phase 2.

In 2021/22, dissolved copper concentrations exceeded BCWQG-STA values at 8 stations. In 2022/23, all copper concentrations met the BCWQG-STA, except for SW-S-12 in January 2023.

## North of the Landfill and Downstream of the North Pad

- Surface water at Boundary Compliance location SW-N-05 continued to exhibit elevated nutrient concentrations in 2022/23, resulting in non-compliant conditions. Nitrate concentrations at Sw-N-05 exceeded BCWQG-STA during the May 2022 and November 2022 sampling events. The elevated nitrate and sulphate concentrations suggest an impact on surface water from quarrying, aggregate production, stockpiling and use. However, the absence of paired ammonia and chloride concentrations indicates the water was not impacted by leachate. The occasionally elevated ammonia concentrations may be associated with nitrate reduction via denitrification under reducing conditions.
- Surface water quality at Boundary Compliance station Sw-N-16 met BCWQG-STA, except for one total iron during February 2023 sampling event. The iron exceedance observed at $\mathrm{Sw}-\mathrm{N}-16$ was likely due to the disturbance of sediment during sampling, as indicated by the reported high TSS concentrations. Surface water quality at Sw-N-16 was not impacted by leachate, but continued to exhibit minor influence from nearby construction activities involving blasting, aggregate production, transport and placement, and excavation of organic soils. Similar aggregate impacts were observed downstream at stations Sw-N-17 and Sw-N-45.
- Surface water at Sw-N-18 reflected dilute landfill leachate impacts and may also have been impacted by aggregate runoff. In November 2022, the plug and diversion measures at surface water station SW-N-18 were removed, directing the discharge to the NWSP. Work is currently underway to install a diversion pipe at station SW-N-18 to direct the contaminated water into the leachate collection system.
- In the Hartland North Pad area, surface water quality at Boundary Compliance stations (Sw-N-41s1 and Sw-N-42s1) met BCWQG-STA in 2022/232, except for TSS. Leachate indicator parameters remained low in 2022/23, indicating surface water was not impacted by landfill leachate or construction activities. The slightly elevated sulphate, nitrate and conductivity concentrations indicate continued minor impacts from aggregate production, stockpiling and use.
- Historically, surface water stations Sw-N-14 and Sw-N-CS2 were used to monitor background conditions north of the landfill, but elevated conductivity, nitrate, and sulphate concentrations observed in 2022/23 indicate that surface water quality at these stations has been impacted by aggregate runoff and they are no longer suitable for use as background monitoring locations. Sw-S-52 consistently showed no signs of impacts related to the landfill, confirming that the water quality remains representative of the background conditions south of the landfill.
- Surface water quality downgradient of the North Pad (Sw-N-41s3) exhibited slightly elevated nitrate concentrations ( 1.28 to $1.57 \mathrm{mg} / \mathrm{L}$ ) corresponding to low to moderate sulphate concentrations ( 17.0 to $20.0 \mathrm{mg} / \mathrm{L}$ ). In the absence of elevated sulphate concentrations, it is difficult to interpret whether the elevated nitrate concentrations reflect a background process, or dilute runoff from aggregate stockpiles originating at Hartland landfill. Historically, nitrate concentrations have generally been below $0.2 \mathrm{mg} / \mathrm{L}$, but they occasionally elevated to a peak level of $10 \mathrm{mg} / \mathrm{L}$ in 2007 . Further downstream of $\mathrm{Sw}-\mathrm{N}-41 \mathrm{~s} 3$, water quality at Sw-N-41s4 was consistent with background conditions, and showed no signs of aggregate or leachate impacts.
- Further downstream to the north of the landfill, at the confluence of Durrance Creek and Tod Creek (Sw-N-64 and Sw-$\mathrm{N}-65$ ), surface water quality showed no impacts from landfill leachate or aggregate runoff. The slightly elevated nitrate ( $<1.5 \mathrm{mg} / \mathrm{L}$ ) and sulphate ( $<20 \mathrm{mg} / \mathrm{L}$ ) concentrations at $\mathrm{Sw}-\mathrm{N}-63$ may have originated from the on-site aggregate runoff, or may be associated with the application of fertilizers to the surrounding agricultural lands.


## South of the Landfill

- Water quality at the Boundary Compliance location (Sw-S-04) met the BCWQG-STA values for all analytes in all samples collected during 2022/23, except for one total iron concentration observed in May 2022. Surface water quality along the south boundary was not impacted by leachate but exhibited impacts from dilute aggregate runoff.
- Water quality at Sw-S-52 (not a compliance location) was representative of background water quality. In 2022/23, concentrations of all parameters were below the BCWQG-STA. Concentrations of leachate indicator parameters were consistent with previously reported values.
- Surface water quality south of the recycling area (Sw-S-03, Sw-S-12) exhibited several BCWQG-STA exceedances, including TSS, dissolved and total iron, and chloride during one or more sampling date. Elevated ammonia, nitrate, conductivity and sulphate concentrations at these stations may be related to aggregate dust from the south face of Phase 1 and runoff from paved areas surrounding the bin facility that experiences heavy traffic and several industrial activities.


### 9.6 Leachate Quality

In 2022/23, the leachate quality observed in the Hartland Valve Chamber followed the requirements of the Waste Discharge Authorization, except for COD exceedances on multiple sampling dates. Based on discussions with the analytical laboratory, CRD confirmed that the noted COD exceedances were due to the use of compromised/expired preservatives that were provided to CRD by the laboratory, and the exceedances do not likely reflect in-situ leachate quality. Overall, average annual leachate concentrations in 2022/23 were comparable with those measured in 2021/22.

### 9.7 Quality Assurance and Quality Control

Upon review of the quality assurance and quality control data collected in 2022/23, groundwater, surface water and leachate sampling and laboratory analysis have produced reliable results that are acceptable for the purposes of this monitoring report.

### 9.8 Compliance with Operating Certificate and Waste Discharge Authorization

Groundwater quality, surface water quality, and leachate quality data were used to assess compliance with the Amended Operational Certificate and Waste Discharge Authorization and are discussed individually below.

### 9.8.1 Groundwater

A total of 36 groundwater monitoring wells were identified as Boundary Compliance Monitoring Wells. Water quality data collected from these wells were compared to the CSR standards for the protection of freshwater aquatic life and drinking water to assess compliance with the landfill Operating Certificate and protect both current and future uses of the groundwater resource.

With respect to groundwater quality, Hartland Landfill remained in compliance with the Operational Certificate in 2022/23 except for one (1) copper exceedance at Boundary Compliance location 21. However, this recorded exceedance may not be representative of in-situ groundwater quality due to high RPD discrepancy. Dissolved copper was not detected in the parent sample but was highly elevated in the duplicate sample. Overall, the copper exceedance is unrelated to landfill activities, as indicated by low concentrations of parameters associated with aggregate runoff and leachate.

### 9.8.2 Surface Water

A total of five (5) surface water monitoring stations have been identified as Boundary Compliance stations surrounding Hartland Landfill. These stations are concentrated along the southern and northern property boundaries, downgradient of areas that have the potential to be impacted by leachate or landfill runoff. Water quality data collected from the Boundary Compliance stations were compared to the BCWQG-STA and BCWQG-LTC criteria to assess compliance with the Landfill Operational Certificate.

Some water quality impacts observed at the Boundary Compliance stations were caused by sources other than landfill leachate or aggregate runoff, including turbid samples collected under low-flow conditions and ongoing construction activities. In 2022/23, surface water quality was slightly deteriorated, exhibiting widespread impacts related to aggregate production and stockpiling. Throughout the monitoring year, highly elevated conductivity, sulphate, nitrate and/or ammonia concentrations consistent with aggregate runoff were observed at Boundary Compliance stations Sw-N-05, Sw-N-14, and Sw-N-16. Nitrate concentrations exceeded the BCWQG-STA at Sw-N-05 during the May and November 2022 sampling events. Additionally, moderately elevated conductivity, sulphate, and nitrate concentrations at Sw-N-41s1 and Sw-N-42s1 indicate minor impacts from aggregate production, stockpiling and use. Ultimately, in 2022/23, surface water quality at Sw-N-05 was not compliant with the Landfill Operational Certificate. Table 9-1 summarizes BCWQG-STA exceedances observed at Hartland in 2022/23.

Table 9-1. $\quad$ Surface Water Quality Compliance at Property Boundary Stations

| Station | General <br> Parameters | Nutrients | Metals | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: |
| North of the Landfill | Nitrate (2) | None | • <br> Sw-N-05 <br> None <br> Nitrate exceedances are associated with aggregate production and <br> stockpiling at Hartland. The nitrate originates from leaching of blasting <br> residue left on the aggregate after blasting. |  |
| Sw-N-16 | None | None | Total Iron (1) | Exceedances are anticipated to be related to turbid flow conditions <br> following a prolonged dry period. Continued monitoring to assess these <br> anomalous results. |
| SW-S-04 | TSS (1) | None | Total Iron (1) | Exceedances are anticipated to be related to turbid flow conditions <br> following a prolonged dry period. Continued monitoring to assess these <br> anomalous results. |

### 9.8.3 Leachate

The Hartland Valve Chamber is the Compliance Monitoring Station for the Waste Discharge Authorization. During the monitoring period, leachate discharges at the Hartland Valve Chamber were in compliance with the Waste Discharge Authorization requirements due to the conclusion that noted COD exceedances were attributed to compromised/expired preservatives provided to CRD by the laboratory, and the exceedances do not likely reflect in-situ leachate quality.

## 10. Recommendations

Based on the findings of this report, our recommendations are summarized in Table 10-1:
Table 10-1. Summary of Recommendations

|  | Leachate Collection System | Status |
| :---: | :---: | :---: |
| 1 | Closely monitor water levels and leachate quality in the north purge wells to verify the effectiveness of the leachate collection system. Water levels in well 52-3-0, adjacent to 52-4-0-P7 have slowly increased since 2021/22 and may indicate diminished drawdown and leachate collection in this area. A step test should be conducted on each north purge well to measure the specific capacity which is an indicator of well performance. The measurements should be compared to historical assessments to determine the need for well rehabilitation. Options for maintaining lower leachate levels in P7 and P9 should be further investigated to continue improving groundwater quality west of the lower leachate lagoon. | New/Ongoing |
| 2 | Closely monitor water levels and leachate quality in the south purge wells to verify the effectiveness of the leachate collection system and identify opportunities for improvements. Several pump failures were reported for south purge wells P3 and P10. Increased water levels above operational targets were observed in P1, reaching a peak of 150.6 m . Pumping elevations in the south purge wells (P2, P3, P4 and P10) should be maintained at elevations below 140 m asl. Pumping elevations in P1 should be maintained near the bottom of the screened interval around 146 m asl. | New/Ongoing |
| 3 | Periodically validate the pumping levels and the extent of the drawdown cones surrounding the north and south purge well systems (next assessment in 2024) to confirm the proper functioning of the purge wells. All procedures should follow the Standard Operating Procedure (SOP) - North Purge Well Drawdown Cone Verification (AECOM 2016), with interpretation of results by a qualified professional. Water levels in purge wells and pump maintenance should be conducted regularly to confirm the efficiency of the purge wells. | Ongoing |
| 4 | Conduct a detailed assessment of the effectiveness of the hydraulic trap and leachate collection systems including the north purge wells and south purge wells based on the design of the Phase $4 / 5 / 6$ quarry and liner system. This is required to confirm the landfill will perform as intended as the landfill extends further north and west, and as additional lifts are constructed. Recent groundwater and surface water characterization between the Phase 2 Basin and the Northwest Sedimentation Pond suggests additional leachate containment or groundwater management measures need to be implemented to mitigate the potential for off-site leachate migration and non-compliant water quality at Sw-N-05. | Ongoing |
| Runoff and Infiltration Associated with Aggregate Stockpiles |  |  |
| 5 | Update the aggregate impact indicator parameters and thresholds based on recent geochemical testing results for aggregate samples and recommendations of the Aggregate Management Plan that is presently being developed. | New |
| 6 | Minimize the spatial extent and volume of aggregate stockpiles outside of the leachate collection system. Where this is not feasible, stockpiles should be covered with low permeability temporary tarps as soon as practical to minimize sulphate, ammonia, nitrate and TSS impacts on downgradient groundwater and surface water quality. Direct runoff from aggregate stockpiles away from natural water courses as it is known to exceed BCWQ guidelines for sulphate and some nitrogenous compounds. This approach proved to be effective for mitigation of historical aggregate impacts at the Hartland North Pad. | Ongoing |
| Groundwater Monitoring Program |  |  |
| 7 | Advance a network of boreholes into the bedrock slope west of the Phase 2 landfill to characterize the geology, hydrogeology and groundwater quality. This will also allow for establishment of a longterm groundwater monitoring network west and upslope of the Phase 2 landfill to support continued evaluations of hydraulic trap performance and monitor groundwater quality. | Ongoing |
| 8 | Groundwater monitoring wells in proximity to the Phase 2 Basin and Northwest Sedimentation Pond should be closely monitored to confirm the hydraulic trap leachate collection system is effectively containing leachate north of Phase 2. Continued quarrying may result in greater connection between the groundwater flow system on the west side of the Highland Fault and the Phase 2 hydraulic trap, resulting in lowering of the surrounding groundwater levels and increased leachate generation. | New/Ongoing |
| 9 | Decommission any monitoring wells that will be affected by quarrying, aggregate stockpiling, landfill development in advance of any damage to satisfy the requirements of the British Columbia Groundwater Protection Regulation. Based on near term construction activities, it is anticipated that monitoring wells 27-1-1, 78-1-1, and 78-2-1 will be inevitably impacted. Destroyed monitoring wells including 27-1-2and 93-1-1 should also be decommissioned. | New |
| 10 | Establish a new background groundwater monitoring well further upgradient of the Northwest Stockpile to replace 94-1-1. Water quality in well $94-1-1$ is no longer considered representative of | New |


|  | Leachate Collection System | Status |
| :---: | :---: | :---: |
|  | background groundwater quality. It is possible that this monitoring well could be coordinated with investigation of the bedrock hydrogeology west of the Phase 2 landfill. |  |
| 11 | The existing Groundwater, Surface Water and Leachate Monitoring Plan should be updated to ensure that it remains effective in monitoring the impacts current and future of landfill operations, including aggregate production, stockpiling, transport and use. This work is currently underway. | Ongoing |
| 12 | Groundwater wells should be surveyed once every five years to verify well condition and ensure geodetic well elevations are accurate (i.e., next survey in 2025). | Ongoing |
| 13 | The elevation of the leachate mound in Phase 1 and 2 should be determined at least once every five years (i.e. next assessment in 2025). | Ongoing |
| 14 | Conduct a review of the landfill development plan and filling plan every two years to ensure the existing monitoring network and monitoring program remain sufficient and interpretation of the data benefits from a complete understanding of the landfill design and operations over the next five years. The next review should be conducted in 2024 following completion of the Phase 4/5/6 quarry and liner design. | Ongoing |
| 15 | As required by the Amended Operational Certificate, the results of the annual monitoring program should continue to be reviewed and interpreted by a Qualified Professional experienced in assessing the impacts of landfill leachate at large municipal landfills similar to Hartland. | Ongoing |
|  | Surface and Leachate Monitoring Program |  |
| 16 | Add sodium to the surface water analytical packages. Analyzing sodium alongside chloride can help determine if elevated chloride concentrations originate from road salt application. | New |
| 17 | Establish a new background surface water station upgradient of the Phase 2 landfill to replace background water quality monitoring locations $\mathrm{Sw}-\mathrm{N}-14$ and $\mathrm{Sw}-\mathrm{N}-\mathrm{C} 52$ which are no longer representative of background conditions. | New |
| 18 | Surface water quality at locations Sw-N41s4, Sw-N-63, Sw-N-64 and Sw-N-65 should be sampled on quarterly basis to delineate the impact of aggregate runoff and assess its effect on the receiving environment. | New |
| 19 | Improve surface water flow monitoring upstream of Sw-N-05 in Heal Creek to ensure it provides an accurate measurement of surface water discharge from the landfill. Accurate flow measurements are important for evaluating environmental impacts and ensuring adequate collection and conveyance capacity. | New |
| 20 | Improve surface water management north of the Phase 2 landfill to minimize impacts of aggregate runoff on groundwater and surface water that is not captured by the leachate collection system. This may require lining of the NWSP and installation of an underdrain to allow for management of groundwater separately from surface water in the area. Additional sediment control measures and efforts to reduce the quantity of blasting residuals contained in aggregate stockpiles may help reduce impacts on water quality as quarry development becomes increasingly close to the northern property boundary and the water quality boundary compliance monitoring stations. | New |
| 21 | Characterize the chemistry of residual wastewater solids and stabilized biosolids (solids and leachate) to allow for future evaluation of any impacts to leachate chemistry. This information may be available from pilot studies or operational monitoring programs. | Ongoing |
| 22 | Determine the source of chloride, ammonia, dissolved copper and nitrate observed in surface water south of the Phase 1 landfill. Additional waste has been placed on the western and southern portions of Phase 2 over recent years and occasional leachate seeps and runoff from the truck wash facility have been noted in the past. Changes in activities at the south end of the landfill and management of impacted surface runoff may play a role. A multilevel monitoring well cluster should be established west of the bin facility and well 85-1-1 to resolve whether the source of impacts to surface water are due to runoff or discharge of leachate impacted groundwater. | New/Ongoing |
| 23 | Resume leachate sampling from the Phase 2 Cleanout as soon as the sampling pump is replaced. This information will be important for tracking changes in leachate chemistry as Phase 2 Cell 4/5/6 are developed. | Ongoing |
| 24 | In addition to the Sewer Use Criteria, leachate quality results for trace organic compounds should be compared to CSR standards for the protection of drinking water and aquatic life to allow for screening of data to identify parameters in leachate that exceed CSR standards and guide any refinements to the monitoring program in future years. Additionally, an updated list of emerging contaminants will be integrated into the monitoring program for the 2023/2024. | Ongoing |
|  | Construction Management |  |
| 25 | Blasting and quarrying activities should continue be to be conducted under the direction of a qualified blasting professional to minimize the potential for blast-enhanced fracturing, with possible negative impacts on hydraulic properties. This has been demonstrated to have important implications on groundwater elevations west of the Highland Fault and the volume of seepage reporting to the Phase 2 Basin as the base of the quarry has been lowered In circumstances where blasting might induce substantial topographic alterations or changes to the elevation of the base of | New |


| $\quad$ Leachate Collection System |  |  |  | Status |  |
| :---: | :--- | :---: | :---: | :---: | :---: |
| $\mathbf{2 6}$ | the Phase 2 quarry, consultation with a hydrogeologist is recommended to evaluate potential <br> implications on the performance of the hydraulic trap and the leachate collection system. |  |  |  |  |
| The placement of aggregate, road salt, dust suppressant and herbicides should be carefully <br> considered and documented to help understand the causes of potential future concentrations of <br> conductivity, ammonia, chloride, nitrate, sulphate and select metals at groundwater and surface <br> water monitoring locations. |  |  |  | Quality Assurance and Quality Control | Ongoing |
| Quality assurance for laboratory analyses should continue to be evaluated quarterly, and any <br> discrepancies should be resolved with the laboratory and CRD sampling personnel within one <br> month of receiving the laboratory results. The appropriate notation should be added to the data files <br> to explain the reason for the low precision and the steps taken, if any, to improve the sampling or <br> laboratory procedures. |  |  |  |  |  |

## 11. Qualifications of the Authors

Matt Martinolich, M.Sc. is a Hydrogeologist / Geochemist with almost two years experience collecting, analyzing, and interpreting hydrogeochemical and hydrogeological data at AECOM in addition to a year of experience in exploration geology. He has been involved in several geological and geochemical investigations at Hartland landfill. Matt contributed to data analysis and authored several sections of this report.

Kun Jia, M.Sc., P.Geo. is a Hydrogeologist / Geochemist with over nine years of experience collecting, analyzing and interpreting hydrogeological and geochemistry data for waste management, mining and contaminated sites projects. Kun has contributed to and authored several monitoring reports at Hartland landfill since 2015. Kun was the primary author and reviewer of this report.

Ryan Mills, M.Sc., P.Geo. is a Senior Hydrogeologist with over 20 years of experience interpreting and analyzing hydrogeological and water chemistry data for waste management, water resources and mining related projects. Ryan has authored several groundwater monitoring reports and conducted numerous site investigations involving drilling and hydrogeologic testing at Hartland Landfill since 2004. He has also undertaken site investigations at numerous other municipal, industrial and small rural landfills throughout British Columbia. Ryan was the senior reviewer of this report.

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## Appendix A

## Monitoring Station and Groundwater Level Data

- A1. Monitoring Well Co-ordinates
- A2. Groundwater Monitoring Plan
- A3. Groundwater Elevations
- A4. Surface Water Station Details

A1. Monitoring Well Co-ordinates

Appendix A-1. Monitoring Well Co-ordinates 2022-2023

| Station Name | Status | Location |  | Elevations |  |  |  |  | Depths |  |  |  | Survey Date | Monitor Class <br> Shallow (S), <br> litermediate <br> (I), Deep ( $(0)$ | Area of Landfill | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Northing (NAD 83) | Easting (NAD 83 ) | Ground Surface Elevatio | Top of Casing Elevation | Top of Piezomete Elevation $\qquad$ | Top of Screen <br> Elevation <br> m ASL | Bottom of <br> Screen Elevation <br> Elevation | Stickup |  | Depth to Top <br> of Screen <br> Below Cround <br> Surface$\|$ | Dopth to <br> Bottom of <br> Screen Below <br> Ground <br> Surface |  |  |  |  |
| 6W-04-3-1 | Active | ${ }_{5} 5375469.0$ | ${ }_{466167.8}$ | ${ }^{\text {mast }} 127.11$ | ${ }^{128.20}$ | ${ }_{1}^{128.09}$ | ${ }^{114.26}$ | ${ }^{\text {mast }} 111.21$ | ${ }_{1} 1.05$ | ${ }^{\text {mbabs }}$ | ${ }^{\text {mides }}$ | ${ }^{\text {m }} 15.90$ | 2005 Mar | I | SP1 |  |
| GW-04-4-1 | Active | 5375466.8 | 466166.1 | 127.13 | 128.08 | 128.01 | 122.98 | 119.93 | 0.88 | 7.20 | 4.15 | 7.20 | 2005 Mar | s | SP1 |  |
| GW-07-1-0 | Active | 5375613.3 | 466177.0 | 140.60 | 142.69 | 142.29 | N/A | N/A | 1.69 | No monitor | No monitor | No monitor | 2005 Mar | D | SP1 |  |
| GW-09-1-0 | Active | 5375774.0 | 466187.0 | 148.96 | 150.14 | 150.14 | N/A | N/A | 1.18 | No monitor | No monitor | No monitor | Historic | D | EP1 |  |
| GW-16-1-1 | Active | 5376345.6 | 466130.2 | 143.48 | 144.28 | 144.06 | 101.98 | 100.48 | 0.58 | 43.00 | 41.50 | 43.00 | 2005 Mar | D | EP1 |  |
| 6W-16-1-2 | Active | 5376345.6 | 466130.1 | 143.48 | 144.28 | 144.00 | 110.98 | 109.48 | 0.52 | 34.00 | 32.50 | 34.00 | 2005 Mar | D | EP1 |  |
| 6W-16-2-1 | Active | 5376347.1 | 466133.7 | 143.31 | 144.09 | 143.67 | 119.81 | 118.81 | 0.36 | 24.50 | 23.50 | 24.50 | 2005 Mar | 1 | EP1 |  |
| GW-16-2-2 | Active | 5376337.0 | 466133.7 | 143.31 | 144.09 | 143.72 | 129.81 | 126.81 | 0.41 | 16.50 | 13.50 | 16.50 | 2005 Mar | 1 | EP1 |  |
| 6W-17-1-1 | Active | 5376186.4 | 466198.0 | 150.99 | 152.17 | 152.08 | 100.49 | 98.99 | 1.09 | 52.00 | 50.50 | 52.00 | 2005 Mar | D | EP1 |  |
| $6 \mathrm{~W}-17-1-2$ | Active | 5376186.5 | 466198.0 | 150.99 | 152.17 | 152.11 | 110.99 | 109.49 | 1.12 | 41.50 | 40.00 | 41.50 | 2005 Mar | D | EP1 |  |
| 6W-17-1-3 | Active | 5376186.5 | 466198.0 | 150.99 | 152.17 | 152.04 | 136.49 | 133.29 | 1.05 | 17.70 | 14.50 | 17.70 | 2005 Mar | 1 | EP1 |  |
| GW-18-1-1 | Active | 53375976.5 | 466694.8 | 168.81 | 169.48 1698 | ${ }^{168.82}$ | 110.64 | 109.14 | 0.19 | 59.67 | 58.17 | 59.67 | ${ }^{2005 \mathrm{Mar}}$ | D | EP1 |  |
| GW-18-1-2 | Active | 5375976.5 | 466194.7 | 168.81 | 169.48 | ${ }^{169.33}$ | 122.61 | ${ }^{121.11}$ | 0.52 | 47.70 | 46.20 | 47.70 | 2005 Mar | D | EP1 |  |
| 6W-18-2-1 | Active | 5375973.0 | 466193.8 | 168.92 | 169.68 | 169.16 | 138.42 | 136.92 | 0.24 | 32.00 | 30.50 | 32.00 | 2005 Mar | D | EP1 |  |
| GW-18-2-2 | Active | 5375973.0 | 466193.7 | 168.92 | 169.68 | 169.12 | 155.92 | 152.92 | 0.20 | 16.00 | 13.00 | 16.00 | 2005 Mar | 1 | EP1 |  |
| 6W-19-1-1 | Active | 5377503.2 | 466125.3 | 132.89 | 133.86 | 133.85 | 96.89 | 95.39 | 0.96 | 37.50 | 36.00 | 37.50 | 2005 Mar | D | SP1 |  |
| GW-19-1-2 | Active | 5377503.2 | 466125.3 | 132.89 | 133.86 | 133.87 | 106.39 | 104.89 | 0.98 | 28.00 | 26.50 | 28.00 | 2005 Mar | 1 | SP1 |  |
| GW-19-2-1 | Active | 5375507.6 | 466124.1 | 132.60 | 133.37 | 133.26 | 117.10 | 115.60 | 0.66 | 17.00 | 15.50 | 17.00 | 2005 Mar | 1 | SP1 |  |
| 6W-19-2-2 | Active | 5375507.6 | 466124.1 | 132.60 | 133.37 | 133.32 | 126.60 | 123.60 | 0.72 | 9.00 | 6.00 | 9.00 | 2005 Mar | s | SP1 |  |
| 6W-20-1-1 | Active | 5376498.3 | 465971.1 | 110.46 | 111.32 | 111.17 | 80.46 | 77.46 | 1.20 | 33.00 | 30.00 | 33.00 | 2005 Mar | D | NP1 | Deactivated in 2010. |
| 6W-20-1-2 | Active | 5376498.4 | 465971.0 | 110.46 | 111.32 | 111.19 | 92.66 | 89.66 | 1.21 | 20.80 | 17.80 | 20.80 | 2005 Mar | 1 | NP1 |  |
| 6W-21-1-1 | Active | 5376483.9 | 465970.8 | 110.92 | 111.79 | 111.69 | 98.02 | 94.92 | 1.25 | 16.00 | 12.90 | 16.00 | 2005 Mar | 1 | NP1 |  |
| GW-21-1-2 | Active | 5376483.9 | 465970.9 | 110.92 | 111.79 | 111.68 | 105.42 | 102.32 | 1.24 | 8.60 | 5.50 | 8.60 | 2005 Mar | s | NP1 |  |
| GW-21-2-1 | Active | 5378482.3 | 465970.0 | 111.10 | 111.87 | 111.80 | Nolog | Nolog | 0.70 | NA | NA | NA | 2005 Mar | s | NP1 |  |
| 6W-25-1-1 | Active | 5377491.7 | 465713.9 | 129.91 | 130.89 | 130.77 | 106.91 | 105.41 | 0.86 | 24.50 | 23.00 | 24.50 | 2005 Mar | 1 | NP2 |  |
| 6W-25-1-2 | Active | 5376491.7 | 465714.0 | 129.91 | 130.89 | 130.78 | 125.61 | 123.41 | 0.87 | 6.50 | 4.30 | 6.50 | 2005 Mar | s | NP2 |  |
| 6W-27-1-1 | Active | 5376358.2 | 465455.8 | 141.09 | 141.91 | 141.57 | 118.09 | 116.59 | 0.48 | 24.50 | 23.00 | 24.50 | 2005 Mar | 1 | BKGND - WP2 |  |
| 6W-28-1-0 | Active | 5376503.6 | 465825.1 | 136.25 | 137.07 | 136.52 | N/A | N/A | 0.27 | NA | NA | NA | 2005 Mar | D | NP1 |  |
| GW-29-1-1 | Active | 5376563.3 | 465898.2 | 113.39 | 114.41 | 14.38 | 100.28 | 98.87 | 0.99 | 14.52 | 13.11 | 14.52 | 2005 Mar | s | NP1 |  |
| 6W-29-1-2 | Active | 5376563.3 | 465898.3 | 113.39 | 114.41 | 114.39 | 110.39 | 105.96 | 1.00 | 7.43 | 3.00 | 7.43 | 2005 Mar | s | NP1 |  |
| 6W-30-1-1 | Active | 5376562.2 | 465978.4 | 109.84 | 110.89 | 110.79 | 95.51 | 94.10 | 0.95 | 15.74 | 14.33 | 15.74 | 2005 Mar | 1 | NP1 |  |
| 6W-30-1-2 | Active | 5376562.3 | 465978.5 | 109.84 | 110.89 | 110.79 | 108.56 | 104.07 | 0.95 | 5.77 | 1.28 | 5.77 | 2005 Mar | s | NP1 |  |
| 6W-31-1-1 | Active | 5376555.2 | 466080.9 | 105.28 | 106.34 | 106.26 | 90.92 | 89.50 | 0.98 | 15.78 | 14.36 | 15.78 | 2005 Mar | 1 | NP1 |  |
| 6W-31-1-2 | Active | 5376555.2 | 466080.9 | 105.28 | 106.34 | 106.26 | 103.84 | 99.41 | 0.98 | 5.87 | 1.44 | 5.87 | 2005 Mar | s | NP1 |  |

Notes:


$\underset{\substack{\text { Shator Class } \\ 1}}{\text { Shato well } 155 \text { doeep }}$





Appendix A-1. Monitoring Well Co-ordinates 2022-2023

| Station Name | Status | Location |  | Elevations |  |  |  |  | Depths |  |  |  | Survey Date | Monitor Class <br> Shallow (S), <br> litermediate <br> (I), Deep ( $(0)$ | Area of Landfill | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Northing (NAD 83) <br> (NAD 83) | Easting (NAD 83) | Ground <br> Elevation <br> - ASL | $\begin{gathered} \text { Top of } \\ \text { Casing } \\ \text { Elevation } \end{gathered}$ | $\begin{gathered} \text { Top of } \\ \text { Piezometer } \\ \text { Elevation } \end{gathered}$ | $\begin{gathered} \text { Top of } \\ \text { Screen } \\ \text { Elevation } \end{gathered}$ | Bottom of Screen Elevation $\qquad$ | Stickup |  | $\begin{array}{\|c\|} \hline \begin{array}{c} \text { Depth to Top } \\ \text { of Screen } \\ \text { Below Ground } \end{array} \\ \text { Surface } \end{array}$ | $\substack{\text { Depth to } \\ \text { Botam of } \\ \text { Scren Below } \\ \text { Ground } \\ \text { Surface }}$ |  |  |  |  |
| ${ }^{6 W-36-1-1}$ | Active | 5376398.9 | 465778.3 | ${ }_{1} 130.21$ | 130.21 | ${ }^{131.51}$ | ${ }^{117.87}$ | ${ }^{114.87}$ | ${ }_{1} 1.30$ | ${ }^{15.34}$ | 12.34 | 15.34 | 2005 Mar | I | NP2 |  |
| 6W-36-2-1 | Active | 5376400.6 | 465776.5 | 130.00 | 131.11 | 131.07 | 90.68 | ${ }^{87.63}$ | 1.07 | 42.37 | 39.32 | 42.37 | 2005 Mar | D | NP2 |  |
| 6W-36-3-1 | Active | 5376401.7 | 465773.6 | 130.01 | 131.01 | ${ }^{130.96}$ | 115.01 | 112.01 | 0.95 | 18.00 | 15.00 | 18.00 | 2005 Mar | 1 | NP2 |  |
| 6W-37-1-1 | Active | 5377432.6 | 465725.6 | 129.59 | 130.12 | 129.98 | 117.35 | 114.35 | 0.39 | 15.24 | 12.24 | 15.24 | 2005 Mar | 1 | NP2 |  |
| 6W-37-2-1 | Active | 5376432.5 | 465727.8 | 129.92 | 130.64 | 130.60 | 89.47 | 86.42 | 0.68 | 43.50 | 40.45 | 43.50 | 2005 Mar | D | NP2 |  |
| 6W-37-3-1 | Active | 5378435.6 | 465726.8 | 129.95 | 130.75 | 130.63 | 119.72 | 115.15 | 0.68 | 14.80 | 10.23 | 14.80 | 2005 Mar | s | NP2 |  |
| 6W-38-1-1 | Active | 5376464.6 | 465797.2 | 131.90 | 132.46 | 132.31 | N/A | N/A | 0.41 | 18.29 | No monitor | 18.29 | 2005 Mar | 1 | NP2 |  |
| 6W-39-1-1 |  | 5376467.2 |  | 129.54 | 130.24 | 130.11 | 111.10 | 108.10 | 0.57 | 21.44 | 18.44 | 21.44 | 2005 Mar |  | NP2 |  |
| 6W-39-2-1 | Active | 5377466.3 | 465874.7 | 129.75 | 130.72 | 130.58 | 95.56 | 92.56 | 0.83 | 37.19 | 34.19 | 37.19 | 2005 Mar | D | NP2 |  |
| $6 \mathrm{~W}-40-1-1$ | Active | 5376432.1 | 465915.2 | 122.00 | 122.78 | 122.68 | 109.76 | 106.76 | 0.68 | 15.24 | 12.24 | 15.24 | 2005 Mar | 1 | NP1 |  |
| 6W-41-1-1 | Active | 5378852.1 | 465190.4 | 149.48 | 150.30 | 150.16 | 143.41 | 140.41 | 0.68 | 9.07 | 6.07 | 9.07 | 2005 Mar | s | HNP |  |
| 6W-42-1-1 | Active | 5376717.6 | 465534.9 | 138.81 | 139.45 | 139.33 | 133.02 | 129.97 | 0.52 | 8.84 | 5.79 | 8.84 | 2005 Mar | s | HNP |  |
| 6W-43-1-1 | Active | 5376883.8 | 465448.7 | 162.60 | 163.05 | 163.10 | 144.31 | 141.26 | 0.50 | 21.34 | 18.29 | 21.34 | 2007 Apr | 1 | HNP |  |
| $6 \mathrm{~W}-44-1-1$ | Active | 5376671.5 | 465322.3 | 161.46 | 162.02 | 161.89 | 153.84 | 150.79 | 0.43 | 10.67 | 7.62 | 10.67 | 2005 Mar | s | HNP |  |
| 6W-46-2-1 | Active | 5376075.5 | 466029.9 | 169.97 | 171.25 | 171.70 | 161.69 | 158.69 | 1.73 | 11.28 | 8.28 | 11.28 | 2006 Apr | s | P1 |  |
| 6W-46-3-1 | Active | 5376085.7 | 466035.7 | 169.83 | 172.46 | 172.38 | ${ }^{137.83}$ | 134.78 | 2.55 | 35.05 | 32.00 | 35.05 | 2006 Apr | D | P1 | Installed during 2005 |
| 6W-46-4-1 | Active | 5376078.5 | 466035.9 | 169.71 | 172.10 | 172.03 | 151.12 | 148.07 | 2.32 | 21.64 | 18.59 | 21.64 | 2006 Apr | 1 | P1 | Installed during 2005 |
| 6W-47-2-1 | Active | 5375888.1 | 465996.7 | 171.84 | 174.46 | 174.40 | 154.78 | 151.73 | 2.55 | 20.11 | 17.06 | 20.11 | 2006 Apr | 1 | P1 | Installed during 2005 |
| 6W-48-1-1 | Active | 5375840.5 | 466031.3 | 169.78 | 171.34 | 171.63 | 160.14 | 157.14 | 1.85 | 12.64 | 9.64 | 12.64 | 2006 Apr | s | P1 |  |
| 6W-48-2-1 | Active | 537815.8 | 466031.6 | 168.87 | 171.28 | 171.22 | 149.97 | 146.92 | 2.35 | 21.95 | 18.90 | 21.95 | 2006 Apr | 1 | P1 | Installed during 2005 |
| 6W-51-1-1 | Active | 5376475.1 | 466048.2 | 110.90 | 111.68 | 111.76 | 106.13 | 103.13 | 0.86 | 7.77 | 4.77 | 7.77 | 2005 Mar | s | NP1 |  |
| 6W-51-2-1 | Active | 5376474.1 | 466045.6 | 110.90 | 111.83 | 111.89 | 100.49 | 97.49 | 0.99 | 13.41 | 10.41 | 13.41 | 2005 Mar | s | NP1 |  |
| 6W-51-3-1 | Active | 5376473.3 | 466042.7 | 110.97 | 111.84 | 111.89 | 93.85 | 90.85 | 0.92 | 20.12 | 17.12 | 20.12 | 2005 Mar | 1 | NP1 |  |
| 6W-52-1-1 | Active | 5378406.0 | 465979.1 | 119.91 | 120.94 | 120.90 | 93.19 | 90.19 | 0.99 | 29.72 | 26.72 | 29.72 | Historic | I | NP1 |  |
| 6W-52-2-0 | Active | 5376391.0 | 465959.0 | 119.99 | 120.65 | 120.65 | N/A | N/A | 0.66 | NA | NA | NA | Historic | 1 | NP1 |  |
| 6W-52-3-0 | Active | 5376389.9 | 465948.8 | 119.8 | 120.4 | 120.4 | N/A | N/A | 0.57 | No monitor | No monitor | No monitor | Historic | I | NP1 |  |
| GW-52-40.0.P7 | Active | 5376388.0 | 465947.0 | 119.80 | 120.60 | 120.60 | N/A | N/A | 0.80 | 22.25 | No monitor | No monitor | Historic | 1 | NP1 | Purge well |
| 6W-53-1-1 | Active | 53765006.2 | 465761.3 | 130.84 | 131.81 | 131.92 | 114.15 | 110.88 | 1.08 | 19.96 | 16.69 | 19.96 | 2005 Mar | 1 | NP2 |  |
| 6W-54-1-1 | Active | 5376187.7 | 466226.9 | 154.63 | 155.58 | 155.65 | 107.91 | 104.91 | 1.02 | 49.72 | 46.72 | 49.72 | 2005 Mar | D | EP1 |  |
| 6W-54-2-1 | Active | 5376185.6 | 466225.5 | 154.69 | 155.62 | 155.66 | 118.55 | 115.55 | 0.97 | 39.14 | 36.14 | 39.14 | 2005 Mar | D | EP1 |  |
| 6W-54-3-1 | Active | 5376183.4 | 466224.7 | 154.66 | 155.50 | 155.53 | 138.05 | 135.05 | 0.87 | 19.61 | 16.61 | 19.61 | 2005 Mar | 1 | EP1 |  |
| GW-55-1-1 | Active | 5376910.6 | 465136.1 | 147.67 | 147.68 | 148.52 | 139.06 | 134.56 | 0.85 | 13.11 | 8.61 | 13.11 | 2005 Mar | s | HNP |  |
| GW-56-1-1 | Active | 5376838.5 | 465287.8 | 148.67 | 149.69 | 149.61 | 139.92 | 131.29 | 0.94 | 17.38 | 8.75 | 17.38 | 2005 Mar | , | HNP |  |
| 6W-57-1-1 | Active | 5378873.9 | 465588.4 | 132.37 | 132.99 | 132.90 | 122.77 | 118.81 | 0.53 | 13.56 | 9.60 | 13.56 | 2005 Mar | s | HNP |  |
| GW-58-1-0 | Active | 537634.8 | 465822.9 | 137.17 | 138.30 | 138.23 | N/A | N/A | 1.06 | 19.20 | No monitor | No monitor | 2009 May | 1 | NP2 |  |



metres above fround divel
metres above enean sea elvel
Shalow well 15 m deep


SP1
EP1 $\begin{gathered}\text { Southo of Phase } 1 \text { Landitil } \\ \text { Eastot Phase } 1 \text { Lanatilit }\end{gathered}$



| $\substack{\text { HNP } \\ P 1 \\ \text { P2 }}$ | $\begin{array}{c}\text { Hatranad } \\ \text { fhase } \\ \text { Phase }\end{array}$ |
| :---: | :---: |

Appendix A-1. Monitoring Well Co-ordinates 2022-2023

| Station Name | Status | Location |  | Elevations |  |  |  |  | Depths |  |  |  | Survey Date | Monitor Class <br>  <br> Shallow (S), <br> Intermediate <br> (I), Deep (D) | Area of Landfill | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Northing (NAD 83) | Easting (NAD 83) | Ground Surface Elevation | $\begin{gathered} \text { Top of } \\ \text { Casing } \\ \text { Elevation } \end{gathered}$ | $\begin{gathered} \text { Top of } \\ \text { Piezometer } \\ \text { Elevation } \end{gathered}$ | Top of Screen Elevation | Bottom of <br> Screen <br> Elevation | Stickup | Borehole Depth Beiow Ground Surface Surace | $\begin{array}{\|c\|} \hline \begin{array}{c} \text { Depth to Top } \\ \text { of çreen } \\ \text { Below Ground } \\ \text { Surface } \end{array} \\ \hline \end{array}$ | $\substack{\text { Depth to } \\ \text { Botaon of } \\ \text { Screen Below } \\ \text { Grund } \\ \text { Surface }}$ |  |  |  |  |
| 6W-60-1-1 | Active | ${ }_{5} 5375636.9$ | ${ }_{466137.2}$ | ${ }^{1411.63}$ | ${ }_{142.40}$ | ${ }_{142.32}$ | ${ }^{122223}$ | ${ }^{\text {mast }} 19.23$ | ${ }_{0} 0.69$ | ${ }^{\text {m2.40 }}$ | ${ }^{\text {m }} 19.40$ | ${ }^{\text {m } 22.40}$ | 2005 Mar | I | SP1 |  |
| 6W-60-2-1 | Active | 5375638.4 | 466137.4 | 141.61 | 142.46 | 142.40 | 129.51 | 126.51 | 0.79 | 15.10 | 12.10 | 15.10 | 2005 Mar | 1 | SP1 |  |
| GW-60-3-1 | Active | 5375640.0 | 466137.5 | 141.74 | 142.60 | 142.49 | 137.84 | 134.84 | 0.75 | 6.90 | 3.90 | 6.90 | 2005 Mar | s | SP1 |  |
| GW-66-1-1 | Active | 5375980.6 | 465523.1 | 212.00 | 213.06 | 212.95 | 146.40 | 143.40 | 0.95 | 68.60 | 65.60 | 68.60 | 2005 Mar | D | BKGND-WP2 |  |
| GW-62-1-1 | Active | 5376609.3 | 465265.5 | 183.44 | 184.25 | 184.25 | 161.32 | 159.82 | N/A | 23.70 | 22.12 | 23.62 | 2008 May | I | HNP | Survey elevations suspect. |
| GW-62-2-1 | Active | 5376610.5 | 465267.3 | 183.06 | 183.96 | 183.96 | 168.70 | 165.70 | N/A | 18.90 | 14.36 | 17.36 | 2008 May | 1 | HNP | Survey elevations suspect. |
| GW-63-1-1 | Active | 5375812.3 | 465609.5 | 197.24 | 198.18 | 198.09 | 168.44 | 165.44 | 0.85 | 31.80 | 28.80 | 31.80 | 2005 Mar |  | BKGND-WP2 |  |
| GW-63-2-1 | Active | 5375809.7 | 465610.5 | 197.21 | 198.11 | 198.03 | 186.71 | 183.71 | 0.82 | 13.50 | 10.50 | 13.50 | 2005 Mar | s | BKGND-WP2 |  |
| 6W-71-1-1 | Active | 5375643.3 | 466260.6 | 144.04 | 144.93 | 144.82 | 116.61 | 113.56 | 0.81 | 31.24 | 27.43 | 30.48 | 2005 Mar | D | SP1 |  |
| GW-71-2-1 | Active | 5375644.1 | 466259.9 | 144.04 | 144.92 | 144.81 | 127.04 | 123.98 | 0.80 | 20.10 | 17.00 | 20.06 | 2005 Mar | 1 | SP1 |  |
| 6W-71-3-1 | Active | 5375645.3 | 466259.1 | 144.05 | 144.95 | 144.90 | 137.04 | 134.00 | 0.85 | 10.10 | 7.01 | 10.05 | 2005 Mar | s | SP1 |  |
| GW-72-1-1 | Active | 5375670.6 | 466188.6 | 143.29 | 144.13 | 144.03 | 115.86 | 112.81 | 0.79 | 30.48 | 27.43 | 30.48 | 2005 Mar | D | SP1 |  |
| 6W-72-2-1 | Active | 5375671.7 | 466186.7 | 143.32 | 144.09 | 144.04 | 126.56 | 123.20 | 0.72 | 20.12 | 16.76 | 20.12 | 2005 Mar | 1 | SP1 |  |
| 6W-72-3-1 | Active | 5375672.7 | 466186.8 | 143.38 | 144.17 | 144.12 | 136.06 | 133.02 | 0.74 | 10.36 | 7.32 | 10.36 | 2005 Mar | s | SP1 |  |
| GW-73-1-1 | Active | 5375532.1 | 466184.1 | 134.52 | 135.47 | 135.31 | 106.91 | 103.86 | 0.86 | 30.66 | 27.61 | 30.66 | 2005 Mar | D | SP1 |  |
| 6W-73-2-1 | Active | 5375533.2 | 466184.1 | 134.50 | 135.40 | 135.36 | 117.40 | 114.38 | 0.86 | 20.12 | 17.10 | 20.12 | 2005 Mar | 1 | SP1 |  |
| 6W-73-3-1 | Active | 5375534.3 | 466184.1 | 134.48 | 135.37 | 135.31 | 127.46 | 124.42 | 0.83 | 10.06 | 7.02 | 10.06 | 2005 Mar | S | SP1 |  |
| 6W-75-1-1 | Active | 5376207.5 | 466035.3 | 154.82 | 155.62 | 155.97 | 127.72 | 121.72 | 1.15 | 33.10 | 27.10 | 33.10 | 2007 Apr | D | P1 |  |
| GW-76-1-1 | Active | 5375966.0 | 466228.4 | 171.08 | 171.79 | 171.69 | 123.23 | 117.13 | 0.61 | 61.90 | 47.85 | 53.95 | 2005 Mar | D | EP1 |  |
| GW-76-2-1 | Active | 5375967.5 | 466227.1 | 171.00 | 171.77 | 171.66 | 133.81 | 127.72 | 0.66 | 43.60 | 37.19 | 43.28 | 2005 Mar | D | EP1 |  |
| GW-76-3-1 | Active | 466226.0 | 5375968.8 | 171.0 | 171.8 | 171.7 | 145.1 | 142.0 | 0.72 | 29.00 | 25.91 | 28.96 | 2005 Mar |  | EP1 |  |
| 6W-77-1-1 | Active | 5376487.8 | 465536.8 | 155.04 | 155.66 | 155.63 | 120.60 | 117.55 | 0.59 | 40.26 | 34.44 | 37.49 | 2006 Apr | D | NP2 | Installed in 2006 |
| 6W-77-2-1 | Active | 5376485.8 | 465536.2 | 154.90 | 155.51 | 155.45 | 139.16 | 135.87 | 0.55 | 20.67 | 15.74 | 19.03 | 2006 Apr | , | NP2 | Installed in 2006 |
| GW-78-1-1 | Active | 5376498.8 | 465648.8 | 142.66 | 143.46 | 143.38 | 113.46 | 110.18 | 0.72 | 33.79 | 29.20 | 32.48 | 2006 Apr | D | NP2 | Installed in 2006 |
| 6W-78-2-1 | Active | 5376500.3 | 465646.7 | 142.59 | 143.40 | 143.36 | 132.42 | 129.14 | 0.77 | 14.11 | 10.17 | 13.45 | 2006 Apr | s | NP2 | Installed in 2006 |
| GW-80-1-0.P8 | Active | 5376397.4 | 465931.0 | 119.61 | 120.29 | N/A | N/A | N/A | N/A | 20.42 | No monitor | No monitor | 2008 May | 1 | NP1 | Purge well |
| 6W-81-1-0.P9 | Active | 5376409.0 | 465910.8 | 122.17 | 122.85 | N/A | N/A | N/A | N/A | 26.82 | No monitor | No monitor | 2008 May | 1 | NP1 | Installed in 2007 |
| GW-82-1-1 | Active | 5376257.5 | 465772.7 | 154.67 | 155.89 | 155.86 | 135.17 | ${ }^{131.50}$ | 1.18 | 23.77 | 19.50 | 23.17 | 2009 May | 1 | P2 | Installed in 2007 |
| GW-83-1-1 | Active | 5378353.3 | 465723.8 | 140.09 | ${ }^{141.40}$ | 141.44 | 124.97 | 121.92 | 1.35 | 18.90 | 15.12 | 18.17 | 2009 May | 1 | P2 | Installed in 2007 |
| 6W-85-1-1 | Active | 5375688.3 | 466068.7 | 149.09 | 150.09 | 150.07 | 142.99 | 139.95 | 0.98 | 9.14 | 6.1 | 9.14 | 2009 May | s | SP1 | Installed in March 2009 |
| GW-87-1-1 | Active | 5376451.7 | 465243.4 | 182.32 | 183.22 | 183.16 | 147.62 | 144.52 | 0.84 | 37.80 | 34.70 | 37.80 | 2015 Feb | D | NP2 | Installed in 2014 |
| 6W-87-2-1 | Active | 5376453.8 | 465242.8 | 182.36 | 183.26 | 183.24 | 165.56 | 162.56 | 0.88 | 21.90 | 16.80 | 19.80 | 2015 Feb |  | NP2 | Installed in 2014 |
| 6W-88-1-1 | Active | 5376467.2 | 465271.7 | 181.47 | 182.24 | 182.16 | 138.77 | 135.77 | 0.69 | 45.70 | 42.70 | 45.70 | 2015 Feb | D | NP2 | Installed in 2014 |
| 6W-88-2-1 | Active | 5376470.1 | 46527.9 | 181.29 | 182.16 | 182.10 | 168.49 | 165.49 | 0.81 | 16.80 | 12.80 | 15.80 | 2015 Feb | 1 | NP2 | Installed in 2014 |
| GW-89-1-1 | Active | 5376028.4 | 466043.4 | 169.25 | 169.45 | 169.45 | 138.62 | 135.57 | 0.19 | 30.18 | 30.63 | 33.68 | 2018 Dec | D | P1 | Installed in 208 |
| 6W-89--1 | Active | 5376026.8 | 466048.7 | 169.141 | 169.34 | 169.34 | 149.48 | 146.43 | 0.20 | 19.05 | 19.66 | 22.71 | 2018 Dec | 1 | P1 | Installed in 2018 |
| GW-90-1-1 | Active | 5376286.6 | 465651.3 | 151.876 | 152.69 | 152.69 | 118.34 | 115.30 | 0.81 | 38.40 | 33.54 | 36.58 | 2019 Oct | D | P2 | leachate mound monitoring well |
| GW-90-2-1 | Active | 5376286.8 | ${ }_{465652.3}$ | 152.00 | 152.75 | 152.75 | 136.84 | 133.79 | 0.75 | 14.94 | 15.16 | 18.21 | 2018 Dec | 1 | P2 | leachate mound monitoring well |
| GW-91-1-1 | Active | 5376604.9 | 465473.7 | 164.61 | 165.33 | 165.33 | 149.01 | 145.97 | 0.72 | 18.64 | 15.60 | 18.64 | 2019 Nov | I | HNP | Installed in 2019 |
| 6W-92-1-1 | Active | 5376535.4 | 465501.9 | 167.68 | 168.57 | 168.57 | 147.76 | 144.72 | 0.88 | 22.96 | 19.92 | 22.96 | 2019 Nov | 1 | HNP | Installed in 2019 |
| GW-94-1-1 | Active | 5376508.8 | 465129.5 | 203.65 | 204.32 | 204.32 | 163.87 | 160.83 | 0.67 | 42.82 | 39.78 | 42.82 | 2019 Dec | D | HNP | Installed in 2019 |
| W-95-1-1 | Active | 5376518.9 | 465698.3 | 128.44 | 129.36 | 129.36 | 126.00 | 124.48 | 0.92 | 4.84 | 2.44 | 3.96 | 2022 Apr | s | NP2 | Installed in April 2022 |


metres below ground level
mAGL
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metese sobve mean sealevel
$\stackrel{S}{5}$ Shallow well 15 m mdeed



$\underset{\substack{H N P \\ P 1}}{\substack{\text { Hatrana } \\ \text { phase } \\ \text { phase }}}$

| Station Name | Status | Location |  | Elevations |  |  |  |  | Depths |  |  |  | Survey Date | Monitor Class <br> Shallow (S), <br> litermediat <br> (I), Deep (D) | Area of Landfill | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Northing (NAD 83) | Easting (NAD 83) | $\begin{aligned} & \text { Ground } \\ & \text { Surface } \\ & \text { Elevation } \end{aligned}$ | $\begin{gathered} \text { Top of } \\ \text { Casing } \\ \text { Elevation } \end{gathered}$ | $\begin{gathered} \text { Top of } \\ \text { Piezometer } \\ \text { Elevation } \end{gathered}$ | $\begin{aligned} & \text { Top of } \\ & \text { Screen } \\ & \text { Elevation } \end{aligned}$ | Bottom of Screen Elevation | Stickup | Depth <br> Ground <br> Surface | Depth to Top <br> of Screen <br> Below Ground <br> Surface | Depth to <br> Bottom of <br> Screen Below <br> Ground <br> Surface |  |  |  |  |
| 6W-96-1-1 | Active | $\begin{gathered} \hline \mathbf{m} \\ \hline 5376510.3 \\ \hline \end{gathered}$ | $\frac{m}{\frac{m}{66595.6}}$ | $\begin{gathered} \hline \text { m ASL } \\ \hline 129.49 \\ \hline \end{gathered}$ | $\frac{\mathrm{m} \mathrm{ASL}}{130.45}$ | $\frac{\mathrm{m} \text { ASL }}{130.45}$ | ${ }_{\text {m ASL }}^{127.05}$ | ${ }_{\text {m ASL }}^{125.53}$ | ${ }_{0}^{\text {m AGL }} 0.95$ | ${ }_{4}^{\text {m BGS }} 4.43$ | ${ }_{2}^{\text {m BGS }}$ | ${ }_{\text {m BGS }}^{3.96}$ | 2022 Apr | s | NP2 | Installed in April 2022 |
| 6W-97-1-1 | Active | 5376503.5 | 465719.6 | 129.12 | 130.00 | 130.00 | 125.16 | 122.11 | 0.88 | 7.95 | 3.96 | 7.01 | 2022 Apr | s | NP2 | Installed in Appril 2022 |
| 6W-98-1-1 | Active | 5376501 | 465698.7 | 130.41 | 131.30 | 131.30 | 125.84 | 122.79 | 0.90 | 8.54 | 4.57 | 7.62 | 2022 Apr | s | NP2 | Installed in April 2022 |
| 6W-99-1-1 | Active | 5376427.6 | 465668.9 | 135.40 | 135.30 | 135.30 | 132.96 | 131.44 | -0.10 | 3.88 | 2.44 | 3.96 | 2022 Apr | s | NP2 | Installed in April 2022 |
| GW-100-1-1 | Active | 53776419.7 | 465668.6 | 135.50 | 135.44 | 135.44 | 130.01 | 128.49 | -0.06 | 7.02 | 5.49 | 7.01 | 2022 Apr | s | NP2 | Installed in April 2022 |
| GW-101-1-1 | Active | 5376414 | 465669.1 | 135.46 | 135.33 | 135.33 | 129.97 | 126.93 | -0.12 | 8.51 | 5.49 | 8.53 | 2022 Apr | s | NP2 | Installed in April 2022 |
| 6W-103-1-1 | Active | 53376400.9 | 465626.0 | 135.74 | 136.67 | ${ }^{136.67}$ | 129.95 | 128.42 | 0.92 | 8.25 | 5.79 | 7.32 | 2022 Apr | s | NP2 | Installed in April 2022 |
| GW-104-1-1 | Active | 5376438.121 | 465666.3 | 135.52 | 135.46 | 135.46 | 129.36 | 127.84 | -0.06 | 7.62 | 6.10 | 7.62 | 2022 Aug | s | NP2 | Installed in August 2022 |
| GW-105-1-1 | Active | 5376431.992 | 465673.0 | 135.51 | 135.42 | 135.42 | 126.58 | 123.53 | -0.09 | 12.19 | 8.84 | 11.89 | 2022 Aug | 1 | NP2 | Installed in August 2022 |
| GW-106-1-1 | Active | 5376423.369 | 465679.7 | 135.43 | 135.34 | 135.34 | 123.15 | 121.62 | -0.09 | 13.72 | 12.19 | 13.72 | 2022 Aug | 1 | NP2 | Installed in August 2022 |
| GW-107-1-1 | Active | 5376417.266 | 465684.4 | 135.54 | 135.40 | 135.40 | 121.88 | 120.35 | -0.14 | 16.76 | 13.52 | 15.05 | 2022 Aug | D | NP2 | Installed in August 2022 |
| GW-107-1-2 | Active | 5376417.266 | 465684.4 | 135.54 | 135.43 | 135.43 | ${ }^{129.33}$ | 128.27 | -0.11 | 16.76 | 6.10 | 7.16 | 2022 Aug | s | NP2 | Installed in August 2022 |
| GW-108-1-1 | Active | 5376441.45 | 465701.6 | 129.80 | 130.69 | 130.69 | 119.26 | 116.21 | 0.89 | 15.24 | 11.43 | 14.48 | 2022 Aug | D | NP2 | Installed in August 2022 |
| GW-109-1-1 | Active | 5376390.117 | 465635.3 | 134.14 | 134.91 | 134.91 | 123.94 | 122.41 | 0.77 | 13.11 | 10.97 | 12.50 | 2022 Aug | s | NP2 | Installed in August 2022 |
| GW-110-1-1 | Active | 5376365.13 | 465593.6 | 137.61 | 138.32 | 138.32 | 134.21 | ${ }^{132.68}$ | 0.71 | 5.94 | 4.11 | 5.54 | 2022 Aug | s | NP2 | Installed in August 2022 |
| VLGW-02-D | Active | 5375782.7 | 465984.2 | 168.58 | 168.93 | 168.79 | 155.58 | 151.58 | 0.21 | 22.00 | 13.00 | 17.00 | 2008 May | 1 | P1 | Landfill gas well |
| VLGW-03-D | Active | 5375776.6 | 465933.6 | 170.04 | 170.30 | 170.30 | 156.29 | 153.29 | 0.26 | 17.00 | 13.75 | 16.75 | 2014 Feb | 1 | P1 | Landifil gas well |
| VLGW-04-D | Active | 5375858.1 | 466056.5 | 169.24 | 169.95 | 169.90 | 156.24 | ${ }^{152.24}$ | 0.65 | 19.00 | 13.00 | 17.00 | 2008 May | 1 | P1 | Landilil gas well |
| VLGW-08-D | Active | 5376088.1 | 466113.6 | 164.42 | 165.28 | 165.14 | 147.41 | ${ }^{137.38}$ | 0.72 | 28.04 | 17.01 | 27.04 | 2008 May | 1 | P1 | Landill gas well |
| VLGW-11-S | Active | 5375996.8 | 466122.6 | 166.06 | 166.53 | 166.48 | 160.42 | 155.78 | 0.42 | 11.28 | 5.64 | 10.28 | 2008 May | 1 | P1 | Landill gas well |
| VLGW-15-D | Active | 5375842.9 | 465997.0 | 170.61 | 171.13 | 171.07 | 156.41 | 148.21 | 0.46 | 23.31 | 14.2 | 22.4 | 2008 May | D | P1 | Landilil gas well |
| VLGW-16-D | Active | 5375901.0 | 466011.3 | 172.47 | 173.38 | 173.31 | 157.02 | 147.48 | 0.84 | 25.90 | 15.45 | 24.99 | 2008 May | 1 | P1 | Landill gas well |
| VLGW-17-D | Active | ${ }^{53775959.1}$ | 466025.5 | 171.40 | ${ }^{1717.81}$ | 1717.77 | ${ }^{157.20}$ | 144.28 | 0.37 | ${ }^{28.04}$ | 14.2 | ${ }^{27.12}$ | 2008 May | D | $\mathrm{P}_{1}$ | Landfill gas well |
| VLGW-18-D | Active | ${ }^{5376017.5}$ | 466039.7 | 170.14 | ${ }^{170.78}$ | ${ }^{170.61}$ | ${ }^{152.94}$ | ${ }^{1414.80}$ | 0.47 | ${ }^{29.25}$ | ${ }_{172}^{17.2}$ | ${ }_{28}^{28.34}$ |  | D | ${ }_{\text {P1 }}$ | Lanatill aga well |
| VLGW-19-D | Active | 5376076.2 | 466054.1 | ${ }^{168880}$ | 169.53 | 169.21 | ${ }^{151.60}$ | ${ }^{1414.37}$ | 0.42 | 28.34 | 17.2 | ${ }^{27.43}$ |  | 1 | P1 | Landifil gas well |
| VLGW-20-D | Active | 5376131.2 <br> 53761772 | 466062.8 | 167888 <br> 16481 | 169.05 16540 | 168.93 16530 | 148.03 14789 | 140.76 <br> 14043 <br> 1 | 1.05 | 28.04 <br> 2.93 <br> 2.80 | 19.85 16.92 | ${ }_{27}^{27.12}$ | 2008 May | I | ${ }_{\text {P1 }}$ | Landifil gas well |
| VLGW-21-D | Active | ${ }^{5376177.2}$ | 466082.0 | 164.81 | 165.40 16457 | 165.30 | 147.89 <br> 1450 | 140.43 | 0.49 | ${ }^{25.93}$ | 16.92 | ${ }_{2}^{24.38}$ | 2008 May | D | P1 | Landifil gas well |
| VLGW-26-D | Active | 5376165.6 | 466022.9 | 163.94 | 164.57 | 164.45 | 145.50 | 136.82 | 0.51 | 28.04 | 18.44 | ${ }^{27.12}$ | 2008 May | D | P1 | Landifil gas well |
| P1 | Active | 5375732.0 | 466026.0 | 157.60 | 155.17 | 1558.17 | 151.50 | 145.41 | 0.57 | 12.50 | 6.10 | 12.19 | 2018 Dec | s | P1 | Purge well, to replace the old well installed in 2005 |
| P2 | Active | 5375733.3 | 466030.9 | 157.37 | 158.72 | 158.62 | 146.87 | 132.37 | 1.25 | 25.00 | 10.5 | 25 | 2005 Mar | 1 | P1 | Purge well |
| P3 | Active | 5375739.0 | 466056.7 | 157.25 | 158.51 | 158.41 | 143.18 | ${ }^{132.25}$ | 1.16 | 25.00 | 14.07 | 25 | 2005 Mar | 1 | P1 | Purge well |
| P4 | Active | 5375751.7 | 466064.7 | 157.73 | 155.88 | 1558.78 | 146.21 | 132.73 | 1.05 | 25.00 | 11.52 | 25 | 2005 Mar | 1 | P1 | Purge well |
| P10 | Active | 5375731.5 | 466023.5 | 157.89 | N/A | 158.01 | 144.79 | 129.89 | 0.12 | 28.00 | 13.1 | 28 | 2010 October | D | P1 | Purge well |
| P11 | Active | 5376434.0 | 465685.9 | 135.01 | 135.69 | 135.69 | ${ }^{134.33}$ | 117.94 | 0.68 | 17.07 | 13.87 | 16.51 | 2022 Aug | D | NP2 | Installed in August 2022 |
| P12 | Active | 5376418.0 | 465702.6 | 135.22 | 135.86 | 135.86 | 134.58 | 116.47 | 0.64 | 18.75 | 13.56 | 14.78 | 2022 Aug | D | NP2 | Installed in August 2022 |
| VLGW-01-D | Inactive | 5375802.0 | 466040.8 | 168.25 | 168.86 | 168.79 | 163.25 | 156.25 | 0.54 | 14.00 | 5.00 | 12.00 | 2008 May | 1 | P1 | Landfill gas well |
| 6W-01-1-1 | Inactive | 5375781.9 | 468852.0 |  |  | 168.56 |  |  |  | 55.54 | 42 | 44.5 |  |  |  |  |
| GW-01-1-2 | Inactive | 5375781.9 | 465852.0 |  |  | 168.53 |  |  |  | 40.10 | 24.4 | 24.8 |  |  |  |  |
| 6W-01-1-3 | Inactive | 5375781.9 | 465852.0 |  |  | 168.53 |  |  |  | 29.53 | 13 | 14.75 |  |  |  |  |
| 6W-01-2-1 | Inactive | 5375787.9 | 468859.0 |  |  | ${ }^{168.25}$ |  |  |  | ${ }^{20.50}$ | 4.7 | 5.1 |  |  |  |  |
| 6W-01-3-1 | Inactive | 5375791.9 | 465851.0 |  |  | 167.81 |  |  |  | 21.02 | 5.11 | 5.91 |  |  |  |  |
| 6W-02-1-1 | Inactive | 5375813.9 | 465996.0 |  |  | 166.90 |  |  |  | 48.58 | 33 | 33.5 |  |  |  |  |
| ${ }^{6} \mathrm{~W}$-02-1-1-2 | Inactive | 5375813.9 | 465999.0 |  |  | 166.90 |  |  |  | 39.29 289 | $\begin{array}{r}23 \\ 12 \\ \hline\end{array}$ | $\begin{array}{r}24.6 \\ \hline 138\end{array}$ |  |  |  |  |
| ¢ ${ }_{\text {GW-02-1-3 }}$ | Inactive Inactive | 5375813.9 5375812.9 | $\begin{array}{\|l\|} \hline 465996.0 \\ \hline 465990.0 \end{array}$ |  |  | 166.90 167.14 |  |  |  | 28.79 22.53 | 12.2 6.25 | 13.8 7.5 |  |  |  |  |
| 6W-03-1-1 | Inactive | 5375684.9 | 466042.0 | 147.41 | 148.21 | 148.05 | 122.91 | 122.21 | 0.64 | 26.03 | 24.5 | 25.2 |  |  |  |  |







Appendix A-1. Monitoring Well Co-ordinates 2022-2023

| Station Name | Status | Location |  | Elevations |  |  |  |  | Depths |  |  |  | Survey Date | Monitor Class <br> Shallow (S), <br> Interneaiae <br> (I), Deep (0) | Area of Landfill | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Northing (NAD 83) | Easting (NAD 83) | Ground Surface Elevation | $\begin{gathered} \text { Top of } \\ \text { Casing } \\ \text { Elevation } \end{gathered}$ | $\begin{gathered} \text { Top of } \\ \text { Piezometer } \\ \text { Elevation } \end{gathered}$ | $\begin{gathered} \text { Top of } \\ \text { Screen } \\ \text { Elevation } \end{gathered}$ | Bottom of Screen Elevation | Stickup | Borehole <br> Depth Ground Surface | Depth to Top <br> of Screen <br> Below Ground <br> Surface | Depth to <br> Bottom of <br> Screen Below <br> Ground <br> Surface |  |  |  |  |
|  |  | m | m | m ASL | m ASL | m ASL | m ASL | m ASL | m AGL | m BGS | m BGS | m BGS |  |  |  |  |
| GW-03-1-2 <br> GW-03-1-3 | Inactive | 5375684.9 5375884.9 | 466042.0 466042.0 | 147.41 147.41 | 148.21 148.21 | 148.05 148.05 | 134.11 143.41 | 133.61 141.61 | 0.64 0.64 | 16.33 <br> 7.03 | $\stackrel{13.3}{4}$ | 13.8 <br> 5.8 |  |  |  |  |
| 6W-03-2-1 | Inactive | 5375685.5 | 466037.7 | 149.20 | 150.01 | 149.28 | 145.15 | 142.10 | 0.08 | 7.10 | 4.05 | 7.1 | 2005 Mar | s | SP1 |  |
| GW-04-1-1 | Inactive | 5375464.0 | 466171.0 | 126.67 | 128.12 | 128.07 | 96.47 | 95.77 | 1.40 | 32.00 | 30.2 | 30.9 |  |  |  |  |
| 6W-04-1-2 | Inactive | 5375464.0 | 466171.0 | 126.67 | 128.12 | 128.07 | 104.27 | 103.67 | 1.40 | 24.10 | 22.4 | 23 |  |  |  |  |
| GW-04-1-3 | Inactive | 5375464.0 | 466171.0 | 126.67 | 128.12 | 128.07 | 116.67 | 115.07 | 1.40 | 11.60 | 10 | 11.6 |  |  |  |  |
| GW-04-1-4 | Inactive | 5375464.0 | 466171.0 | 126.67 | 128.12 | 128.07 | 122.97 | 121.27 | 1.40 | 6.10 | 3.7 | 5.4 |  |  |  |  |
| 6W-04-2-1 | Inactive | 5375468.5 | 466169.3 | 127.14 | 127.95 | 127.86 | 107.49 | 104.44 | 0.75 | 22.70 | 19.65 | 22.70 | 2005 Mar | 1 | SP1 | Deactivated in 2016. |
| 6W-05-1-1 | Inactive | 5376444.9 | 465813.9 | 126.29 | 127.49 | 127.49 | 112.59 | 110.59 | 1.20 | 16.90 | 13.7 | 15.7 |  |  |  |  |
| 6W-05-1-2 | Inactive | 5376444.9 | 465813.9 | 126.29 | 127.49 | 127.49 | 118.69 | 116.69 | 1.20 | 10.80 | 7.6 | 9.6 |  |  |  |  |
| 6W-05-1-3 | Inactive | 5376444.9 | 465813.9 | 126.29 | 127.49 | 127.49 | 122.89 | 121.89 | 1.20 | 5.60 | 3.4 | 4.4 |  |  |  |  |
| GW-05-2-1 | Inactive | 5376426.9 | 465774.9 | 126.00 | 126.80 | 126.80 | 125.20 | 124.20 | 0.80 | 2.60 | 0.8 | 1.8 |  |  |  |  |
| GW-06-1-1 | Inative | 5376148.9 | 465661.9 | 127.70 | 128.41 | 128.40 | 107.00 | 104.90 | 0.70 | 23.50 | 20.7 | 22.8 |  |  |  |  |
| 6W-06-1-2 | Inactive | 5376148.9 | 465661.9 | ${ }^{1277.70}$ | 128.41 | 128.40 | 115.20 | 113.20 | 0.70 | 15.20 | 12.5 | 14.5 |  |  |  |  |
| 6W-06-1-3 | Inactive | 5376148.9 | 465661.9 | 127.70 | 128.41 | 128.40 | 120.80 | 118.80 | 0.70 | 9.60 | 6.9 | 8.9 |  |  |  |  |
| 6W-06-2-1 | Inactive | 5376151.9 | 465669.9 | 127.46 | 127.76 | 127.76 | 125.86 | 124.86 | 0.30 | 2.90 | 1.6 | 2.6 |  |  |  |  |
| 6W-08-1-1 | Inative | 5375867.9 | 465738.0 | 159.26 | 160.28 | 160.26 | 145.86 | 143.76 | 1.00 | 16.50 | 13.4 | 15.5 |  |  |  |  |
| 6W-08-1-2 | Inactive | 5375867.9 | 465738.0 | 159.26 | 160.28 | 160.26 | 152.11 | 150.11 | 1.00 | 10.15 | 7.15 | 9.15 |  |  |  |  |
| 6W-08-1-3 | Inactive | 5375867.9 | 465738.0 | 159.26 | 160.28 | 160.26 | 156.76 | 155.76 | 1.00 | 4.50 | 2.5 | 3.5 |  |  |  |  |
| GW-10-1-1 | Inactive | 5376108.9 | 465701.9 | 135.11 | 135.41 | 135.41 | 133.61 | 132.61 | 0.30 | 2.80 | 1.5 | 2.5 |  |  |  |  |
| 6W-11-1-1 | Inactive | 5376254.9 | 465770.9 | 134.13 | 135.13 | 135.13 | 131.23 | 130.23 | 1.00 | 4.90 | 2.9 | 3.9 |  |  |  |  |
| 6W-12-1-0 | Inactive | 5375701.0 | 466189.0 | 143.02 | ${ }^{143.82}$ | 143.82 | N/A | N/A | 0.80 | 3.40 | No monitor | No monitor | Historic | s | SP1 | Dug well, deactivated in 2010 |
| 6W-13-1-1 | Inactive | 5376166.9 | 466031.9 | 150.43 | 152.73 | 152.73 | 121.93 | 118.93 | 2.30 | 33.80 | 28.5 | 31.5 |  |  |  |  |
| 6W-13-1-2 | Inactive | 5376166.9 | 466031.9 | ${ }^{150.43}$ | 152.73 | 152.73 | ${ }^{126.83}$ | ${ }^{125.33}$ | 2.30 | 27.40 | 23.6 | 25.1 |  |  |  |  |
| 6W-14-1-1 | Inative | 5375968.9 | 465969.9 | 158.73 | 160.22 | 160.22 | 138.53 | 135.53 | 1.49 | 24.69 | 20.2 | 23.2 |  |  |  |  |
| 6W-15-1-1 | Inactive | 5375754.9 | 466007.0 | 159.78 | 160.82 | ${ }^{160.82}$ | 145.68 | 143.38 | 1.04 | 17.44 | 14.1 | 16.4 |  |  |  |  |
| 6W-22-1-1 | Inactive | 5376290.1 | 465751.7 | 138.54 | 138.81 | 139.86 | 111.22 | 109.69 | 1.32 | 28.09 | 27.32 | 28.85 |  |  |  |  |
| GW-23-1-1 | Inactive | 5376260.7 | 465718.7 | 128.37 | 129.52 | 129.50 | 111.51 | 110.09 | 1.13 | 17.57 | 16.86 | 18.28 |  |  |  |  |
| GW-23-1-2 | Inactive | 5376260.7 | 465718.7 | ${ }^{128.37}$ | ${ }^{129.52}$ | 129.49 | ${ }^{124.67}$ | 121.80 | 1.112 | 5.14 | 3.7 <br> 178 <br> 18 | ${ }_{6}^{6.57}$ |  |  |  |  |
| 6W-24-1-1 | Inative | 5376421.9 | 465753.0 | 127.77 | ${ }^{128.88}$ | 128.88 | 109.91 | 108.51 | 1.11 | 18.56 | 17.86 | 19.26 |  |  |  |  |
| GW-24-1-2 | Inactive | 5376421.9 | 465753.0 | 127.77 | 128.88 | 128.87 | 124.10 | 121.26 | 1.10 | 5.09 | 3.67 | 6.51 |  |  |  |  |
| GW-26-1-1 | Inactive | 5376319.8 | 465596.6 | 128.22 | 129.16 | 129.12 | 109.79 | 111.20 | 0.90 | 17.73 | 18.43 | 17.02 |  |  |  |  |
| GW-26-1-2 | Inactive | 5376399.8 <br> 5376358 | ${ }_{4}^{4655996.6}$ | 128.22 <br> 14120 | 129.16 14191 | 129.12 14156 | 124.36 14059 | 121.52 13759 | 0.90 | 5.28 <br>  | 3.86 | 6.7 6.7 | 05 Mar | s | BKGND-WP2 | Destoryed in summer 2022 |
| GW-32-1-1 | Inactive | ${ }^{53776512.9}$ | 4655599.9 | ${ }_{1}^{165.41}$ | 166.44 | 166.34 | ${ }^{123.85}$ | 122.44 | 0.93 | 4.2.27 <br> 4.27 | 4.56 | ${ }^{\text {42.97 }}$ |  |  | - | Destored ${ }^{\text {andmmer } 2022}$ |
| 6W-32-1-2 | Inactive | 5376129.7 | 465569.9 | 165.41 | 166.44 | 166.37 | 157.93 | 152.04 | 0.96 | 10.43 | 7.48 | 13.37 |  |  |  |  |
| 6W-33-1-1 | Inactive | 5376296.5 | 465627.4 | 124.55 | ${ }^{123.43}$ | 124.54 | 111.11 | 109.69 | -0.01 | 14.15 | 13.44 | 14.86 |  |  |  |  |
| 6W-33-1-2 | Inactive | 5376296.5 | 465627.4 | 124.55 | 123.43 |  | N/A | N/A | -124.55 |  | no monitor | no monitor |  |  |  |  |
| 6W-34-1-1 | Inactive | 5376216.8 | 465688.4 | 119.47 |  |  | 104.64 | 101.64 | -199.47 |  | 14.83 | 17.83 |  |  |  |  |
| 6W-34-2-1 | Inactive | 5376214.9 | 465668.9 | 119.67 |  |  | 115.97 | 112.97 | $-119.67$ |  | 3.7 | 6.7 |  |  |  |  |
| GW-35-1-1 | Inactive | 5376207.9 <br> 53762070 | 465677.3 | ${ }_{1}^{120.19}$ |  |  | 108.29 <br> 11713 | 105.29 11413 | $\xrightarrow{-120.19}$ |  | 11.9 3 | 14.9 |  |  |  |  |




$\underset{\substack{\text { ior Class } \\ 1}}{\substack{\text { Shalow well } 155 \mathrm{~m} \text { deep } \\ \text { Mnememediate well bew }}}$
Area of tandifill

${ }_{\substack{\text { SP1 } \\ \text { EP1 }}}$



P2 Phase 2

Appendix A-1. Monitoring Well Co-ordinates 2022-2023

| Station Name | Status | Location |  | Elevations |  |  |  |  | Depths |  |  |  | Survey Date | Monitor Class <br> Shallow (S), <br> Interneaiae <br> (I), Deep (D) | Area of Landfill | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Northing (NAD 83) | Easting (NAD 83) | $\begin{gathered} \text { Ground } \\ \text { Surface } \\ \text { Elevation } \end{gathered}$ | $\begin{gathered} \text { Top of } \\ \text { Casing } \\ \text { Elevation } \end{gathered}$ | Top of Piezometer Elevation | $\begin{gathered} \text { Top of } \\ \text { Screen } \\ \text { Elevation } \end{gathered}$ Elevation | Bottom of Screen Elevatio | Stickup | $\begin{aligned} & \hline \text { Borehole } \\ & \text { Depth } \\ & \text { Below } \\ & \text { Ground } \\ & \text { Surface } \\ & \hline \text { m BGS } \end{aligned}$ | $\substack{\text { Depth to top } \\ \text { of Screen } \\ \text { Below Ground } \\ \text { Surface }}$ | Dopth to <br> Bottom of <br> Screen Below <br> Ground <br> Surface |  |  |  |  |
| 6W-45-1-1 | Inative | ${ }_{5}^{5376163.1}$ | ${ }_{4661006.3}$ | ${ }^{\text {mast }} 16.37$ | mast | mast | ${ }^{1433.71}$ | ${ }^{140.71}$ | ${ }_{-162.37}$ |  | ${ }_{18.66}$ | ${ }_{21.66}$ |  |  |  |  |
| 6W-46-1-1 | Inative | 5376078.3 | 466029.2 | 170.48 | 171.65 | 171.96 | 140.68 | ${ }^{134.68}$ | 1.48 |  | 29.8 | 35.8 |  |  |  |  |
| GW-47-1-1 | Inative | 5375918.3 | 465992.6 | 174.85 | 177.24 | 177.37 | 155.09 | 152.09 | 2.52 | 22.55 | 19.76 | 22.758 | 2008 May | 1 | P1 |  |
| 6W-49-1-0 |  | 5376759.5 | 465218.7 | 160.00 | ${ }^{159.81}$ | 159.81 | N/A | N/A | -0.19 | ${ }^{62.20}$ |  |  |  |  |  |  |
| GW-50-1-1 | Inative | 5376480.9 | 466193.4 | 119.37 | 120.41 | 120.43 | 105.47 | 102.47 | 1.06 | 16.90 | 13.9 | 16.9 | 2005 Mar | 1 | EP1 |  |
| 6W-59-1-1 | Inative | 5376254.5 | 465636.2 | 125.62 | 126.40 | 126.40 | 113.72 | 112.22 | 0.78 | 13.40 | 11.9 | 13.4 |  |  |  |  |
| GW-64-1-1 | Inative | 5375855.0 | 465849.0 | 172.00 |  | 172.78 | 156.20 | 154.50 | 0.78 | 17.50 | 15.8 | 17.5 |  |  |  |  |
| GW-65-1-1 | Inactive | 5375831.0 | 465820.0 | 172.09 |  | 172.79 | 159.79 | 153.69 | 0.70 | 18.40 | 12.3 | 18.4 |  |  |  |  |
| GW-66-1-1 | Inative | 5376088.0 | 465902.0 | 164.23 |  | 165.08 | 157.03 | 150.93 | 0.85 | 13.30 | 7.2 | 13.3 |  |  |  |  |
| GW-67-1-1 | Inative | 5376256.2 | 465774.0 | 155.85 | N/A | 157.55 | 133.95 | 127.85 | 1.70 | 28.00 | 21.90 | 28.00 | 2005 Mar | 1 | P2 | Destroyed. Replaced by 6 W-82-1-1 |
| GW-67-2-1 | Inative | 5376256.2 | 465774.0 | 155.85 |  | 157.55 | 133.95 | 127.85 | 1.70 | 28.00 | 21.9 | 28 | 2005 Mar | 1 | P2 |  |
| 6W-68-1-1 | Inative | 53763353.0 | 465718.0 | 141.56 |  | 142.74 | 130.36 | 124.26 | 1.18 | 17.30 | 11.2 | 17.3 |  |  |  |  |
| GW-68-2-1 | Inactive | 5376353.0 | 465718.0 | 141.56 | N/A | 142.29 | 123.76 | 120.76 | 0.73 | 20.80 | 17.80 | 20.80 | Historic | 1 | P2 | Destroyed. Replaced by GW-83-1-1 |
| GW-69-1-1 | Inative | 5376267.6 | 465665.8 | 140.64 |  | 142.50 | 127.64 | 121.54 | 1.86 | 19.10 | 13 | 19.1 | 2005 Mar | 1 | P2 |  |
| 6W-69-2-1 | Inactive | 5376269.9 | 465667.0 | 140.67 |  | 142.11 | 121.67 | 115.67 | 1.44 | 25.20 | 19 | 25 | 2005 Mar | 1 | P2 |  |
| 6W-70-1-1 | Inative | 5376019.0 | 465891.0 | 167.00 |  | 167.59 | 149.50 | 143.40 | 0.59 | 23.60 | 17.5 | 23.6 |  |  |  |  |
| 6W-74-1-1 | Inactive | 5376013.4 | 466018.4 | 170.94 | 171.56 | 171.60 | 138.04 | 132.04 | 0.66 | 38.90 | 32.90 | 38.90 | 2007 Apr | D | P1 |  |
| GW-74-2-1 | Inative | 5376021.6 | 466032.5 | 170.29 | 172.25 | 172.40 | 153.19 | 150.19 | 2.11 | 20.10 | 17.10 | 20.10 | 2006 Apr | 1 | P1 | Installed during 2005 |
| GW-79-1-1 | Inactive |  | ${ }_{465404.3}$ | ${ }^{182887}$ | 1838.63 <br> 1839 | 188.59 <br> 1839 | 147.82 <br> 1578 | 144.77 | 0.72 | 39.62 <br> 308 | ${ }^{35.05}$ | 38.10 | 2007 Apr | D | NP2 | Installed during 2007, decommmissioned in May 2018 |
| ¢ ${ }_{\text {GW-79-7-1 }}$ | Inactive Inactive | 5376521.9 5376225.9 | 465403.1 465646.9 | 182.96 158.20 | 183.69 158.86 | 183.59 158.79 | 157.05 120.63 | 154.00 117.58 | 0.64 0.59 | 30.48 $\left.\begin{array}{l}37.74 \\ \hline\end{array}\right)$ | 25.991 37.55889 | ${ }_{40.6158889}$ | 2007 Apr <br> 2009 Feb | ! | ${ }_{\text {P2 }}$ | $\frac{\text { Installed during } 2007 \text { decommissioned in May } 2018}{\text { Installed during } 2007}$ |
| GW-84-2-1 | Inative | 5376244.8 | 465648.5 | 158.43 | 159.06 | 161.36 | 133.14 | 130.09 | 2.93 | 15.24 | 25.289839 | 28.339839 | 2009 Feb | s | P2 | Installed during 2007 |
| GW-86-1-1 | Inative | 5376228.4 | 465651.4 | 160.55 | N/A | N/A | N/A | N/A | N/A | 42.98 | N/A | N/A | 2010 October | D | P2 | SG piezo-42.42 mBGS |
| GW-86-1-2 | Inative | 5376228.4 | 465651.4 | 160.55 | N/A | N/A | N/A | N/A | N/A | 42.98 | N/A | N/A | 2010 October | 1 | P2 | SG piezo- 20.85 mbGS |
| GW-86-2-1 | Inative | 537624.8 | 465651.8 | 160.55 | N/A | N/A | N/A | N/A | N/A | 30.78 | N/A | N/A | 2010 October | D | P2 | SG piezo-29.69 mbGS |
| 6W-86-2-2 | Inative | 537624.8 | 465651.8 | 160.55 | N/A | N/A | N/A | N/A | N/A | 30.78 | N/A | N/A | 2010 October | 1 | P2 | SG piezo -8.67 mbGS |
| GW-93-1-1 | Inative | 5376556.5 | 465260.0 | 183.42 | 184.24 | 184.24 | ${ }^{159.34}$ | ${ }^{156.29}$ | 0.83 | 28.04 | 24.08 | 27.13 | 2019 Nov | D | HNP | Installed in 2019 |
| 6W-102-1-1 | Inative | 5376399.4 | 465663.5 | 134.46 | 135.01 | 135.01 | 129.89 | 128.36 | 0.56 | 6.65 | 5.47 | 6.10 | 2022 Apr | s | NP2 | Installed in Apil 2022, decomissione in Q3 2022 |
| P5 | Inactive | 5375775.2 | 466079.9 | 158.35 | 159.46 | 159.36 | 144.43 | 133.35 | 1.01 | 25.00 | 13.92 | 25 | 2005 Mar | 1 | P1 | Purge well |
| P6 | Inative | 5375803.6 | 466098.0 | 159.88 | 161.65 | 161.55 | 147.74 | 134.88 | 1.67 | 25.00 | 12.14 | 25 | 2005 Mar | 1 | P1 | Purge well |

Noles:
Daum Descripiotion
meteses below ground level

$\frac{\text { Montior Class }}{s}$ Shallow well 15 m deep
Invermedale wel beween 15 and 30 o doee
$\frac{D}{\text { Areao } \text { L Landiftl }}$
Intermediale well between 15 and 30 m deep
Deep well 30 mom deep

NP2 North tothasese Llanatiut



## A2. Groundwater Monitoring Plan

Appendix A-2. Groundwater Monitoring Plan 2022-2023

| Station Name | Status | $\begin{gathered} \text { Pipe } \\ \text { Diameter } \\ (\mathrm{mm}) \end{gathered}$ | Monitored in 2022-2023 |  |  |  | Development Method | Sampling Method | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Groundwater QualitySamples |  | Groudwater Level Measurements |  |  |  |  |
|  |  |  | Quarterly (4/yr) | Bi-annual (2/yr) | $\begin{array}{\|c\|c\|c\|c\|c\|c\|c\|c\|c\|c\|} \hline \text { measurement } \\ (4 \mathrm{yr}) \\ \hline \end{array}$ | Presssure <br> transducer <br> (continuous) |  |  |  |
| 6w-04-2-1 | Active | 50 |  |  | Y |  | NA | NA | Dedicated submersible pump installed January 2012. Well recharges very slowly since 2014, soit is no longer sampled. |
| Gw-04-3-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve | Submersile pump removed. Sampled using waterra from May 2019 onwards. |
| 6W-04-4-1 | Active | 50 | $Y$ |  | Y |  | footvalve | footvalve |  |
| 6W-07-1-0 | Active | 220 | Y |  | Y |  | well pump | well pump | Well pump replacement pending April 2023 |
| 6w-09-1-0 | Active | 150 |  |  | Y |  | NA | NA | Sampling discontinued in 2010 due to outdated pump and redundancy. Water levels only. |
| GW-16-11-1 | Active | 50 |  | Y | Y |  | footvalve | footvalve |  |
| GW-16-1-2 | Active | 50 |  | Y | Y |  | footvalve | footvalve |  |
| GW-16-2-1 | Active | 50 |  | Y | $r$ |  | footvalve | footralve |  |
| 6W-16-2-2 | Active | 50 |  | Y | Y |  | footvalve | footralve | 16--2-2 often dry, difificult to develop and sample. Sometimes use bailer. |
| 6W-17-1-1 | Active | 50 |  | Y | Y | Y | footvalve | footvalve | Pressure transducer installed in March 2019. |
| 6W-17-1-2 | Active | 50 |  | Y | Y |  | footvalve | footralve |  |
| GW-17-1-3 | Active | 50 |  | r | Y |  | footvalve | footvalve |  |
| GW-18-1-1 | Active | 50 | Y |  | Y | Y | sample pump | sample pump / footvalve | Dedicated submersible pump installed May 2009. Pump failed in September and November 2018. Waterra tubing from Feb/Mar 2019 onwards. Pressure transducer installed in March 2019. |
| GW-18-1-2 | Active | 50 |  |  | Y |  | sample pump | sample pump | Pump damaged and stuck in well since 2004. Not sampled, only water levels. |
| GW-18-2-1 | Active | 50 | Y |  | $r$ |  | footvalve | footvalve |  |
| GW-18-2-2 | Active | 50 | $Y$ |  | Y |  | footvalve | footralve |  |
| GW-19-1-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve | 19-1-1 sometimes flows artesian. |
| 6W-19-1-2 | Active | 50 | Y |  | Y |  | footvalve | footvalve | 19-1-2 sometimes flows artesian. |
| 6W-19-2-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve | 19-2-1 sometimes flows artesian. |
| 6w-19-2-2 | Active | 50 | $Y$ |  | $r$ |  | footvalve | footralve |  |
| GW-20-1-1 | Active | 50 | Y |  | Y |  | sample pump | footvalve | Dedicated submersible pump installed January 2012. Pump ceased functioning in Nov 2019, was removed for repair and is currently sampled using waterra. Recharges in minutes |
| 6w-20-1-2 | Active | 50 | Y |  | r |  | sample pump | sample pump | Dedicated submersible pump installed January 2012. Recharges completely in under 1 hour |
| GW-21-1-1 | Active | 50 | Y |  | Y |  | sample pump | footvalve | Dedicated submersible pump installed January 2012. Pump replaced with waterra. |
| GW-21-1-2 | Active | 50 | Y |  | $r$ |  | sample pump | sample pump | Dedicated submersible pump installed January 2012. |
| GW-21-2-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve |  |
| GW-25-1-1 | Active | 50 | $Y$ |  | $r$ |  | footvalve | footvalve |  |
| 6w-25-1-2 | Active | 50 | $Y$ |  | Y |  | footvalve | footvalve |  |
| GW-27-1-1 | Active | 50 | Y |  | Y |  | footvalve | footralve |  |
| 6w-27-1-2 | Inactive | 50 |  |  |  |  | footvalve | footvalve | Destroyed in summer 2022. No longer exists. |
| GW-28-1-0 | Active | 150 | Y |  | Y |  | sample pump | sample pump | Open borehole. Dedicated submersible pump installed January 2011. |
| 6w-29-1-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve |  |
| 6w-29-1-2 | Active | 50 | Y |  | $r$ |  | footvalve | footvalve |  |
| 6w-30-1-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve |  |
| GW-30-1-2 | Active | 50 | Y |  | Y |  | footvalve | footvalve | 30-1-2 almost dry in summer, bailer used occasionally. |
| GW-31-1-1 | Active | 50 |  | Y | Y |  | footvalve | footvalve |  |
| GW-31-1-2 | Active | 50 |  | r | r |  | footvalve | footralve |  |
| GW-36-11-1 | Active | 50 |  |  |  | Y | NA | NA | Pressure transducer installed June 1997 records water levels continuously. |
| 6W-36-2-1 | Active | 50 |  |  | $r$ |  | footvalve | footvalve |  |
| 6w-36-3-1 | Active | 50 | Y |  | Y |  | footvalve | footralve |  |
| 6w-37-1-1 | Active | 50 |  |  |  | Y | NA | NA | Pressure transducer installed June 1997 records water levels continuously. |
| GW-37-2-1 | Active | 50 |  |  | Y |  | footvalve | footvalve |  |
| 6w-37-3-1 | Active | 50 | Y |  | $r$ |  | footvalve | footralve |  |
| 6W-38-1-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve | Recharges very slowly, 48 hours required after purging to sample. |
| 6W-39-1-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve |  |
| 6w-39-2-1 | Active | 50 | $Y$ |  | Y |  | sample pump | sample pump | Dedicated submersible pump installed January 2011. |
| GW-40-1-1 | Active | 50 | Y |  | Y | Y | footvalve | footvalve | Pressure transducer installed in September 2008 records water levels continuously. Can be removed as required for sampling. |
| GW-41-1-1 | Active | 50 | $Y$ |  | Y | Y | footvalve | footralve | Pressure transducer installed in May 2018. |
| GW-42-1-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve |  |
| 6W-43-1-1 | Active | 50 | Y |  | Y | Y | footvalve | footralve | Pressure transducer installed in May 2018. |
| 6W-44-1-1 | Active | 50 | Y |  | Y | r | footvalve | footvalve | Pressure transducer installed in May 2018. |
| Gw-46-2-1 | Active | 50 |  |  | Y |  | NA | NA |  |
| 6w-46-3-1 | Active | 50 |  |  | Y |  | NA | NA |  |
| GW-46-4-1 | Active | 50 |  |  | Y |  | NA | NA |  |
| 6w-47-2-1 | Active | 50 |  |  | Y |  | NA | NA |  |
| GW-48-2-1 | Active | 50 |  |  | Y |  | NA | NA |  |
| 6w-50-1-1 | Active | 50 |  |  | Y |  | footralve | NA | Deactivated in from sampling program September 2012 due to poor recharge. Now only monitored for water level. |
| Gw-51-1-1 | Active | 50 |  | Y | Y |  | footvalve | footvalve |  |
| GW-51-2-1 | Active | 50 |  | Y | $r$ |  | footvalve | footvalve |  |
| 6W-51-3-1 | Active | 50 |  | Y | Y |  | footvalve | footvalve |  |

NA - Not Available or Not Applicable

Appendix A-2. Groundwater Monitoring Plan 2022-2023

| Station Name | Status | $\begin{array}{\|c\|} \hline \text { Pipe } \\ \text { Diameter } \\ (\mathrm{mm}) \end{array}$ | Monitored in 2022-2023 |  |  |  | DevelopmentMethod | Sampling Method | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Groundwater Quality Samples |  | Groudwater Level Measurements |  |  |  |  |
|  |  |  | $\underset{(4 / \mathrm{yr})}{\text { Quarterly }}$ | $\begin{gathered} \text { Bi-annual } \\ (2 \text { (2yr) } \end{gathered}$ | $\begin{array}{\|c} \substack{\text { Manual } \\ \text { measurement } \\ (4 / \mathrm{yr})} \\ \hline \end{array}$ | $\begin{array}{\|c\|} \hline \text { Pressure } \\ \text { transucuer } \\ \text { (continuous) } \end{array}$ |  |  |  |
| 6W-52-1-1 | Active | 50 | $Y$ |  | $Y$ | Y | footvalve | footvalve | Pressure transducer installed in 2018 |
| 6W-52-2-0 | Active | 150 |  |  | Y |  | NA | NA |  |
| GW-52-3-0 (P52) | Active | 150 |  |  |  | Y | NA | NA | Pressure transducer records water levels continuously. |
| GW-52-4.0.(P7) | Active | 250 | Y |  |  | Y | NA | spigot | Pressure transducer installed. Purge well in conjunction with 80-1-0-P8. Pressure transducer records water levels continuously. Sampling spigot installed in 2016 . |
| Gw-53-1-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve |  |
| 6W-54-1-1 | Active | 50 |  |  | $Y$ | Y | sample pump | sample pump | Dedicated submersible pump installed January 2012. Sampling discontinued in 2016. Water levels only. Pressure transducer installed in March 2019. |
| 6W-54-2-1 | Active | 50 |  |  | $Y$ |  | footvalve | footvalve | Water levels only. |
| Gw-54-3-1 | Active | 50 |  |  | $Y$ |  | footvalve | footralve | Water levels only. |
| 6w-55-1-1 | Active | 50 |  | Y | Y |  | footvalve | footralve |  |
| 6W-56-1-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve |  |
| 6w-57-1-1 | Active | 50 |  | Y | Y |  | footvalve | footvalve |  |
| GW-58-1-0 | Active | 150 | Y |  | Y |  | sample pump | sample pump | Dedicated submersible pump reinstalled after Phase 2 Cell one closure in October 2012. This well is an 8 " diameter borehole. |
| 6w-60-1-1 | Active | 50 | $r$ |  | $Y$ |  | footvalve | footvalve |  |
| 6w-60-2-1 | Active | 50 | Y |  | $Y$ |  | footvalve | footvalve |  |
| 6w-60-3-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve |  |
| 6W-61-1-1 | Active | 150 |  |  | Y |  | NA | NA | Borehole |
| 6W-62-1-1 | Active | 50 |  | Y | Y | Y | sample pump | sample pump | Portable pressure transducer installed January 2012 - records water levels every three hours. Dedicated submersible pump installed January 2012. |
| GW-62-2-1 | Active | 50 |  | Y | Y |  | footvalve | footvalve |  |
| 6W-63-1-1 | Active | 50 |  | Y | Y |  | sample pump | sample pump | Dedicated submersible pump installed January 2011. |
| 6W-63-2-1 | Active | 50 |  | Y | $Y$ |  | footvalve | footvalve |  |
| 6W-71-1-1 | Active | 50 | Y |  | $Y$ |  | sample pump | sample pump | Dedicated submersible pump installed January 2012 has malfunctioned. |
| 6W-71-2-1 | Active | 50 | Y |  | Y |  | sample pump | footvalve | Dedicated submersible pump installed January 2012. |
| GW-71-3-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve |  |
| 6w-72-1-1 | Active | 50 | Y |  | Y |  | sample pump | sample pump | Dedicated submersible pump installed January 2012. |
| 6W-72-2-1 | Active | 50 |  |  | Y |  | footvalve | footvalve | Well has been recharging very slowly since 2014, so it is no longer sampled. Water levels only. |
| 6w-72-3-1 | Active | 50 | Y |  | $Y$ |  | footvalve | footvalve |  |
| 6W-73-1-1 | Active | 50 | Y |  | $Y$ |  | sample pump | sample pump | Dedicated submersible pump installed January 2012. |
| 6w-73-2-1 | Active | 50 | Y |  | Y |  | sample pump | sample pump | Dedicated submersible pump installed January 2012. |
| 6W-73-3-1 | Active | 50 | Y |  | Y |  | footvalve | footvalve |  |
| 6W-74-2-1 | Active | 50 |  |  | Y |  | NA | NA |  |
| 6W-75-1-1 | Active | 50 |  |  | $Y$ |  | NA | NA |  |
| 6W-76-1-1 | Active | 50 |  |  | Y | Y | sample pump | sample pump | Dedicated submersible pump installed January 2010. Sampling discontinued in 2016. Water levels only. Pressure transducer installed in March 2019. |
| 6W-76-2-1 | Active | 50 |  |  | Y |  | sample pump | sample pump | Sampling discontinued in 2016. Water levels only. Dedicated submersible pump installed January 2012. |
| GW-76-3-1 | Active | 50 |  |  | $Y$ |  | footvalve | footvalve | Sampling discontinued in 2016. Water levels only. |
| GW-77-1-1 | Active | 50 |  | Y | $Y$ | Y | footvalve | footrave | Portable pressure transducer installed January 2012 - records water levels every three hours. Dedicated submersible pump installed. Pump failed in 2020 , replaced by waterra tubing. |
| 6W-77-2-1 | Active | 50 |  | Y | $Y$ |  | footvalve | footrave |  |
| 6W-78-1-1 | Active | 50 |  | Y | Y | Y | footvalve | footralve | Portable pressure transducer installed January 2012 - records water levels every three hours. |
| 6W-78-2-1 | Active | 50 |  | Y | Y |  | footvalve | footvalve |  |
| GW-80-1-0 (P8) | Active | NA | Y |  | $Y$ | Y | sample pump | spigot | Pressure transducer installed. Purge well in conjunction with 52-4-0-P7. Pressure transducer records water levels continuously. Sampling spigot installed in 2016. |
| GW-81-1-0.(P9) | Active | NA | Y |  | Y | r | NA | bailer | Borehole. Pressure transducer installed in 2018. |
| 6W-82-1-1 | Active | 50 |  |  | Y |  | NA | NA |  |
| GW-83-1-1 | Active | 50 |  |  | $Y$ |  | NA | NA |  |
| 6w-85-1-1 | Active | 50 | Y |  | $Y$ |  | footvalve | footrave |  |
| GW-87-1-1 | Active | 50 |  | Y | Y | Y | footvalve | sample pump | Well drilled in December 2014. Dedicated submersible pump (GeoTech) installed in February 2015. Pressure transducer installed March 2016. |
| 6w-87-2-1 | Active | 50 |  | Y | Y | Y | footvalve | waterra | Well drilled in December 2014. Pressure transducer installed in March 2016. |
| GW-88-1-1 | Active | 50 |  | Y | Y | $Y$ | footvalve | waterra | Well drilled in December 2014. Dedicated submersible pump (GeoTech) installed in February 2015 but replaced with waterra tubing after malfunctioning in 2016 . Pressure transducer installed March 2016. waterra tubing after malfunctioning in 2016. Pressure transducer installed March 2016. |
| 6w-88-2-1 | Active | 50 |  | Y | Y | Y | footvalve | waterra | Well drilled in December 2014. Pressure transducer installed in March 2016. |
| 6W-89-1-1 | Active | 50 |  |  | $Y$ | Y | NA | NA | Well drilled in Dec 2018. Pressure transducer installed in March 2019 |
| 6w-89-2-1 | Active | 50 |  |  | Y | Y | NA | NA | Well drilled in Dec 2018. Pressure transducer installed in March 2019 |
| 6W-91-1-1 | Active | 50 | Y |  | Y | Y | Waterra | Waterra | Well drilled in Dec 2019. Pressure transducer installed in March 2020 |
| 6w-92-1-1 | Active | 50 | Y |  | Y | Y | Waterra | Watera | Well drilled in Dec 2019. Pressure transducer installed in March 2020 |
| 6w-94-1-1 | Active | 50 | Y |  | $Y$ | Y | Waterra | Watera | Well drilled in Dec 2019. Pressure transducer installed in March 2020 |
| Gw-95-1-1 | Active | 50 |  | Y | Y |  | Waterra | Waterra | Well drilled April 142022 |
| 6w-96-1-1 | Active | 50 |  | Y | Y |  | Waterra | Waterra | Well drilled April 142022 |
| 6w-97-1-1 | Active | 50 |  | Y | $Y$ |  | Waterra | Waterra | Well drilled April 142022 |
| 6W-98-1-1 | Active | 50 | Y |  | Y |  | Waterra | Waterra | Well drilled April 182022 |
| 6w-99-1-1 | Active | 50 |  | Y | Y |  | Waterra | Waterra | Well drilled April 18 2022. Dry at the time of drilling. Dry during the wet season in 2022 |
| 6W-100-1-1 | Active | 50 | Y |  | Y |  | Waterra | Waterra | Well drilled April 18 2022. Dry at the time of drilling. Not enough volume to sample during the wet season in 2022 |
| GW-101-1-1 | Active | 50 |  | Y | Y |  | Waterra | Waterra | Well drilled April 182022. Dry during the wet season in 2022 |

Notes:
Footvalves are $16 \mathrm{~mm}\left(5 / 8^{\circ}\right)$ unless otherwise noted.
NA - Not Available or Not Applicable.

## Appendix A-2. Groundwater Monitoring Plan 2022-2023

| Station Name | Status | $\begin{gathered} \text { Pipe } \\ \text { Diameter } \\ (\mathrm{mm}) \end{gathered}$ | Monitored in 2022-2023 |  |  |  | $\begin{aligned} & \text { Development } \\ & \text { Method } \end{aligned}$ | Sampling Method | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Groundwater Quality Samples |  | Groudwater Leve Measurements |  |  |  |  |
|  |  |  | Quarterly $(44 / \mathrm{r})$ | $\underset{(2 \text { (2/yr) }}{\text { Bi-anual }}$ | $\begin{array}{\|c} \substack{\text { Manual } \\ \text { measurement } \\ (4 \mathrm{yr})} \\ \hline \end{array}$ | Pressure <br> transducer <br> (continuous) |  |  |  |
| GW-103-1-1 | Active | 50 |  | $Y$ | Y |  | Waterra | Waterra | Well drilled April 182022 |
| GW-104-1-1 | Active | 50 | Y |  | Y |  | Waterra | Waterra | Well drilled July 112022 |
| GW-105-1-1 | Active | 50 | Y |  | Y |  | Waterra | Watera | Well drilled July 112022 |
| GW-106-1-1 | Active | 50 | Y |  | Y |  | Waterra | Waterra | Well drilled July 122022 |
| GW-107-1-1 | Active | 50 |  | Y | Y |  | Waterra | Waterra | Well drilled July 112022 |
| 6W-108-1-1 | Active | 50 |  | Y | Y |  | Waterra | Waterra | Well drilled August 2022 |
| GW-109-1-1 | Active | 50 |  |  | Y |  | Waterra | Waterra | Well drilled August 9 2022. Deeep well. |
| GW-110-1-1 | Active | 50 |  |  | Y |  | Waterra | Waterra | Well drilled August 5 2022. Deep well. |
| P11 | Active | 101.6 | Y |  | $r$ |  | Waterra | Waterra | Well drilled August 3 2022.4" diameter borehole. Submersible pump to be installed. |
| P12 | Active | 101.6 | Y |  | Y |  | Waterra | Waterra | Well drilled August 5 2022. 4" diameter borehole |
| P1 | Active | NA | Y |  |  | Y | NA | spigot | Pressure transducer installed. QED bladder pumps installed and pressure transducer records water levels continuously. Second bladder pump installed April 2013. Electric submersible pump installed 14-May-2015. |
| P10 | Active | NA | Y |  |  | Y | NA | spigot | Pressure transducer i installed. QED bladder pumps installed and pressure transducer records water levels continuously. |
| P2 | Active | NA | Y |  |  | Y | NA | spigot | Pressure transducer installed. QED bladder pumps installed and pressure transducer records water levels continuously. |
| P3 | Active | NA | Y |  |  | Y | NA | spigot | Pressure transducer installed. QED bladder pumps installed and pressure transducer records water levels continuously. |
| P4 | Active | NA | Y |  |  | Y | NA | spigot | Pressure transducer installed. QED bladder pumps installed and pressure transducer records water levels continuously. |
| VLGW002d | Active | NA |  |  | Y |  | NA | NA |  |
| vLGwoo3d | Active | NA |  |  | Y |  | NA | NA |  |
| vLGw004D | Active | NA |  |  | $r$ |  | NA | NA |  |
| VLGw008D | Active | NA |  |  | $r$ |  | NA | NA |  |
| vLgw011s | Active | NA |  |  | Y |  | NA | NA |  |
| vLgwo15D | Active | NA |  |  | Y |  | NA | NA |  |
| vLGW016D | Active | NA |  |  | Y |  | NA | NA |  |
| vLGw017D | Active | NA |  |  | Y |  | NA | NA |  |
| VLGW018D | Active | NA |  |  | Y |  | NA | NA |  |
| vLGw019D | Active | NA |  |  | Y |  | NA | NA |  |
| VLGW021D | Active | NA |  |  | Y |  | NA | NA |  |
| vLGw026D | Active | NA |  |  | r |  | NA | NA |  |
| VLGW020D | Inactive | NA |  |  | Y |  | NA | NA | Became obstructed in 2021 |
| GW-01-1-1 | Inactive | 20 |  |  |  |  |  |  | Site destroyed. |
| Gw-01-1-2 | Inactive | ${ }^{20}$ |  |  |  |  |  |  |  |
| GW-01-1-3 | Inactive | 20 |  |  |  |  |  |  |  |
| GW-01-2-1 | Inactive | 20 |  |  |  |  |  |  |  |
| GW-01-3-1 | Inative | 20 |  |  |  |  |  |  |  |
| GW-02-1-1 | Inactive | 20 |  |  |  |  |  |  | Site destroyed |
| 6W-02-1-2 | Inactive | 20 |  |  |  |  |  |  |  |
| GW-02-1-3 | Inative | 20 |  |  |  |  |  |  |  |
| 6w-02-2-1 | Inactive | 20 |  |  |  |  |  |  |  |
| 6w-03-1-1 | Inactive | 20 |  |  |  |  |  |  | Deactivated at end of 1998, replaced by well $60-1-1$. |
| 6W-03-1-2 | Inactive | 20 |  |  |  |  |  |  | Deactivated at end of 1998, replaced by well $60-2-1$. |
| GW-03-1-3 | Inative | 20 |  |  |  |  |  |  | Deactivated at end of 1998 , replaced by well $60-3-1$. |
| 6W-03-2-1 | Inative | 50 |  |  |  |  |  |  | Destroyed and replaced with 85-1-1 in March 2009. |
| GW-04-1-1 | Inactive | 20 |  |  |  |  |  |  | Deactivated at end of 1998, replaced by new 04-2-1. |
| GW-04-1-2 | Inactive | 20 |  |  |  |  |  |  | Deactivated at end of 1998, replaced by well $04-3-1$. |
| GW-04-1-3 | Inactive | 20 |  |  |  |  |  |  | Deactivated at end of 1998, replaced by well 04-4-1. |
| 6W-04-1-4 | Inactive | 40 |  |  |  |  |  |  | Deactivated during 1998, failed to recharge after purging. |
| GW-04-2-1 | Inactive | 50 |  |  | Y |  | sample pump | sample pump | Dedicated submersible pump installed January 2012. Sampling discontinued due to very slow recharge rate. |
| GW-05-1-1 | Inactive | 20 |  |  |  |  |  |  | Site destroyed during construction of Phase 2 lagoon |
| Gw-05-1-2 | Inactive | 20 |  |  |  |  |  |  |  |
| 6w-05-1-3 | Inactive | 20 |  |  |  |  |  |  |  |
| 6w-05-2-1 | Inactive | 20 |  |  |  |  |  |  |  |
| GW-06-1-1 | Inactive | 20 |  |  |  |  |  |  | Site destroyed during dike contruction in interim filling area. |
| 6w-06-1-2 | Inactive | 20 |  |  |  |  |  |  |  |
| GW-06-1-3 | Inactive | 20 |  |  |  |  |  |  |  |
| Gw-06-2-1 | Inactive | 20 |  |  |  |  |  |  |  |
| GW-08-1-1 | Inactive | 20 |  |  |  |  |  |  | Site destroyed |
| ©w-08-1-2 | Inactive | 20 |  |  |  |  |  |  |  |
| 6W-08-1-3 | Inactive | 20 |  |  |  |  |  |  |  |
| GW-10-1-1 | Inactive | 20 |  |  |  |  |  |  | Site destroyed |

[^1]
## Appendix A-2. Groundwater Monitoring Plan 2022-2023

| Station Name | Status | $\begin{array}{\|c\|} \hline \text { Pipe } \\ \text { Diameter } \\ (\mathrm{mm}) \end{array}$ | Monitored in 2022-2023 |  |  |  | Development Method | Sampling Method | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Groundwater Quality Samples |  | Groudwater Level Measurements |  |  |  |  |
|  |  |  | $\begin{gathered} \text { Quarterly } \\ (4 / \mathrm{yr}) \end{gathered}$ | $\begin{gathered} \text { Bi-annual } \\ (2 \text { lyr) } \end{gathered}$ | $\begin{array}{\|c\|c\|c\|c\|c\|c\|c\|c\|c\|c\|} \hline \text { measurement } \\ (4 \mathrm{yr}) \\ \hline \end{array}$ | $\begin{array}{\|c\|} \hline \text { Pressure } \\ \text { transducer } \\ \text { (continuous) } \\ \hline \end{array}$ |  |  |  |
| GW-11-1-1 | Inactive | 20 |  |  |  |  |  |  | Site destroyed |
| 6W-12-1-0 | Inactive | 220 |  |  |  |  |  |  | Deactivated in 2010. |
| GW-13-1-1 | Inactive | 50 |  |  |  |  |  |  | Site destroyed |
| $6 \mathrm{~W}-13-1-2$ | Inactive | 50 |  |  |  |  |  |  |  |
| GW-14-1-1 | Inactive | 50 |  |  |  |  |  |  | Site destroyed |
| 6w-15-1-1 | Inactive | 50 |  |  |  |  |  |  | Site destroyed |
| 6W-22-1-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed during expansion of interim composting area. |
| 6w-23-1-1 | Inactive | 50 |  |  |  |  |  |  | Site destroyed in November 1994 |
| 6w-23-1-2 | Inactive | 50 |  |  |  |  |  |  |  |
| 6W-24-1-1 | Inactive | 50 |  |  |  |  |  |  | Site destroyed during construction of Phase 2 lagoon |
| 6w-24-1-2 | Inactive | 50 |  |  |  |  |  |  |  |
| Gw-26-11-1 | Inactive | 50 |  |  |  |  |  |  | Site destroyed in August 1994 |
| 6W-26-1-2 | Inactive | 50 |  |  |  |  |  |  |  |
| Gw-32-1-1 | Inactive | 50 |  |  |  |  |  |  | Site destroyed in October 1997 |
| 6w-32-1-2 | Inactive | 50 |  |  |  |  |  |  | Site destroyed in October 1997 |
| 6W-33-1-1 | Inactive | 50 |  |  |  |  |  |  | Deactivated in August 1992. |
| 6w-33-1-2 | Inactive | 50 |  |  |  |  |  |  |  |
| 6W-34-1-1 | Inactive | 50 |  |  |  |  |  |  | Site destroyed in August 1994 |
| GW-34-2-1 | Inactive | 50 |  |  |  |  |  |  |  |
| GW-35-1-1 | Inactive | 50 |  |  |  |  |  |  | Site destroyed in August 1994 |
| 6w-35-1-2 | Inactive | 50 |  |  |  |  |  |  |  |
| GW-45-1-1 | Inactive | 50 |  |  |  |  |  |  | Site destroyed. |
| Gw-46-1-1 | Inactive | 50 |  |  |  |  |  |  | Deactivated in June 1999. |
| 6w-47-1-1 | Inactive | 50 |  |  |  |  |  |  | Deactivated in December 2012. |
| 6w-48-1-1 | Inactive | 50 |  |  |  |  | NA | NA | Obstructed. Unable to measure groundwater level. |
| 6w-49-1-0 | Inactive | 50 |  |  |  |  |  |  | Deactived in June 2005. |
| GW-59-1-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed Feb 2005 by Phase 2 construction. |
| 6W-64-1-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed by interim fill area construction in 2003. |
| 6W-65-1-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed by interim fill area construction in 2003. |
| GW-66-1-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed by interim fill area construction in 2003. |
| 6W-67-1-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed. Replaced by GW-82-1-1 |
| 6w-67-2-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed. Replaced by GW-82-1-1 |
| 6w-68-1-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed. Replaced by GW-83-1-1 |
| 6W-68-2-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed. Replaced by GW-83-1-1 |
| 6w-69-1-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed July 2006 during construction on active face |
| 6w-69-2-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed July 2006 during construction on active face |
| 6W-70-1-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed by interim fill area construction in 2003. |
| 6w-74-1-1 | Inactive | 50 |  |  | Y |  | NA | NA | Decommissioned |
| 6w-79-1-1 | Inactive | 50 |  |  |  |  | sample pump | sample pump | Well decommissioned in May 2018 due to blasting in ivicinity. |
| 6w-79-2-1 | Inactive | 50 |  |  |  |  | footvalve | footvalve | Well decommissioned in May 2018 due to blasting in vicinity. |
| 6W-84-1-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed spring 2009. Recorded continuous water levels until April 2012. No longer functioning. Cannot be repaired because it's buried in garbage. |
| Gw-84-2-1 | Inactive | 50 |  |  |  |  |  |  | Destroyed spring 2009. Recorded continuous water levels until April 2012. No longer functioning. Cannot be repaired because it's buried in garbage. |
| 6w-90-1-1 | Inactive | 50 |  |  |  |  | NA | NA | Well drilled in Oct 2019. No water levels taken in well due to high landfill gas concentrations in area. |
| 6w-90-2-1 | Inactive | 50 |  |  |  |  | NA | NA | Well drilled in Dec 2018. No water levels taken in well since Oct 2019 due to high landilll gas concentrations in area. |
| 6w-93-1-1 | Inactive | 50 |  |  |  |  | Waterra | Waterra | Destroyed September 2022 due to aggregate stockpiling |
| T-86-1-1 | Inactive | NA |  |  |  |  |  |  |  |
| T-86-1-2 | Inactive | NA |  |  | Y | Y | NA | NA | Pressure transducer installed October 2010. Well replaces $\mathrm{GW}-84-1-1$ \& $\mathrm{GW}-84-2-1$. Pressure transducer records water levels continuously. No longer functioning. |
| T-86-2-1 | Inactive | NA |  |  | Y | Y | NA | NA | Pressure transducer installed October 2010. Well replaces GW-84-1-1 \& GW-84-2-1. Pressure transducer records water Ievels continuously. No longer functioning. |
| T-86-2-2 | Inactive | NA |  |  | Y | Y | NA | NA | Pressurre transducer installed October 2010. Well replaces GW-84-1-1 \& GW-84-2-1. Pressure transducer records water tevels continuously No Nonger functioning levels continuously. No longer functioning. |
| VLGw001D | Inactive | NA |  |  |  |  | NA | NA | Well is obstructed. No water level measurements were taken in 2018-2019 |

Footvalves are $16 \mathrm{~mm}\left(5 / 8^{\circ}\right.$ unless
NA- Not Available or Not Apolicab

## A3. Groundwater Elevations

Appendix A-3. Groundwater Elevations 2022-2023
Groundwater Level Elevations (mASL)

| Groundwater Level Elevations (mASL) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Q2 | Q3 | Q4 | Q1 |
| Station | $\begin{gathered} \hline \text { May } 9,2022 \\ \text { to } \\ \text { May 10, } 2022 \\ \hline \end{gathered}$ |  |  | February 2, 2023 <br> to <br> February 4, 2023 |
| 04-2-1 | 127.01 | 125.49 | 125.99 | 127.28 |
| 04-3-1 | 124.38 | 123.07 | 124.54 | 124.62 |
| 04-4-1 | 125.29 | 123.93 | 124.97 | 125.58 |
| 07-1-0 | 141.64 | 141.18 | 141.48 | 141.64 |
| 09-1-0 | 146.40 | 144.20 | --- | 146.87 |
| 16-1-1 | 129.99 | 123.93 | 129.25 | 130.95 |
| 16-1-2 | 131.19 | 123.95 | 129.30 | 131.05 |
| 16-2-1 | 131.59 | 126.10 | 130.55 | 132.15 |
| 16-2-2 | 131.59 | 126.01 | 130.46 | 132.01 |
| 17-1-1 | 144.95 | 142.79 | 143.98 | 147.30 |
| 17-1-2 | 144.94 | 142.80 | 143.97 | 147.29 |
| 17-1-3 | 145.33 | 143.57 | 144.21 | 147.54 |
| 18-1-1 | 151.32 | 149.16 | 150.08 | 151.57 |
| 18-1-2 | 151.15 | 148.99 | 149.92 | 151.46 |
| 18-2-1 | 161.32 | 158.74 | 161.18 | 163.47 |
| 18-2-2 | 161.29 | 158.89 | 161.21 | 163.51 |
| 19-1-1 | 133.54 | 133.27 | 133.44 | 133.54 |
| 19-1-2 | 133.66 | 132.35 | 133.41 | 133.66 |
| 19-2-1 | 133.19 | 132.02 | 133.19 | 133.19 |
| 19-2-2 | 132.68 | 129.94 | 132.61 | 132.52 |
| 20-1-1 | 109.85 | 109.02 | 109.93 | 110.13 |
| 20-1-2 | 109.61 | 108.88 | 109.63 | 109.77 |
| 21-1-1 | 110.28 | 109.54 | 109.85 | 110.52 |
| 21-1-2 | 109.88 | 109.06 | 110.33 | 110.04 |
| 21-2-1 | 109.88 | 109.07 | 109.83 | 110.05 |
| 25-1-1 | 125.31 | 123.98 | 124.87 | 126.06 |
| 25-1-2 | 126.52 | 125.47 | 126.41 | 126.72 |
| 27-1-1 | 137.38 | 136.94 | Not accessible | 140.54 |
| 27-1-2 | 137.43 | 137.40 | Destroyed | Destroyed |
| 28-1-0 | 119.36 | 118.90 | 119.27 | 119.98 |
| 29-1-1 | 112.12 | 111.19 | 112.06 | 112.36 |
| 29-1-2 | 112.22 | 111.29 | 112.17 | 112.40 |
| 30-1-1 | 106.06 | 104.81 | 105.79 | 109.39 |
| 30-1-2 | 107.11 | 105.86 | 106.89 | 109.21 |
| 31-1-1 | 104.66 | 103.61 | 104.42 | 102.22 |
| 31-1-2 | 104.44 | 103.66 | 102.41 | 103.40 |
| 36-1-1 | 123.91 | 124.91 | 125.91 | 125.91 |
| 36-2-1 | 123.02 | 121.45 | 122.74 | 123.38 |
| 36-3-1 | 123.51 | 121.21 | 123.50 | 124.40 |
| 37-1-1 | 125.24 | 126.24 | 127.24 | 127.24 |
| 37-2-1 | 123.08 | 120.91 | 122.61 | 122.76 |
| 37-3-1 | 124.83 | 122.92 | 124.50 | 125.47 |
| 38-1-1 | 119.90 | 119.47 | 119.84 | 120.39 |
| 39-1-1 | 120.72 | 119.39 | 120.50 | 122.15 |
| 39-2-1 | 119.90 | 119.46 | 119.82 | 120.51 |
| 40-1-1 | 117.30 | 118.30 | 119.30 | 119.30 |
| 41-1-1 | 147.37 | 147.00 | 147.28 | 147.55 |
| 42-1-1 | 137.69 | 137.08 | 137.60 | 134.82 |
| 43-1-1 | 158.22 | 157.05 | 158.11 | 158.77 |
| 44-1-1 | 160.21 | 158.94 | 160.10 | 160.43 |
| 46-2-1 | 156.04 | 155.92 | 156.02 | 156.02 |
| 46-3-1 | 151.82 | 151.92 | 150.50 | 152.30 |
| 46-4-1 | 149.27 | 149.36 | 149.36 | 149.41 |
| 47-2-1 | 152.16 | 151.98 | 151.93 | 151.93 |
| 48-2-1 | 149.63 | 148.51 | 148.20 | 149.27 |
| 50-1-1 | 117.44 | 113.45 | 114.89 | 118.04 |
| 51-1-1 | 109.96 | 108.79 | 110.04 | 110.17 |
| 51-2-1 | 110.17 | 108.85 | 110.23 | 110.41 |
| 51-3-1 | 110.26 | 108.87 | 110.27 | 110.51 |
| 52-1-1 | 117.74 | 117.23 | 117.30 | 117.49 |
| 52-2-0 | 118.11 | 117.75 | 118.09 | 118.13 |
| 52-3-0 (P52) | 116.63 | 116.63 | 116.63 | 116.63 |
| 52-4-0 (P7) | 112.94 | 112.94 | 112.94 | 112.94 |
| 53-1-1 | 120.85 | 120.23 | 120.87 | 121.65 |
| 54-1-1 | 149.74 | 147.48 | 148.00 | 150.74 |
| 54-2-1 | 149.32 | 147.10 | 147.60 | 150.29 |
| 54-3-1 | 149.37 | 147.50 | 147.91 | 150.41 |
| 55-1-1 | 142.68 | 140.66 | 141.06 | 143.34 |

Notes:
Bracketed data on bottom of well.
--- - Not measured

Appendix A-3 - Groundwater Elevations 2022-2023

| Groundwater Level Elevations (mASL) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Q2 | Q3 | Q4 | Q1 |
| Station | $\begin{gathered} \text { May } 9,2022 \\ \text { to } \\ \text { May } 10,2022 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { September 6, } 2022 \\ \text { to } \\ \text { September 7, } 2022 \end{gathered}$ | $\begin{gathered} \text { November } 14,2022 \\ \text { to } \\ \text { November } 17,2022 \end{gathered}$ | February 2, 2023 <br> to <br> February 4, 2023 |
| 56-1-1 | 144.28 | 142.27 | 144.40 | 145.87 |
| 57-1-1 | 129.61 | 125.30 | 128.84 | 129.96 |
| 58-1-0 | 128.09 | 126.35 | 127.91 | 128.46 |
| 60-1-1 | 141.70 | 140.97 | 141.56 | 141.79 |
| 60-2-1 | 141.72 | 141.00 | 141.59 | 141.79 |
| 60-3-1 | 141.27 | 140.50 | 141.15 | 141.47 |
| 61-1-1 | 184.79 | 183.13 | 184.10 | 186.47 |
| 62-1-1 | 172.10 | 165.54 | 166.49 | 174.02 |
| 62-2-1 | 172.09 | 166.52 | 165.10 | 174.03 |
| 63-1-1 | 191.53 | 188.85 | 190.75 | 192.35 |
| 63-2-1 | 193.14 | 190.56 | 191.11 | 193.95 |
| 71-1-1 | 141.67 | 140.69 | 141.23 | 141.95 |
| 71-2-1 | 141.82 | 139.91 | 141.46 | 142.08 |
| 71-3-1 | 141.84 | 140.93 | 141.47 | 142.08 |
| 72-1-1 | 141.57 | 140.83 | 141.40 | 141.70 |
| 72-2-1 | 141.81 | 141.15 | 141.35 | 141.77 |
| 72-3-1 | 141.76 | 141.02 | 141.39 | 141.89 |
| 73-1-1 | 133.63 | 131.44 | 133.10 | 134.41 |
| 73-2-1 | 133.67 | 131.45 | 133.15 | 134.40 |
| 73-3-1 | 132.83 | 130.76 | 132.57 | 133.22 |
| 74-2-1 | 152.24 | 152.17 | 152.03 | 152.05 |
| 75-1-1 | 129.79 | 128.50 | 128.13 | 128.67 |
| 76-1-1 | 158.20 | 155.71 | 155.73 | 158.42 |
| 76-2-1 | 161.00 | 158.83 | 159.33 | 162.07 |
| 76-3-1 | 163.03 | 161.09 | 160.17 | 164.55 |
| 77-1-1 | 153.77 | 150.90 | 152.66 | 152.55 |
| 77-2-1 | 152.93 | 149.94 | 151.98 | 152.05 |
| 78-1-1 | 131.54 | 127.98 | 143.40 | 133.33 |
| 78-2-1 | 134.88 | 131.75 | 143.39 | 137.35 |
| 80-1-0 (P8) | 113.01 | 114.01 | 115.01 | 115.01 |
| 81-1-0 (P9) | 116.32 | 115.87 | 116.24 | 116.66 |
| 82-1-1 | 132.84 | DRY | DRY | 132.87 |
| 83-1-1 | 127.52 | 127.31 | 127.30 | 127.65 |
| 85-1-1 | 145.30 | 144.96 | 145.33 | 145.98 |
| 87-1-1 | 175.91 | 161.99 | 159.49 | 176.30 |
| 87-2-1 | 176.14 | 162.58 | 162.56 | 177.13 |
| 88-1-1 | 170.94 | 164.10 | 165.10 | 171.54 |
| 88-2-1 | 171.34 | 165.43 | 165.43 | 172.02 |
| 89-1-1 | 152.47 | --- | --- | 152.67 |
| 89-2-1 | 149.76 | --- | --- | 150.33 |
| 90-1-1 | --- | --- | --- | --- |
| 90-2-1 | --- | --- | --- | --- |
| 91-1-1 | 160.04 | 158.35 | 160.26 | 160.81 |
| 92-1-1 | 161.36 | 158.53 | 160.33 | 161.42 |
| 93-1-1 | Obstructed | Destroyed | Destroyed | Destroyed |
| 94-1-1 | 199.13 | 183.57 | 181.58 | 199.99 |
| LG-02-D | 149.08 | 148.13 | 148.10 | 149.00 |
| LG-03-D | 149.96 | 149.81 | 149.81 | 149.86 |
| LG-04-D | 157.43 | 157.26 | 157.04 | 157.07 |
| LG-08-D | 147.49 | 143.87 | 143.32 | 146.56 |
| LG-11-S | 155.56 | 155.55 | 155.55 | 155.56 |
| LG-15-D | 149.03 | 148.25 | 148.25 | 148.99 |
| LG-16-D | 150.45 | 149.49 | 149.34 | 149.69 |
| LG-17-D | 151.64 | 150.06 | 149.41 | 149.99 |
| LG-18-D | 150.83 | 150.60 | 150.14 | 150.02 |
| LG-19-D | 150.44 | 150.46 | 150.22 | 150.16 |
| LG-21-D | 142.73 | 142.65 | 142.16 | 142.32 |
| LG-26-D | 144.73 | 144.31 | 144.25 | 144.18 |
| 95-1-1 | --- | 125.68 | 126.41 | 127.57 |
| 96-1-1 | --- | 125.76 | 126.78 | 127.93 |
| 97-1-1 | --- | 125.40 | 126.28 | 127.60 |
| 98-1-1 | --- | 125.67 | 126.64 | 127.66 |
| 99-1-1 | --- | Dry | Dry | 131.46 |
| 100-1-1 | --- | Dry | 128.64 | 129.14 |
| 101-1-1 | --- | 127.57 | 128.13 | 128.65 |
| 103-1-1 | --- | 132.40 | 132.22 | 132.86 |
| 104-1-1 | --- | 130.40 | 130.71 | 131.85 |
| 105-1-1 | --- | 128.21 | 128.59 | --- |
| 106-1-1 | --- | 128.01 | 128.32 | 129.05 |
| 107-1-1 | --- | 123.79 | 124.38 | 124.98 |
| 107-1-2 | --- | Dry | Dry | Dry |
| 108-1-1 | --- | 122.25 | 123.90 | 124.30 |
| 109-1-1 | --- | 128.29 | 129.60 | 128.80 |
| 110-1-1 | --- | 132.05 | 132.90 | 133.92 |
| P-11 | --- | 125.32 | 127.83 | 128.14 |
| P-12 | --- | 120.33 | 122.77 | 123.60 |

Notes:
Bracketed data on bottom of well.
--- - Not measured

## A4. Surface Water Station Details

| Station Name | Location |  | Status | Parameter List |  | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Northing | Easting |  |  |  |  |
|  | UTM (NAD83) | UTM (NAD83) |  | Routine |  |  |
|  |  |  |  | (4/yr) | (2/yr) |  |
| Sw-N-05 | 5376534.511 | 465729.384 | Active | Y |  | Heal Creek - 40m from perimeter fence line |
| Sw-N-14 | 5376795.252 | 466228.944 | Active |  | Y | Heal Creek 2/3 of the way from the Northeast Diversion Ditch to Durrance Creek. |
| Sw-N-16 | 5376506.808 | 465968.695 | Active | Y |  | North Wetland at dischage of Weir SF5 into North Wetland Creek, just above confluence with Heal Creek. |
| Sw-N-17 | 5376566.303 | 465903.628 | Active | Y |  | Heal Creek below confluence with 42 Creek and above confluence with North Wetland Creek. |
| Sw-N-18 | 5376429.846 | 465679.709 | Active | Y |  | Manhole at discharge to Northwest Sedimentation Pond. |
| Sw-N-19 | 5376423.322 | 466040.279 | Active | Y |  | Northeast Diversion Ditch below Northeast Sedimentation Pond, just above discharge into North Wetland. |
| Sw-N-41s1 | 5376892.202 | 465171.078 | Active | Y |  | 41 Creek at north side of Willis Point Road, near source. |
| Sw-N-41s3 | 5377102.438 | 465021.226 | Active | Y |  | 41 Creek just above discharge to Durrance Lake. |
| Sw-N-42s1 | 5376753.436 | 465553.804 | Active | Y |  | 42 Creek at discharge from 42 Wetland below southeast end of Yardwaste Pad. Across Willis Point Road from Well 42. |
| Sw-N-45 | 5376605.647 | 465797.403 | Active | Y |  | Heal Creek just above confluence with 42 Creek. |
| Sw-N-50 | 5376353.361 | 465438.872 | Active | Y |  | Toutle Valley break out. |
| Sw-N-51 | 5376323.976 | 465444.733 | Active | Y |  | NW diversion ditch just above confluence with Toutle Valley break out. |
| Sw-N-53 | 5376456.000 | 465693.000 | Active | Y |  | Drainage from High Level Road North Diversion Ditch to Northwest Sedimentation Pond. |
| Sw-N-54 | 5376438.000 | 465713.000 | Active | Y |  | Runoff from northeast face of Phase 2 and Phase 2 Cell 1 closure into Northwest Sedimentation Pond. Replaced Sw-N-47. |
| Sw-N-CSs2 | 5376933.072 | 464896.583 | Active |  | Y | Control station on south side of Willis Point Rd at ephemeral stream and culvert 300 m west of Yardwaste gate. |
| Sw-S-03 | 5375637.813 | 466077.533 | Active | Y |  | Kilarney Creek at culvert discharging from underneath Recycle Road. |
| Sw-S-04 | 5375447.329 | 466171.246 | Active | Y |  | Kilarney Creek below confluence with Southwest Diversion Ditch. |
| Sw-S-12 | 5375661.074 | 465954.884 | Active | Y |  | At the discharge of Weir SF2, upstream of Killarney Creek. |
| Sw-S-20 | 5375607.030 | 465945.921 | Active |  | Y | At the discharge of Weir SF3, in the Southwest Diversion Ditch where it converges with the South High Level Road. |
| Sw-S-21 | 5375419.559 | 466150.441 | Active | Y |  | Southwest Diversion Ditch just above confluence with Kilarney Creek. |
| Sw-S-24 | 5375553.049 | 466179.487 | Active | Y |  | Kilarney Creek just above confluence with Bike Trail Kiosk Creek and below confluence with Southeast Storm Drain. |
| Sw-S-27 | 5375591.468 | 466163.597 | Active | Y |  | Southeast Storm Drain just above confluence with Kilarney Creek. |
| Sw-S-52 | 5376059.905 | 465472.066 | Active |  | Y | Creek from Mt. Work before entering culvert draining to South Diversion Ditch. |
| Sw-N-57 | 5376439.917 | 465744.608 | Active | Y |  | Added December 5 2022. Green pipe outlet at West end of upper lagoon. Collects directed surface water runoff from Shane's pond and the Toutle Valley. |
| Sw-N-58 | 5376430.529 | 465691.127 | Active | Y |  | Added December 5 2022. Manhole near P11, collecting from black pipe in ditch along High Level Road and groundwater pumped out of P11. |
| Sw-N-59 | 5376468.163 | 465532.662 | Active | Y |  | Added December 5 2022. Manhole near GW 77, collecting surfacewater from the kitcen scraps pad (not yet bult as of May 25 2023). |
| Sw-N-60 | 5376504.232 | 465439.352 | Active | Y |  | Added December 5 2022. Manhole, smaller channel within the manhole collecting directed surfacewater runoff from Shane's pond. |
| Sw-N-61 | 5376504.346 | 465135.424 | Active | Y |  | Added December 5 2022. Manhole on the West side of the road collected surfacewater drainage from Shane's pond. On the East side of the road there is a pipe at the base of a block wall which feeds into this manhole. It is the beginning of the collection point which reports to SW-N-57. |
| Sw-N-62 | 5376346.869 | 465503.734 | Active | Y |  | Added December 5 2022. Culvert at the outlet of ditch in Toutle valley. |
| Sw-N-63 | 5377369.181 | 466786.923 | Active | Y |  | Added February 8 2023. Sampling point on Durrance creek after the confluence of Durrance Creek and Heal creek. Located adjacent to a culvert that runs under the road on private property. |
| Sw-N-64 | 5377367.064 | 466825.288 | Active | Y |  | Added February 8 2023. Sampling point on Todd Creek upstream of the Durrance Creek input. |
| Sw-N-65 | 5377383.733 | 466830.580 | Active | Y |  | Added February 8 2023. Sampling point on Todd Creek downstream of the Durrance Creek input. |
| Sw-N-15 | 5377109.531 | 465982.354 | Active | n/a |  | Durrance Creek well above confluence with Heal Creek. |
| Sw-N-41s4 | 5377189.115 | 465130.141 | Active | Y |  | Discharge from Durrance Lake. |

## Appendix B

## Water Quality Data

- B1. Groundwater Quality
- B2. Domestic Well Quality
- B3. Surface Water Quality
- B4. Monthly Leachate Quality Data Hartland Valve Chamber
- B5. Quarterly Leachate Quality Trace Organics
- B6. Monthly Leachate Quality Phase 2 Cleanout
- B7. Monthly Leachate Quality North Purge Wells
- B8. Monthly Leachate Quality Controlled Waste Drainage
- B9. Monthly Leachate Quality South Purge Well
- B10. Monthly Leachate Quality West Face Drainage
- B11. Monthly Leachate Quality Cell 3 Pipe Outlet
- B12. Monthly Leachate Quality Emerging Contaminant

B1. Groundwater Quality



| вc csR |  |  |  | ${ }_{\text {atem }}^{\text {atw waximum（1）}}$ |  |  |  |  |  | 950 | ${ }^{9}$ | 50 10 | （1000 | $\stackrel{2(8)}{8}$ |  | ${ }_{\substack{12000 \\ 5000}}^{\text {col }}$ | ${ }^{\frac{0.54}{5}}$ |  | ${ }_{200}^{1500}$ | ${ }_{\substack{90 \\ 600}}$ | ${ }^{\text {400 }}$ |  | ${ }^{(6)}$ | ${ }^{40-180059} \mid$ | ${ }^{33(9)}$ |  | （6） |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sation | Sample Type | Compliance | Date Sampled | Parameter | PH FFieded | $\begin{gathered} \text { Specific } \\ \text { Conductivity - } \\ 25^{\circ} \mathrm{C} \text { (Field) } \end{gathered}$ | $\underbrace{\text { empeatura }}$ | Akainity | ${ }_{\substack{\text { Harcessas } \\ \text { cacos }}}$ | Auminum | Antinomy | Aseanc | Barium | Benlum | Bismut | ${ }_{\text {boon }}$ | Casmium | atum | Chloride | m | cobar | Copoer | ron | Lead | untum | Manesum | Manganse |
|  |  |  |  | Fraction | тот | тот | тот | тот | ois | ${ }^{1 / 5}$ | ${ }^{1 / 5}$ | ${ }_{\text {on }}$ | ${ }^{\text {onesm }}$ | ois | Dis | Dis | Dis | Ois | ois | 015 | ois | dis | 0is | （tis | ois | ois | Dis |
|  |  |  |  | $\xrightarrow[\text { Metros doiection Limit }]{\text { Unit }}$ | ${ }_{0}^{\text {pH }}$ | ${ }_{\text {ustm }}$ | $\stackrel{.}{0.1}$ | mal | ${ }_{\text {mgh }}^{0.5}$ | ${ }_{\substack{4.51 \\ 0.5}}$ |  |  | ${ }_{\text {a }}^{\text {4．02 }}$ |  | ${ }_{\substack{4.005 \\ 0.005}}$ | ${ }_{\text {ce9 }}^{10}$ |  | ${ }_{\text {men }}^{0.05}$ | m9L | ${ }_{\text {men }}^{0.1}$ | ${ }_{\text {man }}^{\text {ma }}$ | ${ }_{\text {mal }}^{\text {m0，}}$ | ${ }_{\text {men }}^{1}$ |  | ${ }_{\substack{\text { u9，} \\ 0.5}}$ | m91 | ${ }_{\text {man }}^{\text {man }}$ |
| $\frac{5}{51-2 \cdot 1}$ | ss |  |  | Eearacto olueres | ${ }_{\substack{7.9 \\ 6.91}}$ | ${ }^{\text {303 }}$ | 10.6 <br> 115 <br> 1 | ${ }^{\frac{140}{140}}$ | ${ }^{174}$ | ${ }^{236}$ |  | （0．03 | ${ }^{30}$ | ${ }_{0}^{0.01}$ | 0.005 | ${ }^{170}$ | ${ }^{0.0066}$ | ${ }_{4}^{47.9}$ | ${ }^{19}$ | O． | － |  | 1．7 | ${ }_{\text {a }}^{0.0012}$ | ${ }^{1.11}$ | ${ }^{13.1}$ |  |
|  | css |  |  |  | $\xrightarrow{6.89}$ |  | ${ }^{11.5}$ | ${ }^{140}{ }_{140}$ | ${ }_{\substack{185 \\ 185}}^{1}$ | －${ }^{188} 8$ | ${ }_{\substack{0.123 \\ 0.234}}^{0}$ |  | ${ }_{3}^{30.9} 3$ | ${ }_{0}^{0.01}$ | ${ }_{\substack{0.005 \\ \hline .005}}^{\substack{0.055}}$ |  |  |  | ${ }^{28}$ | ${ }^{0.1}$ | ${ }_{\substack{0.0635 \\ 0.15}}^{\substack{0.15}}$ | － |  | （0．0．73 | $\stackrel{1}{1.12}$ | ${ }_{\substack{11 \\ 127}}^{12}$ |  |
|  |  |  |  |  |  | ${ }_{420}^{265}$ | 10．5 10.6 ${ }^{11.6}$ | ${ }^{120}{ }_{110}^{10}$ | ${ }^{168}$ |  | ${ }^{0.136} 0$ | $\underbrace{}_{\substack{0.441 \\ 0.611}}$ | ${ }^{197}$ | ${ }^{\text {O．01 }}$ | （0．005 | ${ }_{\text {che }}^{2225}$ | ${ }_{\substack{0.005 \\ 0.005}}^{\substack{\text { O．}}}$ | （ ${ }_{\text {454，}}^{407}$ | ${ }^{36}$ | ${ }^{0.1}$ | －${ }^{0.0348}$ |  | $\stackrel{19}{155}$ | ${ }^{0.0043}{ }^{0.0003}$ | ${ }_{\text {¢ }}$ | ${ }_{\text {¢ }}^{13,2}$ | （0．0．38 |
| ${ }^{\text {a }}$ | ciss |  | ${ }^{\text {coser }}$ |  |  | ${ }^{\text {cien }}$ | ${ }_{8}^{8.6}$ | ${ }^{130}$ | ${ }_{\text {cti }}^{198}$ | $\stackrel{\substack{276 \\ 220}}{ }$ | ${ }_{0}^{0.139}$ | $\xrightarrow{\text { O．} 0.063}$ | ${ }_{18,3}{ }^{\text {94，}}$ | 0.01 | 0.005 | ${ }_{20}^{20 r^{20}}$ | 0.0063 | ${ }_{\substack{\text { 4．7 } \\ 54.9 \\ \hline \\ \hline}}$ | ${ }^{36}$ | ${ }_{0}^{0.3}$ | ${ }_{0}^{0.084}$ | ${ }_{\substack{1.4 \\ \hline 1.4}}^{\text {0，}}$ | $\stackrel{150}{50}$ | ${ }_{\text {O．0．07 }}^{0.0}$ |  | ${ }_{198}^{13,8}$ |  |
| ${ }_{5}^{52-1}$ | ss ${ }_{\text {ss }}^{\text {ss }}$ |  |  |  | ${ }_{6.98}^{6.98}$ |  |  | ${ }^{880} 8$ | ${ }_{545}^{459}$ | ${ }^{3}{ }^{3} 9$ | $\stackrel{0.14}{0.04}$ | 0.19 | － | ${ }_{0}^{0.05}$ | ${ }_{0}^{0.025} 0$ |  | ${ }^{0.025}$ |  | ${ }^{\frac{1}{180}} 1$ | （1．04 | ${ }^{2083}$ | ${ }^{0.25}$ | ${ }_{\substack{621 \\ 6621}}^{56}$ | －0．025 | ${ }^{5.1}$ | － 40.5 |  |
| ${ }_{\text {k }}^{5}$ | ss <br> ss <br> ss |  |  | Cleareand columes | ${ }_{7}^{725}$ | （isen | ${ }^{124}$ |  | ${ }_{\substack{584 \\ 484}}$ | ${ }_{5}^{5}$ | ${ }_{0}^{0.17}$ | ${ }^{0.212} 0$ | － 116. | ${ }_{0}^{0.05}$ | ${ }_{\substack{0.025 \\ 0.01}}^{\text {col }}$ | －3340 <br> 3 3130 | ${ }_{\substack{0.025 \\ 0.01}}$ | $\underset{\substack{138 \\ 127 \\ 127}}{ }$ | ${ }_{\substack{180 \\ 180}}^{170 .}$ | ＋1．13 | 221 <br> 209 | －0．25 |  | ${ }_{0}^{0.028}$ | 5.9 5.1 5 | ${ }_{40.2}^{40.5}$ | $\underbrace{\substack{\text { 373 }}}_{\substack{368 \\ 373}}$ |
| 隹 |  |  |  | Coarans siolitus |  |  | 18 <br> 18 <br> 18 <br> 18 |  |  | $\stackrel{29}{9,9}$ |  |  |  |  | $\frac{0.01}{0.01}$ |  | ${ }^{0.01}$ |  |  |  | ${ }_{\substack{209 \\ 535}}$ | $\stackrel{0.5}{0.5}$ | ¢ | $\xrightarrow{0.019}$ | ${ }^{\frac{51}{22}}$ | －${ }_{\text {40，5 }}^{522}$ |  |
| ${ }^{\text {che }}$ | ss |  | （1052022 | Moderatey wubde moderatey bown | （1739 | （inco |  | ${ }^{14000}$ | ${ }_{\substack{665 \\ 542}}$ | ${ }^{8}{ }^{8.3}$ | ${ }_{0}^{0.194}$ | ${ }_{\substack{6.92 \\ 7,37}}$ | ${ }_{\text {a }}^{398}$ | ${ }_{0}^{0.02}$ | ${ }^{0.00}$ |  | 0.01 <br> 0.01 | ${ }_{\text {¢ }}^{1254}$ |  | ${ }_{\substack{2.55 \\ 5.39}}$ | ${ }_{6}^{5.72}$ | ${ }_{0}^{0.67} 0$ |  | ${ }_{0}^{0.0046} 0$ | 2.2 <br> 1.6 | ${ }_{6}^{61,4}$ |  |
| ${ }^{524.40 .97)}$ | ss |  | ${ }^{22442023}$ | Coarands colouriss |  | － | ${ }^{13,7}$ | ${ }^{15000}$ | ${ }_{5}^{513}$ | ${ }^{11.5}$ | 0.127 | 5.78 | ${ }^{499}$ | 0.02 | 0.01 | （ | 0.01 | ${ }_{\text {127 }}^{127}$ | － |  | 5， <br> $\substack{\text { 5，} \\ 0.5 \\ \hline}$ | － |  | O．0．7 |  | ${ }_{6}^{62}$ | $\stackrel{1}{1200}$ |
|  | ¢ ${ }_{\text {ss }}^{\text {ss }}$ | $r$ |  | Claerand coouress | （739 <br> 8.02 | （ | ${ }^{10.8}{ }^{11.5}$ | ${ }^{190}$ | ${ }^{241}$ | ${ }^{4.07}{ }_{0}^{4.86}$ | ${ }_{\text {O}}^{0.02}$ | －${ }_{\text {O．1．34 }}^{0.126}$ | ${ }^{19.9}{ }^{19.9}$ |  | ${ }_{\substack{0.005 \\ 0.005}}^{\substack{\text { 0．0 }}}$ | ${ }_{4}^{426}$ | ${ }_{\substack{0.0014 \\ 0.0133}}^{0 .}$ |  | ${ }^{7}$ | ${ }^{0.1}$ |  | ${ }_{0}^{0.005}$ | ${ }_{\substack{232 \\ 313 .}}^{20}$ | $\xrightarrow{0.005}$ | ${ }^{0.5}$ | $\underbrace{}_{\substack{7.65 \\ 6.87}}$ | ${ }^{\frac{204}{169}}$ |
| 1－1 | ss | $\stackrel{r}{r}$ | ${ }^{12262023}$ | Ciearand oloumess | ${ }^{7} 744$ |  | 10 |  | ${ }^{248}$ | ${ }_{\text {l }}^{1.85}$ | ${ }^{0.02}$ | ${ }^{0.086}$ | ${ }^{16.2}$ | ${ }^{0.01}$ | ${ }_{0}^{0.005}$ | ${ }_{5}^{518}$ | 0.0092 | ${ }^{874}$ | ${ }_{6}^{64}$ | 0.1 | 0.25 | ${ }^{0.138}$ | ${ }_{381}$ | ${ }_{0}^{0.0067}$ | 0.5 | ${ }^{7,26}$ | ${ }^{163 .}$ |
| $\frac{3}{55 \cdot 1-1}$ | \％ | r |  | Claarand coluress |  | ${ }^{490}$ |  |  |  |  | ${ }^{0.022}$ |  |  |  | ${ }^{0.0005}$ |  | ${ }^{\text {O．0129 }}$ |  |  |  |  |  |  | ${ }_{\text {O．0．74 }}^{0.005}$ |  | ¢， |  |
| 8，－1／ | ss | $r$ | 121742022 | Claar and colouress | ${ }^{724}$ | ${ }^{297}$ | ${ }^{8.6}$ | ${ }_{10} 10$ | ${ }^{204}$ |  | 0.105 | ${ }^{0.47}$ |  |  | ${ }_{0}^{0.005}$ |  |  |  |  | 0.15 | 0.0388 |  | 10.9 | 0.0084 |  | ${ }^{6} 27$ | ${ }_{6}^{628}$ |
|  | ss | $\stackrel{r}{r}$ | ${ }_{\text {Slil2022 }}$ | cole | ${ }^{7.53}$ | ${ }_{\substack{294 \\ 308}}$ | ${ }^{9.5}$ | ${ }^{150}{ }^{150}$ | ${ }^{201}$ | ${ }^{1.51}$ | ${ }_{0}^{0.037}$ | 0.16 | ${ }^{38,3}$ | 0.01 | 0.005 | ${ }^{18 .}$ | 0.0119 | ${ }^{70.5}$ | ${ }^{9.1}$ | ${ }_{0}^{0.1}$ | ${ }^{0.0067}$ | ${ }^{0.148}$ | ${ }_{5} 5.1$ | － 0.005 | 0.5 | 6.05 |  |
| E－1．1 | ${ }_{\text {ss }}^{\text {ss }}$ | $r$ | ${ }^{\text {a }}$ |  | ${ }_{7}^{7,74}$ |  | ${ }_{9}^{9.4}$ | ${ }_{150}^{150}$ | ${ }^{120}$ | ${ }_{5.97}^{59}$ | ${ }_{0}^{0.085}$ | ${ }_{\text {coin }}^{0.149}$ |  | 0.01 | ${ }_{0}^{0.005}$ | ${ }^{20} 70$. | ${ }_{0}^{0.0282}$ | ${ }^{20.1}$ | ${ }_{6,7}^{6.7}$ |  | ${ }_{\substack{0 \\ 0.0006 \\ 0.0161}}$ | ${ }_{0}^{0.154}$ | ${ }_{8.9}$ | ${ }_{0}^{0.0062}$ | 0．5 | ${ }_{6.26}^{4 .}$ | ${ }_{\text {O．}}^{0.762}$ |
|  |  | $\stackrel{r}{r}$ |  | Iearard coloules |  |  |  |  |  | （117 | ${ }_{0}^{0.04}$ | 0.152 | 22.9 | 0.01 | 0.005 | ${ }^{30}$ | 0.0243 | 66.9 | ${ }_{8}^{82}$ | 0.1 | ${ }^{0.0086}$ | 0.118 | ${ }^{11,3}$ | 0.0052 | 0.5 | ${ }_{5}^{549}$ | 0．223 |
|  | ss | $r$ | 1214142022 | clear and coluriess | ${ }_{7}^{7,56}$ | ${ }^{347}$ | ${ }_{8.8}$ | 170 | ${ }_{\text {237 }}$ | ${ }_{5} 5$ | 0.055 | 0.039 | ${ }^{1.2}$ | 0.015 | 0.005 | ${ }_{\text {83，}}^{\text {83，}}$ | 0.0143 | ${ }^{854}$ | ${ }^{14}$ | 0.1 | 0.085 | ${ }_{0} 0.221$ | 18.4 | 0.0133 | 0.5 | ${ }_{5.68}$ | ${ }^{122}$ |
| ${ }^{58,10}$ | ${ }_{\text {crem }}^{\text {spm }}$ |  |  | Noen of tuplicateses | ${ }_{6}^{6.965}$ |  | ${ }^{20} 193$ | ${ }^{1800}$ | ${ }_{\substack{4040 \\ 1400}}$ | －${ }^{\frac{9}{13,6}}$ | （0，402 |  |  |  |  |  |  |  |  |  |  |  |  |  | ${ }_{2}$ |  |  |
| ${ }^{58,10} 5$ | ss |  | ${ }^{121232022}$ | ciear，very yelow | ${ }_{\text {cis }}^{688}$ | ${ }_{5168}^{565}$ |  | ${ }^{2000}$ | －1590 | ${ }^{13,3}$ | 0.456 | ${ }^{1.53}$ | ${ }^{36,1}$ | 0.02 | 0.01 | ${ }_{\text {cose }}^{\substack{3660 .}}$ | 0.066 | ${ }_{4}^{4188}$ | ${ }^{960}$ | ${ }^{988}$ | 50.8 | 7.74 | 816. | 0.082 | 1. | ${ }^{134}$. | ${ }_{6350} 6$ |
|  |  |  |  | and |  |  |  |  |  |  |  |  |  |  |  |  | $\stackrel{0}{0.0093}$ |  |  |  |  |  |  |  |  |  |  |
|  | ss |  | 92727202 | Clear and coloumes． | 6.57 | ${ }_{840}$ | ${ }^{11,2}$ | ${ }^{240}$ | ${ }^{367}$ | 0.85 | 0.02 | ${ }^{0.288}$ | ${ }^{23}$. | 0.01 | 0.005 | ${ }^{190 .}$ | 0.0083 | ${ }^{1172}$ | 100. | 0.16 | 0.0944 | 0.178 | ${ }^{126 .}$ | 0.0067 | 0.5 | ${ }^{18,3}$ | 9.9 |
| 00，－1．1 | ¢s |  | 3 372023 | Ciara and coloumess | ${ }_{\text {l }}^{1 / 22}$ | ${ }_{\text {c }}^{60}$ | ${ }^{10.1}$ | ${ }^{250}$ | ${ }_{27}^{27}$ | $\stackrel{\text { ¢，}}{\substack{\text { a } \\ \hline}}$ | 0.134 | ${ }^{0.148}$ | ${ }_{15,1}$ | 0.01 | 0.005 | ${ }^{212}$ | 0.085 | ${ }_{94,3}$ | 110. | ${ }_{0}^{0.1}$ | ${ }_{0}^{0.126}$ | 1.49 | ${ }^{18.5}$ | 0.0095 | 0.5 | ${ }^{10} 0$ | ${ }_{30}{ }_{36}$ |
| ${ }^{60.2 .1}$ | ¢ |  |  | ${ }^{\text {ciearana courtess }}$ | \％ <br> 7.85 <br> 7.85 | ${ }^{\frac{502}{540}}$ | ${ }^{\frac{11}{13}}$ | 俍 | ${ }^{\text {288 }}$ | ${ }^{0.5}$ | ${ }_{\substack{0.057 \\ 0.043}}^{0.0}$ |  | 38.9 414 4 | ${ }_{0}^{0.01}$ | （0．005 | ${ }_{\substack{322 \\ 336}}^{\substack{326}}$ |  |  | ${ }^{83}$ |  |  |  | ${ }^{14}{ }_{4}^{148}$ |  | 0.75 |  |  |
|  | ss |  | ${ }^{11122023}$ | Cearand cooumess | ${ }^{7,53}$ | ${ }_{465}^{465}$ | ${ }^{10.3}$ | ${ }^{150}$ | ${ }^{304}$ | ${ }_{587}$ | 0.05 | 0.319 | ${ }^{39,3}$ | 0.01 | 0.005 | ${ }^{328}$ | 0．00193 | ${ }_{958} 9$ | ${ }^{78}$ | 0.1 | ${ }^{0.155}$ | 0.287 | ${ }^{21.3}$ | 0.00102 | 0.88 | 15.6 | ${ }_{51.5}$ |
|  | ss <br> ss |  |  | Cind | ${ }_{\text {T }}^{1727}$ | （int | $\stackrel{10.1}{10.1}$ | ¢190 <br> 100 <br> 1 | ${ }^{\frac{23}{327}}$ |  |  |  | －${ }_{\text {386 }}^{175}$ | ${ }_{0}^{0.01}$ | ${ }_{\text {O．005 }}^{0.005}$ |  | ${ }_{\text {co．}}^{0.027}$ |  | ${ }^{\frac{7}{100}}$ | O．1 0.1 0.1 | ${ }_{\text {O．}}^{\text {O．968 }}$ | － | $\frac{202}{1}$ | ${ }_{\text {O．0．06 }}^{0.005}$ | O．6． <br> 0.5 <br> 0.0 |  | 6．1． <br> 10. <br> 10.0 |
| ${ }^{60.3 .1}$ | ss |  |  |  | ${ }^{6.41}$ | ${ }_{\text {coid }}^{608}$ | ${ }_{103}^{124}$ |  | ${ }_{319}^{281}$ | （1．94 | O．0．088 | －0．173 | －${ }_{154}^{154}$ | ${ }^{0.001}$ | ${ }_{\substack{0.005 \\ 0.005}}^{0.00}$ | ${ }_{\substack{301 \\ 301}}$ | ${ }_{\substack{0.0429 \\ 0.133^{3}}}$ |  | ${ }^{88} 8$ | ${ }^{0.1}$ | ${ }_{\substack{0.106 \\ 0.165}}$ | ${ }^{1,123}$ | ${ }_{4}^{4.6}$ | ${ }^{0.0055}$ | 0．5 | ${ }_{10,3}^{10.4}$ | ${ }_{23}^{234}$ |
| 00．3．1 | ¢ |  | ${ }^{3} 7172023$ | ciearand coumess | ${ }_{122}$ | （ | ${ }_{9} 98$ | 170 | ${ }^{398}$ | ${ }^{3.63}$ | 0.021 | 0.29 | ${ }^{24.4}$ | 0.01 | 0.005 | ${ }^{195}$ | 0.00134 | ${ }^{127 .}$ | ${ }^{24}$ | 0.15 | ${ }^{0.00786}$ | ${ }^{0.296}$ | ${ }_{113 .}$ | 0.00109 | ${ }_{0}^{0.58}$ | ${ }^{19,6}$ | ${ }_{101}$ |
| ${ }^{\frac{122-1}{82,1-1}}$ | ${ }_{\text {ss }}^{\text {ss }}$ |  |  | Ciearara coouless | ${ }^{7}{ }_{7}^{742}$ | ${ }_{\substack{159 \\ 184 \\ 184}}$ | ${ }^{9.8}$ | ${ }^{100}$ | ${ }^{\text {108 }}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{\frac{82}{82} \cdot 1.1}$ | ss |  |  | Cearand | ${ }^{6.98}$ | 179 <br> 179 | ${ }_{9,8}^{9.8}$ | ${ }_{\substack{110 \\ 96}}$ | ${ }^{124}$ | ${ }_{2}^{262}$ | ${ }_{0}^{0.059}$ | ${ }_{0}^{0.095}$ | ${ }^{1.74}{ }^{1.78}$ | 0．01 |  | $\stackrel{27}{275}$ | ${ }_{0}^{0.0062}$ | ${ }_{\text {a }}^{436}$ | $\stackrel{21}{29}$ | ${ }^{0.315}$ | ${ }^{0.0222}$ | ${ }^{0.1086}$ | ${ }^{74}$ | ${ }_{0}^{0.0055}$ | 0.5 0 0 | ${ }_{3}^{3.35}$ | ， |
| 退 | ${ }_{\text {Ns }}$ |  | 9002022 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 既 |  |  |  |  |  | －${ }_{\text {200 }}^{233}$ | $\begin{array}{r}9.4 \\ \hline 10.1\end{array}$ | ${ }^{1300}$ | － | 1.39 <br> 1.01 <br> 1 | ${ }_{0}^{0.078}$ | ${ }_{\substack{0.009 \\ 0.055}}^{0.0}$ | ${ }^{8.22} 11.5$ | ${ }_{0}^{0.01}$ | ${ }_{\substack{0.005 \\ 0.005}}^{0.05}$ | ${ }^{\frac{305}{11}}$ | O．014 | ${ }_{\text {cta．}}^{50.5}$ | ${ }_{5,5}^{25}$ | ${ }^{0.2}$ | ${ }^{\text {O．0．08 } 7}$ | （0．12 | ${ }_{2,}^{29.9}$ | ${ }^{\text {O．0．005 }}$ | ${ }^{\text {O．52 }}$ | ${ }^{4.8}$ | ${ }_{4}^{18,3} 4$ |
|  | ¢ |  |  |  |  | 300 20 219 | －1．4 ${ }^{1.2}$ |  | －${ }_{\text {193 }}^{198}$ |  | 0.02 <br> 0.02 <br> 0.02 | － | －${ }_{8}^{8.94} 8$ | $\stackrel{0.01}{0.01}$ | ${ }_{\substack{0.005 \\ 0.005}}^{0.05}$ | ${ }_{\text {10，}}^{10} 1$. | ${ }^{\text {O．005 }}$ | ${ }^{\text {47，7．}}$ | ${ }_{4}^{4.8}$ | ${ }_{0}^{0.1}$ | ${ }_{\substack{0.0569 \\ 0.499}}$ | $\stackrel{0}{0.06}$ | ${ }_{\substack{224 \\ 229}}^{\text {20，}}$ | ${ }_{0}^{0.005}$ | 0.5 <br> 0.5 <br> 0 |  | （ |
|  | ¢ss |  |  |  | ［ | $\stackrel{2}{25}$ | ${ }^{\text {9，4 }}$ |  |  | ＋1．69 | － | ${ }_{\text {a }}^{0.1038}$ |  | ${ }_{0}^{0.01}$ | （0．005 | ${ }^{10 .}$ | ${ }^{0.0011} 0$ | $\underset{\substack{5.5 \\ 55.4}}{\substack{\text { c．}}}$ | ${ }_{4}^{47}$ | ${ }_{0}^{0.1}$ | ${ }^{0.0027}$（0074 | ${ }_{\text {O }}^{0.252}$ | ${ }^{1.8}$ | ${ }_{\text {O．OO5 }}^{0.0058}$ | －0．5 |  | （i．0． |
|  | Ss |  | ${ }^{1447203}$ | Clearand coumess | \％ 7 7，93 |  | ${ }^{9.4}$ | － 1100 | $\stackrel{177}{1705}$ | ${ }_{2,15}^{2,15}$ | ${ }^{0.027}$ |  | ${ }_{\substack{16,7 \\ 6.66}}$ | $\frac{0.01}{0.01}$ | ${ }_{\text {O．005 }}^{0.005}$ | ${ }_{\text {10，}}^{105}$ | － 0.006 | ${ }_{\text {59，9 }}^{59.9}$ | $\frac{43}{76}$ |  | －${ }^{0.0095}$ | ${ }^{\frac{0222}{020} 5}$ | ${ }_{\text {2 }}^{25}$ | ${ }_{\text {O．005 }}^{0.0055}$ | 0.5 <br> 0.5 <br> 0.5 |  | ${ }_{\text {O．}}^{0.896}$ |
|  | ${ }_{\text {FRM }}$ | $\stackrel{r}{r}$ | 9272022 | Neanor futinctases | ${ }_{\text {Tr }}^{\text {T，75 }}$ | －${ }^{324}$ | ${ }^{10.3}$ |  | ${ }^{\text {ci4．}} 168$ | $\underset{\substack{1.87 \\ 301}}{\substack{\text { a }}}$ | ${ }_{0}^{0.09} 0$ | ${ }_{\substack{0.348 \\ 0.485}}$ |  | ${ }^{0.01}$ | ${ }^{0.005}$ | ${ }_{40}^{40}$ | ${ }_{0}^{0.00065}$ | ${ }_{\text {ck }}^{56}$ | ${ }^{8,2}$ | ${ }^{0.105}$ | － | ${ }_{\substack{0.1075}}^{0.052}$ | ${ }_{\substack{34.7 \\ 365}}$ | －0．0065 | 0.5 | ${ }_{\text {7 }}{ }_{7}^{732}$ | ${ }^{94.1}$ |
| $\frac{7}{71-1}$ | $\underset{\substack{\text { remm } \\ \text { rem }}}{ }$ | $r$ | ${ }_{\text {l }}$ | Mean（fupleaes | ${ }^{17,73}$ | ${ }_{\substack{236 \\ 340}}$ | ${ }^{10.1}$ | ${ }_{1}^{140}$ | ${ }_{149}^{149}$ |  | ${ }_{\text {lol }}^{0.005}$ |  |  |  | ${ }_{\substack{0.0005}}^{0.005}$ | ${ }_{481}^{48 .}$ |  |  |  |  |  | ${ }_{0}^{0.0025}$ |  |  |  |  |  |
| ${ }^{\frac{12}{12,-1}}$ | ${ }_{\substack{\text { ss } \\ \text { ss }}}$ | r |  |  | ${ }_{\text {che }}^{7.45}$ | $\underset{\substack{337 \\ 362}}{\substack{\text { and }}}$ | ${ }^{10}$ | ${ }_{2}^{220}$ | ${ }^{235}$ |  | 0.025 <br> 0.02 <br> 0 | ${ }_{0}^{0.312}$ | ${ }^{1591} 14.5$ | ${ }^{0.01}$ | （0．005 | ${ }^{\text {302 }}$ | ${ }^{0.00738}$ | ${ }_{\substack{812 \\ 798}}$ | ${ }^{8.9}$ | ${ }^{0.1}$ | －0．114 | ${ }^{0.224}$ | ${ }_{211}^{213}$ | ${ }^{0.0043}$ | －0．5 | ${ }_{\text {\％}}^{7.95}$ | ${ }^{80,3}$ |
| ${ }^{717.2 .2 .1}$ | Ss | $\stackrel{r}{r}$ |  |  | ${ }_{7}^{762}$ |  | ${ }^{9} 9$ | ${ }_{2}^{210}$ | ${ }^{249}$ |  | $\stackrel{0.021}{0.02}$ | ${ }^{0.21}$ | 29.9 | 0.01 | 0.005 | ${ }_{3}^{315 .}$ | ${ }^{0.115}$ | ${ }_{85,9} 8$ | ${ }^{8.8}$ | 0.1 | 0.144 | 0.412 | ${ }_{20.3}^{20 .}$ | 0.00616 | 0.5 | ${ }_{8,48}$ | ${ }_{87,3}$ |
| ${ }^{\frac{172.3 .1}{713.1}}$ | Ss | $r$ | ${ }^{51272022}$ | Ciarand counilss | $\stackrel{.64}{6.84}$ | $\stackrel{\text { 203 }}{283}$ | ${ }_{9,8}^{9,}$ | ${ }_{150}$ | ${ }_{176}$ | $\stackrel{1.71}{ }$ | ${ }_{0} 0.02$ | $\stackrel{0.052}{0.002}$ | ${ }_{5}^{5.95}$ | 0.01 | 0.005 | ${ }^{\frac{235}{251}}$ | $\xrightarrow{0.0095}$ | ${ }_{6}^{628}$ | ${ }^{8.1}$ | ${ }_{0}^{0.1}$ | $\stackrel{0}{0.039}$ | ${ }_{0}^{0.866}$ | ${ }_{2}$ | ${ }^{0.005}$ | ${ }^{0.5}$ | ${ }_{4.48}^{4.05}$ | ${ }^{\frac{15.51}{0.21}}$ |
|  | ¢ | r |  | Sill |  | ${ }^{330}$ | ${ }^{11,9}$ | ${ }^{170}$ | ${ }^{199}$ | ${ }_{\text {cher }}^{\substack{365 \\ 565}}$ | 0．021 | ${ }^{0.0075}$ | ${ }^{10.4}$ | ${ }^{0.01}$ | ${ }_{0}^{0.005}$ | ${ }_{\substack{346 \\ 3 \\ 345}}$ | ${ }^{0.0 .0313}$ | $\frac{7,}{77}$ | ${ }^{6.6}$ |  | ${ }^{0.0039}$ | ${ }^{0.8}{ }_{0}^{081}$ | ${ }_{\substack{8.7 \\ 118}}^{\text {18，}}$ | （0．087 |  |  | S19 |
| ${ }^{17.3 .7}$ | ${ }_{\text {ss }}^{\text {ss }}$ | $r$ | 3 372023 |  | ${ }_{6}^{693}$ | ${ }^{310}$ | ${ }^{10.4}$ | ${ }_{160}$ | ${ }_{\substack{1168 \\ \hline 188}}$ | － | ${ }_{0} 0.025$ | ${ }_{0}^{0.077}$ | ${ }_{7}^{7.85}$ | ${ }_{0}^{0.01}$ | ${ }_{0}^{0.005}$ | ${ }_{284}$ | 0.0164 | ${ }_{66.2}$ | 10. | ${ }_{0}^{0.1}$ | ${ }^{0.0528}$ | － | ${ }_{3,1}$ | ${ }_{0}^{0.005}$ | 0.5 <br> 0.5 |  | （e．292 |
| $\frac{1}{2 \times-1}$ | ${ }_{\text {rerm }}^{\text {Rram }}$ | $r$ |  | Neanotainicaes | ${ }_{\text {\％}}^{\text {\％}}$ |  |  |  | ${ }_{2025}^{2085}$ | ＋133 | O． |  |  | ${ }^{0.02}$ | ${ }^{\text {O．01 }}$ |  | ${ }^{0.005}$ |  | 25 |  |  |  | 516．5 |  |  |  |  |
| 2，1． | ${ }_{\text {frim }}$ | $r$ | ${ }_{\text {l }}^{1 / 1112023}$ | Neanof fotpineates | ${ }_{7}^{7,84}$ | ${ }_{4}^{43}$ | 10.5 | ${ }_{150}$ | ${ }^{2200}$ | ${ }^{203}$ | ${ }_{0}^{0.03}$ | $\stackrel{0}{0.0085}$ | $\stackrel{\text { 12，}}{12}$ | ${ }^{0.0015}$ | ${ }_{0}^{0.000}$ | ${ }^{\text {c／} 565}$ | ${ }_{0}^{0.0075}$ | ${ }_{8}^{805}$ | ${ }_{65,5}$ | ${ }_{0}^{0.16}$ | ${ }_{\substack{0.0799}}^{0.754}$ | ${ }_{0}^{0.005}$ |  | ${ }_{0}^{0.0075}$ | ${ }_{0.75}^{0.5}$ | ${ }_{128}^{138}$ | ${ }_{110.5}^{110}$ |
|  | ${ }_{\text {erem }}^{\text {en }}$ | $r$ | ${ }^{\frac{36}{26203}}$ | Meanofaturaies | ${ }^{\text {7，}}$ | ${ }_{600}^{600}$ | ${ }^{10.3}$ | ${ }^{\text {100 }}$ | ${ }_{\text {2635 }}^{263}$ | －${ }_{\text {3，} 625}$ | ${ }^{0.038}$ | ${ }_{\text {one }}^{0.142}$ | ${ }_{\text {l }}^{129}$ | ${ }^{0.015}$ | 0.0075 | ${ }^{1645}$ | 0.0075 | ${ }_{\text {ckis }}^{8.35}$ | ${ }_{66}^{66}$ | ${ }_{0}^{0.16}$ | ${ }^{0.0028}$ | ${ }_{0}^{0.0085}$ |  | 0．008 | O． 0.78 |  |  |
| 23， | ss | $r$ | ${ }^{10442022}$ | Clarand colounses | ${ }_{8,16}$ | ${ }_{473}$ | ${ }^{121}$ | ${ }^{200}{ }^{200}$ | ${ }^{249}$ | ${ }^{1.48}$ | ${ }_{0}^{0.027}$ | ${ }_{0}^{0.284}$ | ${ }^{19} 9$ | 0.01 | ${ }_{0}^{0.005}$ | ${ }_{487}^{487}$ | 0.046 | ${ }_{88,1}$ | ${ }^{58}$ | ${ }^{0.1}$ | ${ }^{0.0812}$ | ${ }_{0}^{0.355}$ | ${ }_{4.6}$ | 0.005 | ${ }_{0}^{0.5}$ | ${ }_{6} .98$ | ${ }^{2.16}$ |
| ${ }^{212,3.3}$ | ¢ss | $\stackrel{r}{r}$ | $\underbrace{}_{\substack{1 / 22023 \\ 362023}}$ |  | ${ }^{7} 7.85{ }^{7}$ | － $\begin{gathered}4.48 \\ 426\end{gathered}$ | 10.8 10.4 10.4 | （200 | 262 <br>  <br> 227 <br> 22 | ¢ | －0．034 | ${ }_{0}^{0.284} 0$ | ${ }_{\substack{18,8 \\ 15,1}}$ | ${ }^{0.01}$ | ${ }_{\text {O．0．05 }}^{0.005}$ | ${ }_{\text {Sl }}^{510}$ | （0．026 |  | ${ }^{41}$ | 0.1 0.11 | （0．0022 | ${ }_{\text {O，}}^{0.358} 0$ | $\stackrel{9}{21.7}$ | （0．079 | 0.54 <br> 0.54 <br> 0.54 |  | － |
| ${ }^{\frac{1}{23} 31.1}$ | ${ }_{\substack{\text { rem } \\ \text { RRM }}}$ | $\stackrel{r}{r}$ |  | Meanofotiricies | －${ }_{\text {7．45 }}^{7}$ |  | －${ }_{\text {10，}}^{10.6}$ | ${ }^{\text {coio }}$ | $\stackrel{2095}{20.5}$ | ${ }^{0.6}$ |  |  |  |  | （0．005 |  |  |  |  | （0．105 |  | ${ }_{\substack{0.4555 \\ 0.652}}$ | ＋1， | ${ }^{\text {0．0．141 }}$ |  | ${ }_{\text {lin }}^{11.55}$ | ${ }_{\substack{112 \\ 884}}^{128}$ |
|  | ${ }_{\text {rem }}^{\text {RRM }}$ | $\stackrel{r}{r}$ | $\xrightarrow{11122023}$ | Meanototiplicies | ${ }^{747}$ | ${ }^{329}$ | ${ }_{9} 98$ | ${ }_{\substack{180 \\ 180 \\ 180}}$ | ${ }^{230}$ | ${ }_{10,09}^{1.09}$ | ${ }_{\substack{0.035 \\ 0.035}}^{0.0}$ | ${ }_{\substack{0 \\ 0.2075 \\ 0.295}}$ | － | ${ }_{0}^{0.01}$ | ${ }_{0}^{0.0005}$ |  | （0．0235 |  | ${ }^{21}$ | ${ }_{0}^{0.17}$ |  | －0．495 | ${ }_{5}^{525}$ |  | 0.5 <br> 0.5 | ${ }_{1}^{124}$ | （17．3． |
| $\frac{8}{38.2 .1}$ | ${ }_{\text {ss }}$ | r | ${ }^{515252022}$ | Coarandofolouress | ${ }_{7}^{1725}$ | ${ }_{\substack{306 \\ 306}}$ | ${ }^{10.3}$ | ${ }^{160}$ |  | ${ }^{4.78}$ | ${ }_{0}^{0.02}$ |  |  |  | ${ }_{0}^{0.005}$ |  |  | ${ }_{64,5}$ |  |  |  |  |  |  |  |  |  |
|  | ${ }_{\text {ss }}^{\text {ss }}$ | r | ${ }_{\text {a }}^{\text {9272022 }}$ | Ciear and ooumeses | \％ | ${ }_{\substack{334 \\ 305}}$ | －${ }_{\text {10，7 }}^{10.3}$ |  | ${ }_{204}^{204}$ | ${ }_{\substack{1.188 \\ 1.188}}$ | ${ }_{0}^{0.02}$ | ${ }^{0.004}$ | ${ }^{104}$ |  | ${ }^{0.0005}$ |  | ${ }_{\substack{0.0598 \\ 0.019}}^{\substack{\text { a }}}$ | ${ }_{\substack{64.3 \\ 652}}^{6}$ |  |  | ${ }^{0.0938}$ | ${ }_{0}^{0.399}$ | ${ }^{17.1}$ | ${ }^{0.005}$ | 0．5 | ${ }_{\text {l }}^{10.7}$ | ${ }_{\text {454 }}^{434}$ |
| 32．1 | ${ }_{\text {ss }}$ | $r$ | 3772023 | Clearand columess | ${ }_{7.2}$ | ${ }_{480}$ | 10 | 180 | ${ }^{204}$ | 1.67 | 0.02 | ${ }_{0}^{0.1}$ | 9．79 | 0.01 | 0.005 | 74 | 0．0168 | 63.5 | 20. | 0.1 | ${ }_{0}^{0.104}$ | －0278 | 10.4 | 0.005 | 0.5 | ${ }^{11.1}$ | ${ }_{47.6}$ |



\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline \multicolumn{4}{|c|}{${ }_{\text {bc css }}$} \& ${ }_{\text {atem }}^{\text {atw Maximum (1) }}$ \& \& \& \& \& \& 950 \& ${ }_{6}^{90}$ \& $\stackrel{50}{10}$ \& ${ }_{\substack{10000 \\ 1000}}$ \& $\stackrel{298}{8}$ \& \& ${ }^{12000}$ \& ${ }^{0.5445}$ \& \& ${ }_{250}^{1500}$ \& ${ }_{\substack{9000 \\ 600}}$ \& ${ }_{\text {40 }}^{14(9)}$ \& ${ }_{\text {20.900 (5) }}^{1500}$ \& (6) \& ${ }^{40-1800]}$ \& ${ }_{33}$ (9) \& \& (6) <br>
\hline \multirow{3}{*}{Station} \& \multirow{3}{*}{Sampeo Type} \& \multirow{3}{*}{Compliance} \& \multirow{3}{*}{Date Sampled} \& Parameer \& PH(Fibel \& Specific
Conductivity
$25^{\circ} \mathrm{C}$ (Field) \&  \& Akaliny \&  \& Alum \& Antinom \& Assenic \& Batium \& Senlium \& Bisuun \& Boon \& ${ }_{\text {casmum }}$ \& Calcum \& Choride \& Chromium \& Cooat \& copper \& ${ }_{\text {Ion }}$ \& Lead \& Lutium \& \& Manganese <br>
\hline \& \& \& \& Fraction \& Tot \& тот \& тот \& Tot \& Dis \& ${ }_{\text {on }}$ \& Dis \& ${ }_{\text {dis }}$ \& dis \& ${ }_{\text {on }}$ \& 0 ls \& 015 \& Dis \& ois \& 0 ois \& ${ }_{0} \mathrm{is}$ \& Dis \& dis \& 0is \& ${ }^{15}$ \& ${ }_{\text {Lis }}$ \& ois \& ois <br>
\hline \& \& \& \& $\xrightarrow{\text { method Dotatioction Limit }}$ \& ${ }_{0}^{\text {pH }}$ \& $\stackrel{\text { ustm }}{1}$ \& c

0.1
0. \& mgh \& ${ }_{\text {mgh }}^{0.5}$ \& ${ }_{\text {man }}^{\text {ma }}$ \& $\xrightarrow{\text { nen }}$ \& $\stackrel{\text { Hen }}{\substack{\text { en }}}$ \&  \&  \& ${ }_{0}^{\text {nen }} 0$ \& $\stackrel{\text { ma }}{10}$ \& ${ }_{\text {Len }}^{\text {nen }}$ \& men \& mg \& ${ }_{\text {ren }}^{0.1}$ \&  \& $\xrightarrow{\text { nen }}$ \& $\stackrel{\text { ma }}{1}$ \&  \& ${ }_{\text {mal }}^{0.5}$ \& ${ }_{\text {mon }}^{\text {ma }}$ \& (1) <br>

\hline 10, 10.1 \& ${ }_{\text {ss }}^{\text {ss }}$ \& \&  \& Noderatev wrob, moderaely yey \& $\xrightarrow{7,14}$ \& ${ }^{2162}$ \& - 20.9 \& ${ }^{630}$ \& ${ }_{5}^{507}$ \& ${ }_{\text {201. }}^{20.5}$ \&  \& - \& - \& ${ }_{0}^{0.01}$ \& ${ }_{\text {¢ }}$ \& ${ }^{864}$ \& | O.005 |
| :--- |
| 0.136 |
| 0.053 | \& ¢, \& ${ }^{140 .}$ \&  \& (0.005 \&  \& ${ }^{251}$ \&  \& $\stackrel{\text { O. }}{\substack{\text { 3.54 }}}$ \& ${ }_{\substack{\text { a } \\ \hline 0.3 \\ 31.3}}$ \& - <br>

\hline ${ }^{10.10 \cdot 1.1}$ \& (css \& \&  \&  \& + 7.5 \&  \& ${ }^{14,1}$ \& ${ }^{160}$ \&  \&  \& - \& (0.483 \& - ${ }_{\text {13, }}^{38.2}$ \& 0.018 \& ${ }_{0}^{0.005}$ \& ${ }^{173}$. \& ${ }_{\substack{0.0053 \\ 0.0153}}^{0.0}$ \&  \& (14. \& - ${ }_{4.188}$ \&  \& $\stackrel{2.84}{7.5}$ \& ${ }^{101}$ \&  \& $\stackrel{0.5}{0.57}$ \&  \&  <br>
\hline 107.1. \& ¢ss \& \&  \& Veer utibas isiony gey \& (7, \& ${ }_{\text {c }}^{1100}$ \& ${ }_{\substack{21 \\ 13.1}}$ \& ${ }^{180}{ }^{180}$ \& ${ }_{\text {cke }}^{598}$ \& ${ }^{40.1} 80.9$ \& O. ${ }^{0.374}$ \& ${ }_{0}^{0.23}$ \&  \& ${ }_{0}^{0.01}$ \& ${ }_{0}^{0.005}$ \&  \& (0.024 \& ${ }^{174}{ }^{172}$ \& (15. \& 0.19
0.1
0.1 \& ${ }_{\text {1.17 }}^{1.97}$ \& ${ }_{4}^{3.95}$ \& ${ }_{4}^{74.3}$ \& ${ }^{0.0363}$ \&  \& ${ }_{\substack{27.1 \\ 18.6}}$ \& (iser <br>
\hline  \& (is \& \& ${ }_{\text {a }}^{317172033}$ \& Nosamper equired dis manh \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& <br>
\hline (1081-1 \& ¢s \& \& ${ }^{12772022}$ \& Sols \& ${ }_{\text {c }}$ \& ${ }_{\substack{600 \\ 519}}^{\text {col }}$ \& ${ }^{\frac{13,3}{11}}$ \& ${ }^{120}$ \& ${ }_{\substack{\text { 286 } \\ 316}}$ \& ${ }_{\text {c, }}^{\text {S.2 }}$ \&  \& $\stackrel{\text { 0.67 }}{0.67}$ \& ${ }^{\text {Tri. }}$ \& O.01 \& ${ }_{0}^{0.005}$ \& ${ }^{114}$ \& ${ }_{0}^{0.0005}$ \& $\underset{8}{\substack{76.6 \\ 81.4}}$ \& ${ }^{6.5}$ \& ${ }_{0}^{0.1}$ \&  \&  \& ${ }^{12}$ \& ${ }_{0}^{0.00058}$ \&  \& ${ }_{\text {2 }}^{27.4}$ \&  <br>
\hline ${ }^{1099}$ \& ¢ \& \& - \& Nosmpere eaured his monh \& $\stackrel{\text { 8,74 }}{ }$ \& ${ }_{460}$ \& ${ }^{21,9}$ \& $\frac{170}{140}$ \& ${ }^{24,4}$ \& $\frac{1}{11.1}$ \& 0.815 \& $\stackrel{18.9}{18.9}$ \& $\stackrel{-1}{10.1}$ \& $\underline{0.01}$ \& 0.005 \& $\stackrel{-}{237}$ \& 0.005 \& $\frac{-7}{6.52}$ \& $\stackrel{5}{5.9}$ \& $\stackrel{-}{0.1}$ \& $\stackrel{0.055}{ }$ \& ${ }_{0}^{0.492}$ \& ${ }^{77}$ \& 0.00192 \& ${ }^{3,78}$ \& -1.98 \& ${ }_{4}^{4.82}$ <br>
\hline ${ }^{1097}$ \& ss \& \& ${ }^{1131302022}$ \& Moderaty utude, modeatey gey \& 7.98 \& ${ }^{387}$ \& ${ }_{14.6}$ \& 170 \& 28. \& ${ }^{8.64}$ \& 0.696 \& 16.4 \& 6.06 \& 0.01 \& 0.005 \& ${ }^{281}$ \& 0.005 \& 7.12 \& 3.5 \& 0.1 \& 0.036 \& 0.61 \& 12.5 \& 0.041 \& 5.09 \& 261 \& ${ }_{3.83}$ <br>
\hline $\frac{109-1 / 1}{100 \cdot 1 /}$ \& - $\frac{\text { Ns }}{\text { Ns }}$ \& \& ${ }^{36712023} 9$ \& Nosampe equired dis monh \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& <br>

\hline ${ }^{110 \cdot 9.1}$ \& ¢ | Ns |
| :---: |
| Ns | \& \& ${ }_{\substack{113502022}}^{3 / 7172023}$ \& No smple (ooeminiod not oberepeseenalave) \& \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - \& - <br>

\hline \& ¢ \& \&  \& Coarand colouress \& $\xrightarrow{709}$ \& ${ }^{12061}$ \& ${ }^{16}$ \& ${ }^{600}$ \& ${ }_{\text {cki }}^{279}$ \& $\stackrel{64}{61}$ \& ${ }_{0}^{0.135}$ \& $\stackrel{9.65}{\substack{707}}$ \&  \& ${ }_{0}^{0.02}$ \& ${ }_{0}^{0.015}$ \& $\underset{\substack{1610 . \\ 2.500}}{ }$ \& 0 \&  \& $\stackrel{10}{10}$ \& $\stackrel{0}{0.94}$ \& $\underset{\substack{194 \\ \text { ien }}}{ }$ \& ${ }_{0}^{02}$ \&  \& ${ }_{0}^{0.162}$ \& ${ }^{1224}$ \& ${ }^{182}$ \& ${ }_{\substack{184 \\ 651}}^{18 .}$ <br>
\hline P1 \& (tss \& \&  \& Clearand coluress \& $\frac{7.1}{7.1}$ \&  \&  \& ${ }^{790}$ \& ${ }_{\substack{281 \\ 288 \\ 208}}$ \& $\stackrel{51}{42}$ \&  \& (17.78 \& ¢ \& ${ }_{\substack{0.05 \\ 0.05}}$ \& ${ }^{0.025}$ \&  \& ${ }_{\text {O}}^{0.025}$ \& ¢ \&  \&  \& ${ }_{\substack{3.61 \\ 3,74}}$ \& ${ }^{0.25}$ \& $\underset{\substack{132000}}{\substack{12000}}$ \& ${ }^{0.034} 0$ \& - \& ${ }^{23.5}$ \& ${ }^{\frac{65,1}{102}}$ <br>
\hline ${ }^{\frac{p}{P 2}}$ \& ss \& \&  \& Coearanos siontury yelow \& $\stackrel{\text { T, } 722}{ }$ \& ${ }^{\text {9020 }}$ \&  \& ${ }^{860}$ \& ${ }^{206}$ \& ${ }^{\frac{32}{74}}$ \& ${ }_{0}^{0.14}$ \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& <br>
\hline \& ss \& \& 914142022 \& \& ${ }^{7,75}$ \& ${ }^{21577}$ \& ${ }^{16.5}$ \& 980 \& ${ }^{337}$ \& $\stackrel{73}{78}$ \& 0.1 \& 0.86 \& ${ }_{467}$ \& 0.05 \& 0.025 \& ${ }^{3450}$. \& ${ }^{0.025}$ \& ${ }^{828}$ \& ${ }^{210}$. \& ${ }^{1.27}$ \& ${ }^{4.5}$ \& 0.91 \& 140. \& 0.047 \& ${ }^{25}$ \& ${ }_{31.6}$ \& ${ }_{3} 30$. <br>
\hline P2 \& ¢ \& \& ${ }^{22442023}$ \& vensolmy \& ${ }_{7}{ }_{7}{ }^{2}$ \& ${ }_{2120}^{2100}$ \& ${ }^{15,5}$ \& ${ }^{800}$ \& ${ }_{241}$ \& ${ }_{4.1}^{4 .}$ \& ${ }_{0} 0.048$ \& ${ }_{0}^{0.787}$ \& ${ }_{480}$ \& ${ }_{0}^{0.02}$ \& ${ }_{0}^{0.009}$ \& ${ }_{2630}$ \& 0.01 \& $\stackrel{621}{621}$ \& ${ }^{200}$ \& $\stackrel{1}{1.04}$ \& ${ }_{3,34}$ \& ${ }_{0.026}^{0.029}$ \& ${ }_{5609}^{560}$ \& ${ }_{0}^{0.031}$ \& ${ }_{3.1}{ }^{2.5}$ \& ${ }^{3}$ \& ${ }^{364}$ <br>

\hline ${ }^{\text {P3 }}$ \& ¢ss \& \&  \& 隹 \& | 722 |
| :--- |
| 7.13 | \& ${ }^{2197}$ \& ${ }_{16,3}$ \& 100 \& \& ${ }^{\text {6.8 }}$ \& ${ }^{0.1}$ \& | 1.98 |
| :--- |
| 1.14 |
| 1.4 |
| 1 | \&  \& ${ }_{0}^{0.05}$ \& ${ }_{0}^{0.025}$ \& ${ }^{3680}$ \& ${ }^{0.025}$ \& | 58, |
| :--- |
| 717 |
| 1.7 | \& ${ }^{190}{ }_{20}^{10 .}$ \&  \& ${ }^{4.8}$ \& \& ${ }_{\substack{684 \\ 280}}^{\text {280 }}$ \& | 0.025 |
| :--- |
| 0.025 |
| 0.025 | \& ${ }^{4.8}$ \&  \&  <br>

\hline \& ${ }_{\text {ss }}$ \& \& ${ }^{121712022}$ \& Silathy ubba, Iflyly orave \& ${ }^{2} 7.16$ \& ${ }^{12225}$ \& ${ }^{11.6}$ \& ${ }^{960}$ \& ${ }^{287}$ \& ${ }_{4}^{4}$ \& ${ }^{0.0042}$ \& 1.12 \& ${ }_{\text {cta }}^{540}$ \& ${ }^{0.00}$ \& 0.005 \& ${ }_{3}^{32200}$ \& 0.005 \&  \& $\underset{\substack { 20 \\ \begin{subarray}{c}{20 \\ 120{ 2 0 \\ \begin{subarray} { c } { 2 0 \\ 1 2 0 } } \\{\hline}\end{subarray}}{ }$ \& $\stackrel{0.97}{0.98}$ \& ${ }^{3,77}$ \& 0.311 \& ${ }^{1880}$ \& 0.01 \& ${ }^{1.1 .15}$ \& ${ }^{29,5}$ \& ${ }^{389}$ <br>
\hline ${ }^{\frac{18}{84}}$ \& ¢ss \& \&  \&  \& ${ }^{\text {7,36 }}$ \& ${ }^{2004}$ \& ${ }^{14.4}$ \& ${ }^{940}$ \& ${ }^{\text {219 }}$ \& ${ }^{\text {4 }}$ 3 8 \& -0.046 \& - \& ¢ \& $\stackrel{0.02}{0.02}$ \& \& ${ }^{3230}$ \& \& \&  \& O.98
0.87 \& ${ }^{\frac{3}{3} 3.8}$ \& ${ }^{0.28}$ \&  \& -0.028 \&  \& ${ }^{22}{ }^{22,6}$ \&  <br>
\hline P4 \& ss \& \& 91/42022 \& \& 8.07 \& ${ }^{2300}$ \& ${ }^{16}$ \& ${ }^{890}$ \& ${ }^{321}$ \& 8 \& ${ }^{0.1}$ \& $\stackrel{1.87}{ }$ \& ${ }^{171}$ \& 0.05 \& ${ }^{0.025}$ \& ${ }^{2460}$ \& ${ }^{0.025}$ \& ${ }_{80,6}^{80,6}$ \& ${ }^{180}$ \& 0.76 \& ${ }^{282}$ \& ${ }^{1.81}$ \& ${ }^{7} 7380$. \& ${ }_{0}^{0.053}$ \& ${ }^{2.5}$ \& ${ }^{29}$ \& ${ }^{654}$ <br>

\hline ${ }^{84}$ \& \% \& \& ${ }^{22424023}$ \& 为 \& ${ }_{7}^{7.09}$ \& 2300 \& ${ }^{13.4}$ \& ${ }_{900}$ \& - $\begin{array}{r}\text { 311 } \\ -11\end{array}$ \& \% ${ }_{3}$ \& ${ }_{0}^{0.04}$ \& $\stackrel{1.25}{1.25}$ \& ${ }_{\text {cob }}^{108 .}$ \& ${ }_{0}^{0.02}$ \& ${ }^{0.009}$ \& ${ }_{2}^{2300}$ \& ${ }_{0}^{0.01}$ \& | $\frac{78.6}{78.2}$ |
| :---: |
| 8. | \& ${ }^{170 .}$ \& ${ }^{0.122}$ \& ${ }_{2,273}^{2 / 2}$ \& ${ }_{0}^{0.173}$ \&  \& O.0.02 \& - \& ${ }_{\substack{28 \\ 20 . \\ 0.0}}$ \&  <br>

\hline ${ }^{\text {P10 }}$ \& \& \& ${ }^{5126202022}$ \& Sighly stity nas silighly frane \& ${ }^{7} \mathbf{7}$ \& ${ }^{11110}$ \& ${ }^{1588}$ \& ${ }^{560}$ \& ${ }^{329}$ \& ${ }_{\text {236 }}^{232}$ \& 0.036 \& ${ }_{0}^{0.855}$ \& ${ }^{797}$ \& ${ }^{0.01}$ \& ${ }_{0}^{0.005}$ \& ${ }^{1330}{ }^{130}$ \& ${ }_{0}^{0.00783}$ \&  \& ${ }^{100}$ \& ${ }_{0}^{0.54}$ \& ${ }^{1386}$ \& ${ }^{0.326}$ \& ${ }_{504}^{504}$ \& 0.012 \& ${ }^{324}$ \& ${ }^{198}$ \& <br>
\hline ${ }^{\text {P10 }}$ \& ${ }_{\text {ss }}^{\text {ss }}$ \& \& ${ }_{1}^{121712022}$ \& Coars. Sfloly yelow \& $\xrightarrow{8.02}$ \& ${ }_{1414}^{1449}$ \& ${ }^{159}$ \& ${ }_{7}^{600}$ \& ${ }^{323}$ \& ${ }_{64}{ }_{6}$ \& ${ }_{0}^{0.051}$ \& 0.979 \& ${ }_{921}$ \& ${ }_{0}^{0.02}$ \& ${ }_{0}^{0.01}$ \& ${ }^{1790}{ }^{1909}$ \& ${ }_{0}^{0.01}$ \& ${ }_{93.1}$ \& ${ }_{140}$ \& ${ }_{0}^{0.92}$ \& ${ }_{2,56}$ \& ${ }_{0}^{0.47}$ \& ${ }_{566}$ \& 0.072 \& ${ }_{4}^{4.9}$ \& ${ }^{21.9}$ \& ${ }_{50}^{50 .}$ <br>

\hline \& ¢ \& \&  \& vers sifity youv, on utbibly \& \% 7.7 \& - 1200 \& ${ }^{\frac{1488}{109}}$ \& ${ }^{\text {550 }}$ \& ${ }^{294}$ \& | 4.15 |
| :--- |
| 354 | \& ${ }^{0.034}$ \& (1.12 \&  \& ${ }_{0}^{0.01}$ \& ${ }_{0}^{0.0055}$ \& ${ }^{11300}$ \& ${ }^{0.005}$ \& ${ }_{\substack{8.8 \\ \hline 13 \\ \hline 18}}$ \& 54. \& ${ }_{0}^{0.54}$ \& (1.188 \& ${ }_{0}^{0.041}$ \& ${ }_{\text {cies }}^{4 .}$ \& ${ }_{0}^{0.0564}$ \& ${ }_{\text {3, }}^{3}$ \& - 16.8 \& ${ }_{\frac{427}{428}}$ <br>


\hline \& ss \& \& ${ }^{1129292022}$ \& ciearand colouress \& ${ }_{6.87}$ \& ${ }_{645}^{645}$ \& ${ }^{11,8}$ \& 110 \& ${ }^{341}$ \& | 10.6 |
| :--- |
|  |
| 108 | \& ${ }^{0.186}$ \& ${ }^{0.0 .87}$ \& ${ }^{16,5}$ \& ${ }^{0.01}$ \& ${ }_{0}^{0.005}$ \& ${ }^{1887}$ \& $\stackrel{0}{0.0367}$ \& ${ }^{114 .}$ \& ${ }_{5} 5$ \& 0.19 \& ${ }^{2488}$ \& ${ }_{6}^{6.93}$ \& ${ }^{21.7}$ \& ${ }_{0}^{0.0025}$ \& ${ }_{0}^{0.5}$ \& ${ }^{13.7}$ \& ${ }^{1296}$ <br>


\hline \& ¢s \& \&  \&  \& ${ }^{6.964}$ \&  \& | 11.1 |
| :--- |
| 102 | \& ${ }^{18} 8$ \& | 888 |
| :--- |
| 473 |
| 8 | \&  \& $\stackrel{0.16}{0.16}$ \& ${ }_{\substack{0.068 \\ 0.005}}^{0.0}$ \& ${ }^{20.35} 8.13$ \& $\xrightarrow{0.02}$ \& ${ }_{0}^{0.005}$ \& $\stackrel{1235}{1235}$ \& | 0.0695 |
| :--- |
| 0.021 | \& ${ }_{\substack{266 \\ 155}}$ \& | 10 |
| :--- |
| 8. |
| 8. | \& | 0.21 |
| :--- |
| 0.1 |
| 0. | \&  \& ${ }^{\text {c, }}$ \& ${ }_{\text {c }}^{40}$ \& ${ }^{0.0023}$ \& | 1. |
| :---: |
| 0 | \& ${ }^{34.9}{ }^{31}$ \& ${ }_{\text {l }}^{109.5}$ <br>

\hline ${ }^{\frac{12}{12}}$ \& ¢ ${ }_{\substack{\text { Ns } \\ \text { Ns }}}$ \& \& $\frac{9.112022}{111292022}$ \& Nosanpe Well was dy. \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& <br>
\hline \& ss \& \& (1262033 \& Coter \& ${ }_{6}^{684}$ \& ${ }^{783}$ \& $\stackrel{11.2}{-2}$ \& 110 \& ${ }^{531}$ \& 9.71 \& 0.112 \& ${ }^{0.123}$ \& ${ }^{22.7}$ \& $\stackrel{0.0}{ }$ \& ${ }^{0.005}$ \& ${ }^{120}$ \& 0.024 \& \& ${ }_{19}$ \& $\stackrel{0}{0}$ \& $\stackrel{0.52}{ }$ \& ${ }_{6.76}^{6.6}$ \& 314 \& 0.0019 \& 0.5 \& 2.7 \& 30.6 <br>
\hline
\end{tabular}


simme











| вc csR |  |  |  | ${ }_{\text {ata }}^{\text {atw Mximum（1）}}$ |  | ［1．31－184．43］ | ${ }^{0.2 .2(4)}$ | ${ }^{400} 10$ | ${ }^{400}$ | ${ }^{2500^{20 / 500(5)}}$ |  |  | ${ }_{\substack{20 \\ 10}}$ |  | ${ }^{\frac{0.5175(5)}{20}}$ | 17000 | 2500 | $\frac{88.2929}{500}$ |  |  |  |  | ${ }^{25}$ |  |  | ${ }^{\frac{85}{85}}$ | ${ }^{20}$ | ${ }^{\frac{15}{523000}} 3$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sation | Sample |  | Dates Sampled | Paramear | Morvedenm | $\underbrace{\text { a }}_{\substack{\text { Nitogen－} \\ \text { ammona }}}$ | Nitosen－ |  |  | Noctel | Prosolous | Poassum | Seserium | Silion | Siver | Sodium | ${ }_{\text {Stronium }}$ | Suphate | Suphur |  | Tralium |  | Tin | Tratium |  | Uatium | varaium |  |  |
|  |  |  | Fraction | 015 | 015 | 015 | Dis | dis | is | ${ }^{015}$ | ois | 015 | Dis | ${ }^{015}$ | Dis | ${ }^{0}$ | Dis | dis |  | Dis |  | 015 | Dis |  | 015 | ${ }^{\text {ils }}$ | ${ }^{\text {dis }}$ | ${ }^{\text {ois }}$ |
|  |  |  |  | $\xrightarrow{\text { man }}$ | ${ }_{\text {mon }}$ | ${ }_{\text {mon }}^{\text {mos }}$ | ${ }_{\text {m91 }}$ | ${ }_{\text {m9 }}$ | －909 | $\stackrel{1}{4}$ | \％919 |  | ${ }_{\text {eng }}^{50}$ | ${ }_{\text {Hen }}^{\text {nen }}$ | ¢mı | ${ }_{\text {m9 }}^{0.05}$ | m92 | $\stackrel{10 n}{ }$ |  |  |  | ${ }_{\text {H1 }}^{102}$ |  |  |  |  |  | ${ }_{\text {m }}^{0}$ |
| $\frac{5}{55-1.1}$ | ss |  |  | 51772022 | Coara and colowres | ${ }_{0} 0.538$ | 0.015 |  | ${ }^{129}$ | 129 | ${ }_{0}^{0.128}$ |  | 0.298 | ${ }_{0} 0.935$ | 10000. |  | ${ }^{\text {4，}}$ | ${ }_{358}{ }^{35}$ | ${ }^{190}$ | ${ }_{53,4}$ |  |  |  | ${ }_{0}^{02}$ |  |  |  | ， | $\stackrel{0}{0.39}$ |  |
|  |  |  |  | ${ }^{9112022}$ | Ciear cooburess | 0.745 | ${ }^{0.0015}$ | ${ }_{0}^{0.0065}$ | ${ }_{5}^{588}$ | ${ }_{5}^{588}$ | －0．138 | ${ }_{4}^{54}$ | ${ }_{0}^{0.368}$ | ${ }_{0}^{0.004}$ | ${ }_{8}^{8810}$ | ${ }^{0.0005}$ | ${ }_{4}^{4.6}$ |  | ${ }^{130}$ | ${ }_{\substack{4,5 \\ 568}}$ |  | ${ }^{0.002}$ |  | 0.2 | ${ }_{0}^{0.5}$ |  | 0.395 | 0.44 | 1.05 | 0.1 |
|  | ss |  | ${ }_{\substack{1262022 \\ 3 \\ 3 / 7 \text { P20 }}}$ | Claarand coouress | 0．642 | ${ }_{0}^{0.0015} 0$ | ${ }_{\substack{0.006 \\ 0.006}}^{0.0}$ |  | ${ }_{\substack{4.84 \\ 3.54}}$ | ${ }_{\text {O．}}^{0.349} 0$ |  | － | 0.44 <br> 1. |  | ${ }^{0.0005}$ | ${ }_{4}^{4.17} 4$ | （tar | ${ }_{\substack{190 \\ 100}}$ | ${ }_{\substack{56.6 \\ 478}}^{5}$ |  | ${ }_{\text {O．002 }}^{0.002}$ |  | 02 <br> 0.2 | 0.5 <br> 0.5 |  | （e．28 | － | O．8 <br> 1.14 | 0.1 <br> 0.1 <br>  |
| 551．2 | ss |  | ${ }^{51772022}$ | Ciearano coloures | 0.422 | 0.015 | 0.0069 | ${ }_{0} 0.49$ | ${ }_{0}^{0.496}$ | ${ }^{0.338}$ | ${ }_{5}^{5 .}$ | ${ }^{0.185}$ | ${ }^{0.186}$ | ${ }_{8}^{8130}$ | 0.005 | ${ }^{2.9}$ | ${ }^{104}$ | ${ }^{53}$. | 15.9 |  | 0.002 |  |  | 0.5 |  | 0.0628 |  |  |  |
|  |  |  | ${ }^{12112022}$ | Coarandoumosouses | ${ }_{0}$ | 0.015 | ${ }_{0}^{0.0064}$ | ${ }_{0.649} 0$ | ${ }_{0}^{0.655}$ | ${ }_{1035}$ | ${ }_{46}^{4.6}$ | ${ }_{0}{ }_{0}$ | 029 |  | 0．005 |  |  | ${ }^{85}$ |  |  | O．002 |  |  |  |  | ${ }_{0} 0$ | ${ }^{0.25}$ | ${ }^{03}$ |  |
| Sti－2 | ${ }^{35}$ |  | ${ }^{31462023}$ | Arand count | $\stackrel{0.039}{0.093}$ | 0.015 | ${ }_{0} 0.005$ | ${ }_{0} 0.212$ | ${ }_{0}^{0.212}$ | 0.505 | ${ }_{2}$ | ${ }_{0}^{0.91}$ | 0.221 | ${ }_{820 .} 8$. | 0.005 | $\stackrel{3.12}{ }$ | ${ }_{114}^{14}$ | ${ }_{54}$ | ${ }_{\substack{28,8 \\ 18.8}}$ |  | 0.002 |  | 0.2 | ${ }^{0.5}$ |  | ${ }_{0}^{0.0827}$ | ${ }_{0} 029$ | ${ }_{0}^{0.42}$ | 0.1 |
| 27－1．2 | ¢ |  | ${ }^{1214242022}$ | No sampe，weolosatioed |  |  |  |  |  |  |  |  |  |  |  |  | $\stackrel{10}{10}$ |  | $\stackrel{14}{14.4}$ |  |  |  |  | $\stackrel{-7}{0.5}$ |  | ${ }_{0}^{0.0871}$ |  |  |  |
|  | ${ }_{\text {ss }}$ |  |  | Sighy yubas．sighy grey | ${ }^{\text {O．362 }}$ | ${ }_{0}^{0.042}$ | ${ }_{0}^{0.0146}$ | ${ }_{0}^{0.02}$ | ${ }_{0}^{0.023}$ | ${ }_{0}^{0.262}$ | ${ }_{7}^{4.8}$ | ${ }_{0}^{0.1215}$ | ${ }_{0}^{0.04}$ |  | ${ }_{0}^{0.0005}$ | ${ }_{\text {¢ }}$ | ${ }_{10}^{115}$ | ${ }_{4}{ }_{4}$ |  |  | ${ }_{0}^{0.002}$ |  | ， | O． |  |  | ${ }_{0}^{0.45}$ |  | 0.1 |
| 27－1－1 | ss |  | ${ }^{31442023}$ | Clara and cooviless | －0．41 | 0.015 | ${ }_{0}^{0.047}$ | 0.02 | 0.02 | ${ }_{0} 0.23$ | 28. | 0.2 | 0.58 | ${ }^{174700 .}$ | ${ }_{0} 0.01$ | ${ }_{123}$ | ${ }^{106}$ | ${ }_{46}$ | ${ }^{14 .}$ |  | 0.026 |  | ${ }^{0.4}$ | 1.2 |  | 0.198 | ${ }_{0} 0.79$ | 8， 88 | 0.2 |
| 退 $27-1.2$ | ¢ |  |  | Wert |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  | ${ }_{5}^{51312022}$ | Coar and coobiness | 0.358 | 0.015 | 0.005 | ${ }^{258}$ | ${ }^{258}$ | $\stackrel{0.488}{ }$ | 29 | 0．357 | $\stackrel{0.163}{ }$ | ${ }^{580}$. | 0.005 | ${ }^{28.5}$ | ${ }^{192}$ | 40. | 12 |  | 0.0025 |  | 0.2 | 0.5 |  | $\stackrel{0.088}{ }$ | ${ }^{0.2}$ | 0.26 | 0.1 |
|  | ${ }_{\text {cis }}^{\text {ss }}$ | r |  | Feragh | 0.321 | 0.015 | 0.005 | 225 | 225 | 0.474 | ${ }_{3}$ | 0.422 | 0288 | 6790． | 0.005 | ${ }^{31}$ | ${ }^{240}$ ． | ${ }_{56}$ | 19.4 |  | 0.0046 |  | 0.21 | 0.5 |  | 0.0481 | 0.21 | 0.4 | ${ }_{0}^{0.1}$ |
| 29，－2 | ${ }_{\text {ss }}^{\text {ss }}$ | $r$ | ${ }^{21272023}$ | Claerand colounss | ${ }_{0} 0.39$ | 0.015 | 0.005 | ${ }_{0} 0.903$ | ${ }_{0} 0.003$ | ${ }_{0.458}$ |  |  |  | ${ }_{6490}$ | ${ }_{0}^{0.005}$ | ${ }^{29,6}$ | ${ }_{212}^{212}$ |  | ${ }^{14,5}$ |  |  |  |  | ${ }_{0}^{0.5}$ |  | ${ }_{0.0879}$ | 0.2 |  |  |
|  | ss | $r$ |  |  |  | 0.015 |  |  |  |  |  | 0.96 | 0.005 | 976． |  | ${ }^{3.69}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | ${ }_{\text {ss }}^{\text {ss }}$ | $\stackrel{r}{r}$ |  |  | ${ }^{1.1 .1}$ | ${ }_{0}^{0.0015}$ | ${ }_{0}^{0.005}$ | ${ }^{0.544}$ | ${ }_{0}^{0.544}$ | ${ }_{\text {O．}}^{0.342}$ | ${ }^{32}$ | ${ }_{0}^{0.217}$ | ${ }_{\text {coin }}^{0.146}$ |  | ${ }_{0}^{0.005}$ | ${ }_{\substack{3,3 \\ 3,3}}$ |  | ${ }^{39}$ | ${ }^{125} 8$ |  | ${ }^{0.002}$ |  | ${ }_{0}^{02}$ | ${ }_{0}^{0.5}$ |  | ${ }_{0}^{0.242}$ | ${ }_{0}^{0.88}$ | ${ }_{0}^{0.95}$ | ${ }_{0}^{0.1}$ |
|  |  | $r$ | ${ }^{31414223}$ | Cliarand colouness | 0.947 |  |  |  |  |  |  |  |  |  | 0.0065 |  | ${ }^{334}$. |  |  |  |  |  |  | 1.04 |  |  |  |  |  |
|  |  | $r$ |  |  | － | ${ }_{0}^{0.02}$ | ${ }^{\text {O．O．005 }}$ |  |  | － | ${ }_{8565}$ | － |  |  | ${ }^{\text {O．005 }}$ | ${ }_{23,}^{20,}$ |  | ${ }_{\substack{31.5 \\ 51.5}}$ | ${ }^{14 .}$ |  | ${ }_{\text {－}}^{0.002}$ |  | ${ }_{0}^{0.255}$ | ${ }_{\text {－}}^{0.5}$ |  | ${ }^{0.0059}$ | ${ }^{0.46}$ | － <br> $\substack{0.23 \\ 0.35}$ | ${ }_{0}^{0.1}$ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.0894 |  |  |  |
| $\frac{20.1}{30 \cdot 11}$ | ¢ss | r |  | Clarand cooutes | ， | ${ }_{0}^{0.0015}$ | ${ }^{\text {O．005 }}$ |  | － | ${ }^{0.5235}$ | ${ }_{21,6}^{21.6}$ | ${ }_{\text {O．}}^{0.096}$ | ${ }^{\frac{0}{0.152}}$ |  | ${ }_{0}^{0.005}$ | ${ }_{\text {20，5 }}^{20.5}$ | ${ }_{\text {cose }}^{\text {238．}}$ | ${ }^{29}$ | ${ }^{12.4}$ |  | ${ }^{\text {O．002 }} 0$ |  | ${ }_{0}^{02}$ | ${ }_{\text {3，}}^{0.5}$ |  | －0．0．43 | ${ }^{\frac{0}{0.39}}$ | － | ${ }_{0}^{0.13}$ |
|  |  | r |  | Claer and colouress． | ${ }^{1.26}$ | 0.015 | 0.0371 | 0.02 | 0.046 | 0.307 | ${ }_{8.1}$ | 0.637 | 0.04 | 7680. | 0.005 | 18.6 | ${ }^{486}$ ． | ${ }^{26}$ | 8.9 |  | 0.0039 |  | 0.2 |  |  | 0.335 |  | 0.4 |  |
| （eald | ciss | $r$ | 隹 | 隹 | $\stackrel{1.12}{1.12}$ | 0.015 | 0.009 | 0.02 | ${ }_{0}^{0.023}$ | 0.7215 | ${ }_{9} 9$ | ${ }_{0}^{0.583}$ | 0.04 | 8080. | ${ }^{0.005}$ | 19，4 | $5{ }_{531}$ | $3{ }^{30}$ | ${ }_{9}{ }_{9}$ |  | 0.0026 |  | 0.2 | ${ }_{0}^{0.5}$ |  | ${ }_{0}^{0.321}$ | 0.21 | $\stackrel{0.84}{0.8}$ | 0.1 |
| － | ¢ss | $r$ |  | Somer |  | ${ }_{\text {a }}^{0.0015}$ | ${ }_{0}^{0.005}$ | ${ }^{\text {i．}}$ |  | （0．008 | ${ }^{71.2}$ | （1．62 | ${ }_{\text {0．0．07 }}^{0.0}$ |  | ${ }^{0.0005}$ | $\stackrel{\substack{352 \\ 29.6}}{ }$ |  | 30 <br> 34. <br> 4. | ${ }^{\frac{9}{13,6}}$ |  | 0．002 |  | －${ }_{0}^{0.2}$ | ${ }_{0}^{0.5}$ |  | ${ }_{\substack{0.0099 \\ 0.35}}$ | 0.2 <br> 0.2 |  | ${ }_{0}^{0.1}$ |
| 0.12 |  |  |  | Heas Seange did not have access |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{31-1-2}$ | ¢ | r | ${ }^{\frac{2152023}{5162022}}$ | Claarancoloultes | ${ }_{0}^{0.503}$ | 0．0．15 | ${ }^{\text {O．005 }}$ | ${ }_{\text {2，}}^{0.36}$ | ${ }_{\text {206 }}^{\text {2092 }}$ | － | ${ }^{79}$ | ${ }_{\text {\％}}^{\text {¢ }}$ | $\stackrel{0.057}{0.057}$ | ${ }_{5} 5450$ | 0.0 | ${ }_{\text {er }}^{\text {20，}}$ | －${ }^{3122}$ | ${ }^{20}$ | ${ }^{8}$ |  | ${ }_{0}^{0.0026}$ |  | ， | ， |  | $\stackrel{0}{0.459}$ | 0.0 |  |  |
|  |  |  |  | Clear and colouness． |  |  |  |  |  |  |  | 0.936 | 0.06 |  |  | ${ }_{648}^{648}$ | ${ }^{266 .}$ |  | ${ }^{612}$ |  |  |  |  |  |  | ${ }^{0.378}$ |  | ${ }^{1.63}$ |  |
| $\frac{3}{31-2-1}$ | ${ }_{\text {ckis }}^{\text {SRM }}$ | $r$ | ${ }^{\frac{12142022}{}}$ | Cearan cooutes | ${ }^{0.593}$ | ${ }_{0}^{0.0015}$ | ${ }^{\text {O．005 }}$ | ${ }_{\text {\％}}^{\text {3．395 }}$ | ${ }^{\text {c．}}$ | ${ }^{0.122}$ | ${ }_{5}^{245}$ | ${ }_{\text {O．}}^{0.647}$ | ${ }^{\frac{0.255}{0.045}}$ |  | ${ }_{0}^{0.005}$ | ${ }^{6.83}$ | ${ }_{3}{ }_{3625}^{362}$ | （20． | ${ }^{\frac{8986}{6485}}$ |  | ${ }^{\text {O．O．02 }}$ |  | ${ }_{0}^{02}$ | ${ }_{\text {O．5 }}^{0.5}$ |  | ${ }^{\text {O．528 }}$ | $\stackrel{\text { O．45 }}{0.25}$ | ${ }_{0} 0295$ | ${ }_{0}^{0.1}$ |
|  |  |  | ${ }^{921212022}$ | Meno fotupiciates | 0.9755 | 0.015 | 0.005 | ${ }_{0}^{02345}$ | 0.2345 | ${ }^{0.233}$ | 4.05 | 0.8725 |  |  | 0.005 | ${ }^{6.27}$ |  | ${ }^{1955}$ |  |  |  |  |  |  |  | 0.326 |  |  |  |
|  | ¢ |  |  | Nean ofuplatas | O． 0.85 | ${ }_{0}^{0.16}$ | ${ }_{\text {O．005 }}^{0.05}$ | － | $\xrightarrow{\text { L．} 0.95}$ | ${ }_{\substack{0.365 \\ 4.22}}^{0.20}$ |  | ${ }^{\text {O．} 0.63}$ | ${ }_{\text {O．}}^{0.104}$ |  | ${ }^{0.0055}$ | ${ }_{\text {¢ }}^{6.05}$ |  | $\stackrel{\substack{200 \\ 100}}{ }$ | ${ }^{\frac{1825}{432}}$ |  | $\stackrel{\text { O．002 }}{0.009}$ |  | $\stackrel{02}{0.2}$ | $\stackrel{0.5}{0.5}$ |  | ${ }^{0.3526}$ | －0．37 | ${ }_{0}^{0.057}$ | ${ }_{0}^{0.1}$ |
|  |  |  |  | Coar，comuness | ${ }^{0.236}$ |  |  |  |  | 0.027 | ${ }^{7} 9$ | ${ }_{0}^{0.784}$ |  | ${ }^{93360 .}$ | ${ }^{0.0005}$ |  | ${ }^{433}$ |  | ${ }^{71,6}$ |  |  |  |  |  |  | ${ }^{0.306}$ | 0.26 |  |  |
|  | ss <br> ss <br> ss |  |  |  | ${ }^{0.328}$ | －0．015 | －0．0．059 | ${ }^{0.542}$ | －${ }_{\text {O．942 }}^{0.542}$ | （e．as | ${ }^{\text {9，2 }}$ | $\xrightarrow{0.04}$ | ${ }^{0.002}$ |  | ${ }^{\text {cous }}$ |  | ${ }_{4}^{450}$ | ${ }^{200}{ }^{200}$ |  |  |  |  | ${ }_{0}^{02}$ | ${ }_{\substack{1.05 \\ 0.5}}^{\text {¢，}}$ |  |  |  |  | 0.1 |
|  | ¢ |  |  |  | － | （0．018 | ${ }^{\text {O．005 }}$ | （o．099 |  | － $\begin{aligned} & \text { O．378 } \\ & 3.02\end{aligned}$ | $\frac{8.9}{5.8}$ | ${ }_{\text {O．}}^{0.88}$ | ${ }_{\text {coiol }}^{0.092}$ |  | ${ }_{\text {O．005 }}^{0.005}$ |  |  |  |  |  | ${ }^{\text {O．0057 }}$ |  | ${ }^{0.2}$ | 0．5 <br> 0.5 <br> 0 |  | －0.324 <br> 0.651 | － 0.2 | ${ }^{\text {a，42 }}$ | ． |
| ${ }^{37.3 .1}$ | ${ }_{\text {crsm }}^{\text {crss }}$ |  | ${ }_{\text {l }}^{12882022}$ | Meanof fuplicies | ${ }^{0.0885}$ | ${ }_{0}^{0.45}$ | ${ }^{0.0055}$ | － 0.0405 | ${ }^{0.0405}$ | ${ }^{3.36}$ | ${ }_{44}^{44}$ | ${ }^{0.0225}{ }_{0}^{0.48}$ | ${ }^{0.04}$ | ${ }_{\substack{11350 \\ 10500}}^{\substack{1000}}$ | ${ }^{0.0005}$ | ${ }_{5}^{6.99}$ | ${ }_{\text {203 }}^{203}$ | ${ }^{205}$ | ${ }_{\substack{6335 \\ 308}}^{\substack{\text { che }}}$ |  | ${ }^{0.001315}$ |  | 0. | 0．5 |  | ${ }_{0}^{0.517}$ | ${ }_{0}^{0.325}$ | ${ }_{0}^{0.995}$ | 0.1 |
| ${ }^{388.1}$ | ${ }_{\text {ss }}^{\text {ss }}$ |  | ${ }^{511832022}$ | Ciearand cooumess | ${ }^{289}$ | ${ }^{0.015}$ | 0.005 | ${ }^{0.388}$ | ${ }^{0.388}$ | ${ }^{4.73}$ | ${ }_{4}^{4.5}$ | ${ }^{0.364}$ | ${ }^{0.435}$ |  | ＜ 0.005 | ${ }^{126}$ | ${ }_{654} 5$ | ${ }^{50}$ | ${ }^{13,8}$ |  | 0.002 |  | 0.2 | 0.5 |  | ${ }^{0.566}$ | ${ }^{1.91}$ | ${ }^{1.17}$ | 0.1 |
| ${ }^{3}$ | ss |  |  | Tiarand olouress | ${ }_{\substack{2.97 \\ 1.97}}$ | ${ }_{0}^{0.0015}$ | ${ }_{0}^{0.005}$ | ${ }^{0.319} 0$ | －${ }_{0}^{0.319}$ |  | ${ }^{4.5}$ | ${ }_{\text {a }}^{0.373}$ | ${ }_{\text {a }}^{0.189}$ | ${ }_{\substack{15200 \\ 15000}}^{15}$ | ${ }_{0}^{0.0005}$ |  |  |  | ${ }_{\substack{13,5 \\ 13,4}}$ |  | ${ }_{\text {O．007 }}^{0.002}$ |  |  | 0.5 |  |  | ${ }^{212}$ |  |  |
| ${ }^{\frac{1.1}{1-1}}$ |  |  | ${ }^{31442023}$ | learand colurses | 1.66 <br> 24 |  |  |  | ${ }^{\text {0．4．89 }} 0$ | ${ }^{3.79}$ |  |  |  | ¢ | ${ }_{0}^{0.0053}$ |  | ${ }_{5}^{519}$ | ${ }^{4 .}$ | ${ }_{\text {l }}^{\text {11，3 }}$ |  | 0.002 |  | 02 | 0.79 |  | ${ }^{0.449}$ | 1.89 | 1.63 | 0.1 |
| 9，1－1 | ${ }_{\text {ss }}$ | $\stackrel{r}{r}$ | 91582022 | coara，colouness | ${ }^{248}$ | 0.015 | 0.005 | 0.078 | 0.078 | ${ }^{0.188}$ | ${ }_{2}$ | ${ }_{0}^{0285}$ | ${ }_{0}^{0.187}$ | ${ }_{\text {cke }}$ | ${ }^{\text {O．005 }}$ | ${ }_{3,22}$ | ${ }^{236}$ | ${ }^{24}$ | $\stackrel{\text { ¢ }}{7.1}$ |  | 0．002 |  | ${ }_{0}$ | ${ }^{0.5}$ |  | ${ }_{0}^{0.13}$ | ${ }_{1}^{1.71}$ | ${ }^{1.52}$ | 0.1 |
|  |  | $\stackrel{r}{r}$ | ${ }^{\text {chirl202 }}$ | Clear and cooures | 243 <br> 225 <br> 2 | ${ }^{0.015} 0$ | ${ }_{\substack{0.0226 \\ 0.3785}}^{0.0}$ |  |  |  | $\frac{3}{2}$ <br> 26 | ${ }_{\text {O．}}^{0.309}$ | ${ }_{\text {a }}^{0.123}$ 0．195 |  | ${ }^{\text {co．005 }} 0$ | ${ }_{\substack{3.98 \\ 4.515}}^{\text {a }}$ |  | ${ }^{26}{ }^{26}$ | 7．8 <br>  <br> 7.05 |  | ${ }_{\text {a }}^{0.002}$ |  | 0.2 <br> 0.2 | 0.5 <br> 0.5 |  | （e．138 | $\stackrel{209}{1.615}$ |  | 0.1 |
|  |  | $\stackrel{\text { r }}{ }$ | ${ }_{\substack{12272022}}^{\substack{302023}}$ | Ciear and coloumes | ${ }_{\text {¢ }}^{1.128}$ | ${ }_{0}^{0.015}$ | （0．005 | －0．094 | ${ }^{0.0984}$ | －0．112 |  | ${ }^{0.2}$ | ${ }^{0.007}$ | ${ }^{8440}$ | ${ }^{0.0055}$ |  | ${ }^{327}$ | ${ }^{26}$ | ${ }_{8}^{74}$ |  | ${ }^{0.0022}$ |  |  |  |  | ${ }^{0.107}$ |  |  |  |
| ${ }^{302 \cdot 2 \cdot 1}$ | ¢ ${ }_{\text {ss }}^{\text {ss }}$ | $r$ |  |  | ${ }^{1.1 .88}$ | ${ }^{1.3}$ | ${ }_{0}^{\text {O．O．032 }}$ |  | ${ }_{0}^{1.097}$ | ${ }_{0}^{0.087}$ | 2 | ${ }_{0}^{0.19}$ | ${ }^{0.1066}$ | $\stackrel{7}{7000}$ | ${ }^{0.005}$ | ${ }_{3,25}$ | ${ }_{264}$ | ${ }_{20}^{20}$ | $\stackrel{6}{5.9}$ |  | ${ }^{0.002}$ |  | 0.2 | ${ }^{0.5}$ |  | $\stackrel{0}{0.135}$ | ${ }_{1}^{1.7}$ | ${ }_{0.19}^{0.19}$ | 0.1 |
| ${ }^{392.2 .1}$ | ¢ |  |  | Cliear and coouriss | － 12.29 | －0．015 | ${ }^{\text {0．005 }} 0$ | ${ }_{0}^{0.071}$ |  | O．0．068 | $\stackrel{21}{21}$ | ${ }^{0.19}$ | ${ }^{0.2079} 0$ |  | ${ }^{\text {O．005 }}$ | ${ }_{\substack{3,39 \\ 935}}$ |  | ${ }_{\substack{20 \\ 32}}$ |  |  | ${ }^{0.002}$ |  | －02 | ${ }_{\substack{0.5 \\ 0.58}}$ |  | ${ }^{0.16}$ | ${ }^{\frac{1144}{08}}$ | $\frac{0.1}{1.41}$ |  |
| 20－1－1 | ${ }_{\text {ss }}$ |  | ${ }_{\text {g } 92222022}$ |  | ${ }^{1.09}$ | ${ }_{1}^{1.6}$ |  |  | ${ }^{1.41}$ |  | $\stackrel{436}{438}$ | ${ }^{3,7}$ |  |  | 0．005 |  | ${ }_{394} 3$ | ${ }_{45}^{45}$ |  |  |  |  |  |  |  | ${ }_{0}^{0.65}$ |  |  |  |
| 0 | ${ }_{\text {ss }}$ |  | 1214142022 | Ciaran andolounses | ${ }^{1.16}$ | 0.95 | 0.0284 | ${ }^{1.48}$ | ${ }^{1.51}$ | 1.05 |  | ${ }_{385}$ | 0.05 | ${ }_{8390} 8$. |  | ${ }^{10.3}$ | ${ }_{\text {sor }}$ | ${ }_{56}{ }^{\text {b }}$ | ${ }^{19,1}$ |  | 0.0066 |  | 0.2 | 0.5 |  | ${ }^{0.8}$ | 0.35 | 229 | 0.1 |
|  | $\stackrel{\substack{\text { spm } \\ \text { FRM } \\ \hline}}{ }$ | $r$ |  | Cearan coloures | $\begin{array}{r}1.98 \\ \hline 1.98 \\ \hline\end{array}$ | ¢ 0.03 <br> 0.03 | ${ }_{\substack{0.0 .055 \\ 0.005}}$ |  | $\stackrel{\text { ¢ }}{ } \stackrel{6.17}{ }$ | O． | ${ }^{2,285}$ |  | ${ }_{0}^{0.04}$ |  | ${ }_{0}^{0.0055}$ | － |  | ${ }^{\text {47．5 }}$ | $\stackrel{13,9}{13,9}$ |  | ${ }^{0.0007}$ |  | ${ }^{0.2}$ | ${ }_{\text {－}}^{0.5}$ |  | ${ }^{0.173}$ | ${ }_{0}^{0.05}$ | coid <br> 0.39 <br> 0.39 | 0 |
| $\frac{4.1 .1}{41-1.1}$ | ${ }_{\substack{\text { rem } \\ \text { FRM }}}$ | $\stackrel{r}{r}$ | （1922022 | Meano otuplialas | ${ }_{2}^{2415}$ | ${ }^{0.0525}$ | ${ }_{0}^{0.0005}$ | ${ }^{\text {O．0445 }}$ | ${ }_{\substack{0.0445 \\ 0.035}}$ |  | 22 <br> 265 <br> 28 | ${ }^{1.951}$ | ${ }_{\text {coiob }}^{0.0065}$ |  | ${ }^{0.0005}$ | ${ }_{\text {L，}}^{\text {4，995 }}$ | ${ }^{191}$ | ${ }^{56}{ }^{56}$ | ${ }_{\substack{17,95 \\ 17.95}}^{\text {c，}}$ |  | ${ }_{\text {0．002 }}^{0.005}$ |  | －${ }_{0}^{0.2}$ | ${ }^{0.5}$ |  | ${ }_{\substack{0.261 \\ 0.268}}^{0.20}$ | ${ }^{0.2}$ | ${ }_{0}^{0.315}$ |  |
| 401－1 | ${ }_{\text {fr m }}$ | $r$ | ${ }^{22712023}$ | veano foturicates | ${ }^{12955}$ | 0.0215 | 0.00595 | 0.3635 |  |  |  | 0.86 | 0.0025 | ${ }_{8160} 8$. | 0.005 | 4.975 | ${ }^{189,5}$ | ${ }^{59}$ | ${ }^{16,5}$ |  |  |  |  |  |  |  |  |  |  |
| ${ }^{2 \times 2.1}$ | ${ }_{\text {cren }}^{\text {ren }}$ | $\stackrel{\text { r }}{\text { r }}$ |  | Naenotaplatas |  | 0.0076 | ${ }^{\text {O．005 }}$ | ${ }_{0}^{0.02}$ | 0.02 <br> 0.02 |  | －${ }^{512}$ | －${ }_{\text {0．906 }}^{1.16}$ | ${ }^{0.2096}$ |  | ${ }^{\text {O．005 }}$ | － |  | ${ }_{86}{ }_{86}$ | －${ }_{\text {20，9 }}^{20.9}$ |  | 0．002 |  | ${ }^{02}$ | ${ }^{0.1}{ }_{1}^{0.72}$ |  | －0．0．05 |  | － | 01 |
| ${ }^{\frac{424-1}{42,-1}}$ | ${ }_{\text {Fer }}^{\text {Rem }}$ |  | ${ }^{12175252022}$ | Meanot tuplcates | ${ }^{0.236}$ | ${ }_{0}^{0.0745}$ | 0.005 | 0.02 | 0.02 | 0.1435 | ${ }^{9.15}$ | ${ }_{0}^{0.897}$ | 0.04 | ${ }^{7775}$ | ${ }^{0.005}$ | ${ }_{\text {O，} 9.55}$ | ${ }_{\text {a }}^{\text {435．}}$ | ${ }^{78.5}$ | ${ }_{20.05}$ |  | 0.002 |  | 0.2 | 0.785 |  | 0.0683 | ${ }^{1.105}$ | ${ }_{0}^{0.38}$ |  |
|  | ${ }_{\text {sfs }}^{\text {ss }}$ |  | ${ }^{2 \text { 2772023 }}$ | Hean ofupleates | － | ${ }^{0.0075}$ | ${ }_{0}^{0.0057}$ | ${ }_{0}{ }_{0}^{0.029}$ | ${ }_{0}^{0.0025}$ | ${ }^{\text {c．as8 }}$ | ${ }^{5} 5$ | ${ }_{\substack{0.8775 \\ 0.429}}^{\text {a }}$ | ${ }_{\text {O }}^{0.04}$ |  | ${ }_{0}^{0.005}$ | ${ }_{\text {8．945 }}^{9.967}$ | ${ }_{\text {3 }}^{\text {3915 }}$ 465． | ${ }_{\text {liob }}^{80.5}$ |  |  | ${ }^{\text {0．0．022 }}$ |  | 02 | 0.5 <br> 0.5 <br> 0 |  | ${ }^{0.0 .492} 0$ | ${ }_{\text {O．76 }}^{0.01}$ | ${ }_{\substack{0.18 \\ 1.99}}^{1}$ |  |
| ${ }^{\frac{33-1.1}{43-1.1}}$ | ss |  | 9812022 | Ciarar columess | －0．82 | ${ }^{0.0015}$ | ${ }^{0.005}$ | －0．071 | ${ }_{0}^{0.0071}$ | ${ }^{0.064}$ | －${ }^{36}$ | ${ }_{\text {O．481 }}^{0.465}$ | ${ }^{0.076}$ | ${ }_{\text {a }}^{\text {9810 }}$ | ${ }^{0.005}$ | ${ }_{\substack{8,74 \\ 888}}^{\text {c，}}$ | ${ }_{\text {che }}^{4.45}$ | ${ }^{21}$ | ${ }_{\substack{2,8 \\ 24}}$ |  | 0．002 |  | ${ }_{0}^{02}$ | ${ }_{0}^{0.5}$ |  | ${ }^{0.1988}$ | ${ }_{1}^{1.46}$ |  |  |
| 星 | ${ }_{\text {ss }}^{58}$ |  | ${ }_{\text {322023 }}$ | 隹 | ${ }_{\substack{0.821}}^{0.092}$ |  |  |  |  | ${ }_{0}^{0.718}$ | 6. |  |  |  | 0.0071 |  | ${ }_{497}^{498}$ | ${ }_{72}$ |  |  |  |  |  |  |  | ${ }^{0.146}$ | ${ }_{1}^{1.35}$ | ${ }^{1.20}{ }^{1.27}$ | 1 |
| $\frac{44+1}{44+1-1}$ | ¢ |  |  | Cotarand coluress | － | ${ }^{0.0015}$ | ${ }^{0.005}$ | ${ }_{\text {O }}^{0.021}$ | ${ }_{0}^{0.021}$ | －0．399 | \％${ }^{8.6}$ | ${ }^{0.3037}$ | ${ }_{0}^{0.044}$ |  | ${ }^{0.005}$ | ${ }^{5} 5$ |  | ${ }^{36}$ |  |  | 0．002 |  | ${ }^{0.2}$ |  |  | （0，188 | － |  |  |
|  | ss |  | ${ }^{126820222}$ | diearand colousess | ${ }^{1.03}$ | 0.015 | 0.005 | $\stackrel{0}{0.02}$ | 0.02 | 0.529 | ${ }_{9} 9$ | 0.389 | ${ }_{0}^{0.04}$ |  | 0．005 | ${ }_{6}^{6.27}$ | ${ }^{2055}$ | ${ }^{30}$ |  |  | 0.002 |  |  | 0.5 |  | ${ }^{0.174}$ | ${ }_{0}^{0.35}$ |  |  |
| 退 | －${ }_{\text {ss }}^{\text {ss }}$ |  | ${ }_{5}^{51929022}$ | Niarar fooluess | ${ }^{0.095}$ | 0．015 | 0．005 | ${ }_{0}^{0.851}$ | ${ }_{0}^{0.851}$ | ${ }_{0}^{0.305}$ | ${ }^{46}$ |  |  |  |  | $\stackrel{\text { a，}}{3,32}$ | ${ }^{203}{ }^{233}$ |  |  |  | ${ }^{0.0028}$ |  |  |  |  |  | ${ }_{0}^{0.23}$ |  |  |
| ${ }^{\frac{5151-1}{51-1 / 1}}$ | ${ }_{\text {ss }}^{\text {ss }}$ |  |  | coind | －0．189 | ¢ 0.015 | －${ }_{5}^{0.0005}$ | ${ }_{\substack{0.221 \\ 2.74}}$ |  | － | 24 | ${ }^{0.396}$ | ${ }_{0}^{0.04}$ | ${ }_{\substack{47200 \\ 6080}}$ | －${ }_{\substack{0.005 \\ 0.005}}$ | ${ }^{3.5}$ | ${ }_{\substack{174 \\ 188 .}}^{\text {18，}}$ | 40. | ${ }_{128}^{12.1}$ |  | ${ }_{\substack{0.0027 \\ 0.022}}^{\substack{0}}$ |  | O2 | ＜ 0.5 |  |  | 0．3 | （0．72 | ： 0.1 |




| ${ }_{\text {sc css }}$ |  |  |  | ${ }_{\text {atem }}^{\text {atw } \text { aximum (1) }}$ | ${ }_{\substack{1000 \\ 250}}$ | 1.31-18.43) | ${ }^{0.22(4)}$ | 400 10 | ${ }^{400}$ | $\underbrace{250-1500(5)}$ |  |  | ${ }_{20}^{20}$ |  | ${ }^{0.5 .515(5)}$ | 1700(9) | 2500 | ${ }^{128.429} 5$ |  |  | 3 |  | 25 | 1000 | ${ }^{85}$ | ${ }^{20}$ | ${ }_{\substack{15.2900 \\ 3000}}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Station | Sample Type | (camplane ( | ${ }^{\text {Date Sampled }}$ | Parameer | m | $\underset{\substack{\text { Nitroen- } \\ \text { ammonal }}}{\text { and }}$ |  | $\substack{\text { Nirosen- } \\ \text { nitate }}$ | Nitrogen - nitrate plus | Niokel | Prospows | ${ }_{\text {Potassum }}$ | Seanium | Sillon | ${ }_{\text {Siver }}$ | Sodium | Stronium | ate | rr |  | Tralum |  | Tn | Tranum | UTasum | m | zno | Erame |
|  |  |  |  | Fraction | dis | 015 | 015 | 015 | dis | 015 | ${ }^{1 / 5}$ | ${ }^{\text {dis }}$ | ${ }^{0} 15$ | ${ }^{\text {dis }}$ | Dis | Dis | 015 | 015 | dis |  | 015 |  | 015 | 015 | 015 | 015 | Dis | 015 |
|  |  |  |  |  | $\xrightarrow{\text { men }}$ | ${ }_{\substack{\text { mgn } \\ 0.015}}$ | ${ }_{\substack{\text { m91 } \\ 0.005}}$ | ${ }_{0}^{\text {m91 }}$ | ${ }_{\text {mgn }}^{\substack{\text { ma }}}$ | $\xrightarrow{\text { ne9 }}$ | $\stackrel{\text { ren }}{2}$ |  | $\xrightarrow{\text { min }}$ | ${ }_{\text {men }}^{50}$ |  | ${ }_{\substack{\text { mgh } \\ 0.05}}$ |  | m92 | $\stackrel{\text { m }}{3}$ |  |  |  | $\stackrel{\text { m91 }}{0.2}$ | ${ }_{\text {mel }}^{0.5}$ |  | ${ }_{0.2}$ | ${ }_{0}^{19.1}$ | ${ }_{\text {men }}^{0.1}$ |
| 106 | ss |  | 9112022 |  | - |  |  | -0.0.9 |  |  | ${ }^{\text {20.9 }}$ |  |  |  |  |  |  | ${ }^{250}$ | ${ }_{\text {74.2 }}$ |  | ${ }^{0.002}$ |  | 0.2 <br> 0.93 | $\stackrel{\text { 2. }}{2.12}$ |  | $\stackrel{1.22}{1.22}$ |  | O.44 |
| (1007-1.1 | ss |  |  |  | ${ }_{\substack{127 \\ 5.71}}^{10}$ | 2.8 1.6 | ${ }_{0}^{0.1083} 0$ |  |  | ${ }_{\text {2 }}^{232} 4$ | ( $\begin{array}{r}37.9 \\ 76.9\end{array}$ | ${ }_{2}^{2,15}$ | ${ }_{\text {coind }}^{0.198}$ |  | ${ }_{\substack{0.005 \\ 0.0061}}^{\substack{\text { O. }}}$ |  |  | ${ }_{\substack{100 \\ 200}}^{1}$ |  |  | $\underbrace{0.002}_{0}$ |  | - ${ }_{0}^{0.2}$ | 0.5 <br> 322 | (0.924 | ${ }_{\substack{0.72 \\ 9.25}}$ | $\underset{\substack{0.04 \\ 6.07}}{ }$ | 0.1 0.19 |
|  | ¢ss |  | ${ }_{\text {a }}^{\text {9912022 }}$ | Veen utidisisibiy gey | (1928 | ${ }^{0.1}$ | - | ${ }_{\substack{23.9 \\ 9.07}}$ |  | -3, <br> 215 <br> 15 | 79.9. <br> 10. | 1.21 <br> 1.04 <br> 1 | ${ }_{\text {O }}^{0.225}$ | ${ }_{\text {cke }}^{5680}$ | $\underbrace{0.005}_{0}$ | ${ }_{\substack{10.1 \\ 6.07}}^{1}$ | cois | ${ }_{\substack{20 . \\ 100}}^{\substack{100}}$ | ${ }_{\substack{854 \\ 58.8}}^{\substack{\text { che }}}$ |  | ${ }^{0.0 .118} 0$ |  | - | ${ }_{\substack{0.95 \\ 0.65}}$ | (i.0.58 | ${ }_{\text {0. }}^{0.89}$ |  | 0.1 |
|  | ¢ |  |  |  | $\stackrel{110.5}{11.5}$ | $\stackrel{0}{0.02}$ | $\stackrel{\text { O. }}{0}$ | $O$ | $\stackrel{0}{0.0}$ |  | $\stackrel{+}{7.4}$ | $\stackrel{1.04}{1.14}$ | - |  | 0 | $\stackrel{.0}{14.5}$ | ${ }_{788} 7$ | ${ }_{200}^{200 .}$ | $\stackrel{56.8}{64.1}$ |  | 0008 |  |  | -0. | $\stackrel{.5}{0.512}$ | - | ${ }_{0}^{0.35}$ | $\frac{-1}{0.1}$ |
| 1089,-1 | ${ }_{\text {ss }}^{\text {ss }}$ |  | ${ }^{12772022}$ | Sinhtury ubis, isionly yey | ${ }^{14.1}$ | ${ }_{0} 0.025$ | ${ }_{0}^{0.0098}$ | ${ }_{0} 0.02$ | 0.02 | ${ }_{1}^{1.35}$ | ${ }_{125}$ | ${ }_{1.14}$ | ${ }_{0} 0.102$ | ${ }_{1}{ }^{12800}$ | 0.005 | ${ }^{11.3}$ | ${ }_{\text {of }}$ | ${ }_{200}^{20 .}$ | ${ }_{74,1}$ |  | ${ }_{0}^{0.0252}$ |  | ${ }_{0.2}$ | ${ }^{0.5}$ | ${ }_{0}^{0.509}$ | ${ }_{0}^{0.53}$ | ${ }_{0.3}$ | ${ }_{0}^{0.1}$ |
| ${ }^{10.1}$ | ¢ |  |  | Nosemperequired the month | ${ }^{131}$ | 0.41 | 0.0223 | 0.044 | 0.066 | 0.889 | ${ }^{322}$ | ${ }^{124}$ | ${ }^{247}$ | ${ }^{11900}$. | 0.005 | ${ }^{85,9}$ | ${ }^{832}$ | ${ }^{76}$ | ${ }^{24.8}$ |  | ${ }_{0}^{0.0285}$ |  | 0.2 | ${ }_{0}^{0.5}$ | ${ }^{1.52}$ | 5.04 | 12 | ${ }_{0}^{0.1}$ |
| ${ }^{10971.1}$ | ss |  | ${ }_{11302022}$ | Moderatey ubtod, modealey yey | 160. | 0.38 | 0.024 | 0.029 | 0.053 | 0.439 | ${ }^{35,3}$ | 1.11 | 1.06 | 13300. | 0.005 | 10. | 109. | 89. | ${ }^{27.6}$ |  | 0.0039 |  | 0.2 | 0.5 | 236 | ${ }_{1} .5$ |  | 0.1 |
| $\frac{10.9}{\frac{10.7}{10.1}}$ | cis |  |  | No sample equered his mont | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  | - |  |  | - | - | - | - | - |
| $\frac{10,1-1}{1010}$ | Ns |  |  | Vosampe (eetemined dot obe erepsenenate) | - |  | - | - | - |  | - | - | - |  | $=$ |  |  |  |  |  | $=$ |  |  |  |  |  |  |  |
| P1 | ¢ |  |  |  | $\stackrel{\substack{\text { 14, } \\ 16.3}}{\square}$ | $\stackrel{-}{42}$ | ${ }_{0}^{0.005}$ | -0, | $\stackrel{-}{0.02}$ | $\stackrel{\square}{783}$ | $\stackrel{\substack{869 \\ 164}}{\square}$ | $\underset{\substack{376 \\ \text { S0. }}}{ }$ | $\stackrel{-08}{0.08}$ |  | O-01 | $\stackrel{\square}{120}$ |  | $\stackrel{-}{1}$ |  |  | -0.004 |  | $\underline{0.43}$ | ${ }_{2}^{1}$ | $\stackrel{-}{0.024}$ | $\stackrel{-6}{0 .}$ | $\xrightarrow{722}$ | - |
| P1 | ¢ |  |  | Clazarand columess | ${ }^{14.0}$ | ${ }_{72}$ | ${ }_{0}^{0.0071}$ | $\stackrel{0.029}{0.0}$ | ${ }_{0}^{0.036}$ | ${ }^{\text {14,6 }}$ |  | ${ }_{532}$ | ${ }_{0} 0.2$ | ${ }_{20}^{2000}$ | ${ }_{0}^{0.025}$ |  | $\stackrel{1000}{1000}$ | ${ }_{4}{ }^{1}$ | ${ }^{15} 5$ |  | 0.01 |  |  |  | ${ }^{0.033}$ |  |  |  |
|  | cos |  |  |  | - | ${ }^{23}$ | ${ }^{0.005}$ | -0.020 | ${ }^{0.02}$ |  |  | ${ }_{\text {20, }}^{20.4}$ | O.08 <br> 0.2 <br> 0.0 | 18200 <br> 11800 <br> 18. | ${ }^{0.015}$ | ${ }^{245}$ | $\stackrel{1060}{\substack{1022 \\ 902}}$ | $\frac{12}{1 .}$ | ${ }^{\frac{6}{15}}$ |  | 0.004 <br> 0.01 |  | O. | ${ }^{\frac{1}{25}}$ | ${ }^{0.0 .299} 0$ | 0.4 <br> ${ }_{2}^{22}$ | (in30.8 <br> 1.88 | $\begin{array}{r}0.2 \\ 0.5 \\ \hline\end{array}$ |
|  | ¢ss |  |  | Slighly whide, sloply yelow | -14.8 <br> 15.5 | - ${ }^{82}$ 92. | ${ }^{0.00159} 0$ | 0.069 0.02 0 | ${ }_{0}^{0.095}$ | (9,984 | ${ }^{53}$ | ${ }^{76}$ | - ${ }_{0}^{0.155}$ |  | ${ }_{\substack{0.025 \\ 0.005}}$ |  | (1210. | . | ${ }^{15}$ |  | ${ }_{0}^{0.00}$ |  | ${ }^{029}$ | ${ }^{2.5}$ | $\underbrace{0.0017}_{0}$ | ${ }_{2}^{24}$ |  | ${ }_{0}^{0.5}$ |
|  | ss |  | ${ }^{22424203}$ | vers sifth yelow, notutbidy | ${ }^{125}$ | ${ }^{13}$ | 0.005 | 0.02 | 0.02 | ${ }^{7,23}$ | ${ }_{227}$ | ${ }_{59,1}^{59 .}$ | 0.095 | ${ }^{11900}$. |  | ${ }_{1}^{199 .}$ | ${ }^{924}$ | 1. |  |  | 0.004 |  | 0.4 |  |  | ${ }^{227}$ | ${ }_{297}^{298}$ | ${ }_{0}^{0.28}$ |
|  | ¢ |  |  | Ceara and sighly orame | 19.9 <br> ${ }^{13,5}$ |  | ${ }^{\text {0.0.5 }} 0$ | ${ }^{\text {O. }}$ | $\stackrel{0.2}{0.174}$ | $\stackrel{10.9}{ }$ | ${ }^{41}$ |  | -0.2 | ${ }_{\text {l }}^{147000}$ | -0.025 |  | ( ${ }_{\text {a }}^{\substack{\text { a37 } \\ 1000}}$ | 1 | ${ }_{\text {¢ }}^{15}$ |  | ${ }^{0.01}$ |  |  | ${ }^{2.5}$ | ${ }^{0.0 .13} 0$ | 2,9 <br> 1.9 <br> 2 | - | 0.5 <br> 0.5 |
| ${ }^{\text {P3 }}$ | ss |  | ${ }^{121712022}$ | Siphly ubded sifuly orang | ${ }_{143}^{14}$ | ${ }_{81}^{91}$ | ${ }^{0.0389}$ | ${ }^{0.022}$ | ${ }_{0}^{0.055}$ | ${ }^{7}{ }^{7,95}$ | (167 | ${ }_{\substack{75.6 \\ 699}}$ | ${ }^{0.16}$ | (13200. | ${ }_{\text {coiol }}^{0.005}$ | ${ }_{203}^{24 .}$ | (1010. | ${ }^{23}$ | ${ }^{3}$ |  | ${ }_{0}^{0.002}$ |  | -02 | ${ }_{0}^{0.5}$ | - ${ }_{\text {0.0.10 }}^{010}$ | - | 0.9 287 | 0.2 0.21 0 |
|  | ¢s |  | ${ }_{\text {S5262022 }}$ | Coara and moderately orange | ${ }^{14.7}$ | ${ }^{76}$ | 0.935 | ${ }^{0.95}$ | ${ }^{1.89}$ | ${ }_{7}^{7.73}$ | $\stackrel{247}{24}$ | ${ }_{64,5}$ | 0.15 | ${ }_{14100}$ | 0.01 | ${ }^{221}$ | ${ }^{\text {a26. }}$ |  | ${ }^{6}$ |  | 0.004 |  |  |  | 0.0124 | 202 <br>  <br>  <br>  <br> 28 | $\stackrel{1.17}{1.17}$ | ${ }_{0}^{0.2}$ |
| ${ }^{\text {P4 }}$ | ss |  | ${ }^{\text {9,4/42022 }}$ |  | ${ }^{16,4}$ | 7r | 0.0143 | ${ }^{0.033}$ | ${ }_{0}^{0.047}$ | ${ }_{8}^{8,1}$ | ${ }^{43}$ | ${ }_{60}^{60}$ | 0.2 | ${ }_{1}^{14300}$ | ${ }^{0.025}$ | ${ }^{224}$ | ${ }^{887}$ |  |  |  | ${ }^{0.01}$ |  |  | ${ }^{25}$ | 0.01 | ${ }^{1.9}$ | ${ }_{6}^{6.98}$ |  |
| ${ }^{24}$ | ¢ |  |  | cemen | ${ }^{14.1}$ | ${ }_{7}$ | -0.0567 | $\stackrel{0.025}{0.025}$ |  | ${ }_{\text {c, }}^{1.18}$ |  |  | - ${ }_{\text {O.146 }}^{0.17}$ |  | ${ }_{0}^{0.000}$ | ${ }_{213}^{213 .}$ |  | 1. | $\stackrel{3}{6}$ |  | ${ }^{\text {O.002 }}$ |  | - 0.24 | ${ }_{\text {a }}^{1.05}$ | (0.0175 | ${ }_{208}^{208}$ | +10.05 | ${ }_{0}^{0.2}$ |
| ${ }^{\text {P10 }}$ | ¢s |  |  |  | ${ }_{4}^{4.06}$ | ${ }_{\text {¢ }}^{\substack{33 \\ 47}}$ | ${ }^{0.0083}$ | (0.288 | (0.237 | ${ }_{6}^{4.05}$ | ${ }^{6.6}$ | ${ }_{4}^{29.9}$ | ${ }_{0}^{0.076}$ | ${ }_{\text {cose }}^{\substack{14700 \\ \hline 1400}}$ | ${ }^{0.005}$ |  | - |  |  |  | ${ }^{0.0022}$ |  | ${ }_{0}^{0.4}$ | $\begin{array}{r}0.5 \\ \\ \\ \hline 25\end{array}$ | ${ }^{0.0012}$ | 0.74 | ${ }_{\substack{201 \\ 426}}$ | ${ }_{0}^{0.1}$ |
| P10 | ¢s |  |  |  |  | ${ }_{5}^{4 .}$ | ${ }_{\substack{0}}^{\substack{0.0089}}$ | $\stackrel{\text { cose }}{0.024}$ | ${ }_{\text {N }}^{0.039}$ | ${ }_{6}^{6.687}$ | ${ }^{222}$ | ${ }_{4}^{4.7}$ | ${ }^{0.106}$ | ¢ | O.020 |  |  | . | ${ }_{6}{ }^{\text {c/ }}$ |  | ${ }_{0}^{0.0049}$ |  | ${ }^{0.4}$ |  | ${ }_{0}^{0.0166}$ | 0.9 | $\stackrel{4.105}{205}$ | - |
|  |  |  |  |  | ${ }^{\frac{3}{4} 4}$ | ${ }^{\text {33, }}$ | ${ }^{0.005}$ | (0.02 | ${ }^{0.02}$ |  | ${ }^{\frac{97}{112}}$ | ${ }^{27.9}$ | ${ }_{0}^{0.064}$ |  | ${ }^{0.0005}$ | ${ }_{8}^{842}$ |  |  | ${ }^{3}$ |  | 0.002 |  |  |  | ${ }^{0.00107}$ | ${ }_{0}^{0.77}$ | ${ }^{1.19}$ |  |
| P11 | ss |  | ${ }^{11129202022}$ | Vear and colountess | ${ }^{1.34}$ | ${ }_{7} 7.9$ | ${ }_{0}^{0.0565}$ | ${ }^{322}$ | ${ }^{32} 3$ | ${ }^{1.73}$ | ${ }_{6}^{6.3}$ | ${ }_{3.85}^{3.8}$ | 0.69 | ${ }_{6520 .}$ | ${ }_{0}^{0.005}$ | ${ }^{13.8}$ |  | ${ }^{190 .}$ | ${ }_{58,4}$ |  | ${ }_{0}^{0.0072}$ |  |  |  | ${ }_{0}^{0.202}$ |  | ${ }_{4}^{488}$ |  |
| ${ }^{\text {P11 }}$ | ss |  |  |  | ${ }_{\text {l }}^{1.162}$ | ${ }_{1}^{24}$ | ${ }^{0.0093}$ | ¢ | ${ }_{5}^{59.9}$ | ${ }_{\text {l }}^{1.295}$ | $\stackrel{1585}{\frac{15}{6.5}}$ | ${ }^{2445}$ | ${ }^{2.345}$ |  | ${ }_{0}^{0.005}$ | $\stackrel{10,75}{149}$ |  | ${ }_{\substack{565 \\ 10}}^{50}$ | ${ }_{\text {che }}^{174.5}$ |  | ${ }_{0}^{0.009} 0$ |  | ${ }_{0}^{0.2}$ | 0.5 | ${ }_{0}^{0.773}$ | ${ }_{0}^{0.575}$ | (incti | ${ }^{0.2}$ |
| P12 | ${ }^{\text {Ns }}$ |  | 9112022 | sampe well wasady. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| P12 | ${ }_{\text {Ns }}^{\text {ss }}$ |  | ${ }^{\text {che }}$ |  | 0.972 | 04 | 0.0 | ${ }^{274}$ | ${ }^{274}$ | ${ }^{221}$ | 16.8 | 209 | ${ }_{0.808}$ | ${ }_{6} 620$. | 0.005 | ${ }_{13}$ | 459. | 310. | ${ }_{964}$ |  | 0.0013 |  | 0.2 | 0.5 | $\stackrel{-144}{ }$ | 0.9 | 1.45 | $\stackrel{-}{0.1}$ |
| P12 | ss |  | 3882023 | Doy- No enough waero Sosampe |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |













## B2. Domestic Well Quality

| Parameters | Criteria |  | Well Number | Site \# 1 | $\begin{gathered} \hline \text { Site \#2 } \\ \hline \text { Kiowa Ave } \\ \hline \end{gathered}$$24$ | $\begin{gathered} \hline \text { BFD of Site \#2 } \\ \hline \text { Kiowa Ave } \\ \hline 24 \end{gathered}$ | $\begin{gathered} \hline \text { Site \#3 } \\ \hline \text { Spotts Close } \\ \hline 25 \end{gathered}$ | $\begin{gathered} \hline \text { Site \#11 } \\ \hline \text { Farmington Rd } \\ \hline 36 \end{gathered}$ | $\begin{gathered} \hline \text { Site \#12 } \\ \hline \text { Wallace } \mathrm{Dr} \\ \hline \end{gathered}$ | Site \#13Wallace Dr | $\begin{gathered} \hline \text { Site \#10 } \\ \hline \text { Meadowbrook } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\begin{gathered} \hline \text { Hartland Avenue } \\ \hline 99 \end{gathered}$ |  |  |  |  |  |  |  |
|  | Canadian | British Columbia |  | 37 (at house) |  |  |  |  | 38 (at well head) | $\begin{gathered} \hline \text { Meadowbrook } \\ \hline 61 \end{gathered}$ |  |
|  | DWQ Guidelines ${ }^{\text {(1) }}$ | SDWQ Guidelines ${ }^{(2)}$ | Sample Date | 15-Jun-2022 | 14-Jun-2022 | 14-Jun-2022 | 14-Jun-2022 | 14-Jun-2022 | 14-Jun-2022 | 14-Jun-2022 | 15-Jun-2022 |
| Metals |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum, total | $0.1 \pm$ | 9.5 | mg/L | 0.0134 | 0.0005 | < 0.0005 | 0.00128 | 0.00118 | 0.00073 | 0.00075 | 0.0239 |
| Antimony, total | 0.006 | 0.006 | mg/L | 0.00002 | < 0.00002 | < 0.00002 | 0.00002 | 0.00002 | 0.000085 | 0.000084 | 0.00002 |
| Arsenic, total | 0.010 | 0.010 | mg/L | 0.000039 | 0.000035 | 0.000021 | 0.000036 | 0.00002 | 0.00157 | 0.00162 | 0.000049 |
| Barium, total | 2 | n/a | mg/L | 0.00391 | 0.00138 | 0.00135 | 0.0024 | 0.000668 | 0.00914 | 0.0094 | 0.00156 |
| Beryllium, total | n/a | n/a | mg/L | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 |
| Bismuth, total | n/a | n/a | mg/L | < 0.000005 | < 0.000005 | < 0.000005 | < 0.000005 | 0.000005 | 0.000005 | 0.000005 | 0.000005 |
| Boron, total | 5 | 5 | mg/L | 0.01 | 0.164 | 0.16 | 0.138 | 0.719 | 0.28 | 0.28 | 0.01 |
| Cadmium, total | 0.007 | 0.005 | mg/L | < 0.000005 | 0.0000277 | 0.0000272 | 0.0000065 | < 0.000005 | 0.0000122 | 0.000005 | 0.0000081 |
| Calcium, total | n/a | n/a | mg/L | --- | --- | --- | --- | --- | --- | --- | --- |
| Chromium, total | 0.05 | 0.05 | mg/L | 0.0001 | < 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.00028 | 0.0001 | 0.00031 |
| Cobalt, total | n/a | 0.001 | mg/L | 0.0000226 | 0.0000157 | 0.0000147 | 0.0000263 | 0.000005 | 0.0000367 | 0.000061 | 0.0000194 |
| Copper, total | $2^{(1)}$ and $1 \dagger$ | $2^{(2)}$ and $1 \dagger$ | mg/L | 0.00743 | 0.0116 | 0.0119 | 0.0563 | 0.000124 | 0.00731 | 0.00147 | 0.0524 |
| Iron, total | 0.3 † | 0.3 † | mg/L | 0.0299 | < 0.001 | 0.001 | 0.0012 | 0.0104 | 0.0103 | 0.001 | 0.011 |
| Lead, total | 0.005 | 0.005 | mg/L | 0.000258 | 0.000113 | 0.000111 | 0.000994 | 0.0000179 | 0.000658 | 0.000417 | 0.00087 |
| Lithium, total | n/a | n/a | mg/L | 0.0005 | < 0.0005 | < 0.0005 | 0.0005 | 0.0011 | 0.00115 | 0.00111 | 0.0005 |
| Magnesium, total | n/a | n/a | mg/L | 1.07 | 4.64 | 4.51 | 5.68 | 2.75 | 9.3 | 9.44 | 5.23 |
| Manganese, total | $0.12^{\text {"1) }}$ and $0.02 \dagger$ | $0.12^{(2)}$ and $0.02 \dagger$ | mg/L | 0.00346 | 0.000174 | 0.000128 | 0.000376 | 0.000005 | 0.0207 | 0.0637 | 0.000275 |
| Mercury, total | 0.001 | 0.001 | mg/L | < 0.0000019 | < 0.0000019 | < 0.0000019 | 0.0000019 | 0.000124 | < 0.0000019 | < 0.0000019 | 0.0000019 |
| Molybdenum, total | n/a | 0.088 | $\mathrm{mg} / \mathrm{L}$ | 0.000061 | 0.000101 | 0.0001 | 0.000101 | 0.0104 | 0.00114 | 0.00113 | 0.000071 |
| Nickel, total | n/a | 0.08 | $\mathrm{mg} / \mathrm{L}$ | 0.000249 | 0.000183 | 0.00015 | 0.000274 | 0.0000179 | 0.000458 | 0.000447 | 0.000195 |
| Phosphorus, total | n/a | n/a | $\mathrm{mg} / \mathrm{L}$ | 0.0036 | < 0.002 | 0.002 | 0.003 | 0.0011 | 0.0326 | 0.0323 | 0.0077 |
| Potassium, total | n/a | n/a | mg/L | 0.123 | 0.628 | 0.608 | 1.84 | 0.46 | 1.29 | 1.28 | 0.227 |
| Selenium, total | 0.05 | 0.01 | $\mathrm{mg} / \mathrm{L}$ | 0.00004 | 0.00004 | < 0.00004 | 0.00004 | 0.00004 | 0.00006 | 0.000049 | 0.00009 |
| Silicon, total | n/a | n/a | $\mathrm{mg} / \mathrm{L}$ | 1.97 | 9.13 | 9.25 | 9.19 | 8.96 | 8.32 | 8.37 | 9.59 |
| Silver, total | n/a | n/a | mg/L | < 0.000005 | < 0.000005 | < 0.000005 | 0.000005 | < 0.000005 | < 0.000005 | 0.000005 | 0.000005 |
| Sodium, total | 200 † | n/a | mg/L | 3.09 | 12.7 | 12.2 | 11.4 | 37.7 | 21.4 | 21.6 | 6.51 |
| Strontium, total | 7.0 | 7.0 | $\mathrm{mg} / \mathrm{L}$ | 0.0134 | 0.0614 | 0.0606 | 0.0726 | 0.0234 | 0.157 | 0.16 | 0.0412 |
| Sulphur, total | n/a | n/a | $\mathrm{mg} / \mathrm{L}$ | 3 | 3.5 | 3.1 | 6.2 | 4.9 | 9.9 | 9.6 | 3 |
| Thallium, total | n/a | n/a | mg/L | < 0.000002 | < 0.000002 | < 0.000002 | 0.000002 | 0.000002 | 0.000002 | 0.000002 | 0.000002 |
| Tin, total | n/a | n/a | mg/L | < 0.0002 | < 0.0002 | < 0.0002 | 0.0002 | 0.0002 | 0.0002 | 0.0002 | 0.0002 |
| Titanium, total | n/a | n/a | $\mathrm{mg} / \mathrm{L}$ | < 0.0005 | < 0.0005 | < 0.0005 | 0.0005 | < 0.0005 | 0.0005 | 0.0005 | 0.00245 |
| Uranium, total | 0.02 | 0.02 | mg/L | 0.0000048 | 0.0000862 | 0.0000876 | 0.000175 | 0.000002 | 0.00034 | 0.000348 | 0.0000052 |
| Vanadium, total | n/a | n/a | mg/L | < 0.0002 | 0.00058 | 0.00057 | 0.00086 | < 0.0002 | 0.00046 | 0.00047 | 0.00066 |
| Zinc, total | $5 \dagger$ | $3^{(2)}$ and $5 \dagger$ | mg/L | 0.00412 | 0.00878 | 0.00808 | 0.0344 | 0.00239 | 0.0058 | 0.00255 | 0.00608 |
| Zirconium, total | n/a | n/a | mg/L | 0.0001 | 0.0001 | < 0.0001 | 0.0001 | < 0.0001 | 0.0001 | 0.0001 | 0.0001 |
|  |  |  |  |  |  |  |  |  |  |  |  |
| Hardness (as CaCO3) | n/a | n/a | mg/L | 16.1 | 176 | 173 | 158 | 35.5 | 144 | 145 | 68.3 |
| Chloride, dissolved | 250 † | 250 † | mg/L | 4.5 | 15 | 15 | 38 | 9.2 | 15 | 15 | 8.6 |
| Total dissolved solids | 500 † | n/a | $\mathrm{mg} / \mathrm{L}$ | 38 | 240 | 260 | 280 | 180 | 230 | 230 | 110 |
| Specific Conductivity (Lab) | n/a | n/a | $\mu \mathrm{s} / \mathrm{cm}$ | 50 | 400 | 400 | 400 | 240 | 390 | 390 | 170 |
| Conductivity (Field) | n/a | n/a | $\mu \mathrm{s} / \mathrm{cm}$ | 41 | 305 | 305 | 305 | 176 | 287 | 285 | 125 |
| pH (lab) | 7.0 to $10.5 \dagger$ | n/a | pH | 7.22 | 8.09 | 8.07 | 7.66 | 8.32 | 8.37 | 8.36 | 6.84 |
| pH (field) | 7.0 to 10.5 t | n/a | pH | 7.54 | 7.04 | 7.04 | 6.76 | 8.78 | 7.73 | 7.7 | 6.86 |
| Temperature (field) | $15 \dagger$ | $15 \dagger$ | ${ }^{\circ} \mathrm{C}$ | 14.2 | 13.9 | 13.9 | 13.5 | 11.8 | 12.7 | 12.3 | 11.2 |
| Ammonia | n/a | n/a | mg/L | --- | --- | --- | --- | --- | --- | --- | --- |

Notes:

1) Heath Canada (updated in September 2020 version). Guidelines for Canadian Drinking Water Quality-Summary Table
${ }^{(2)}$ ) British Columbia Approved Source Drinking Water Quality Guideline (SDWQGs) (2020 edition)
${ }^{\text {F }}$ + Limit for this parameter is an operational guideline (OG) only.
imit tor this guideline is an iterim value.
Limit for this guideline is an terim valu
BOLD 'Sample concentrations expressed in bold exceed $A O$ drinking water quality guidelines
n/a-no drinking water qualitirs yudeline is asaiallew for this parameter water quality guidelines.
n/a - no drinking water quality guideline is available for this parameter.

|  |  |  |  | Site \#9 |  | Site \#4 |  | Site \# 6 |  | Site \#7 |  | Site \#8 |  | Site \# 16 |  | Site \# 5 |  | Site \#14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameters |  |  | Well Number | Wildview Cr |  | Lohr Rd |  | Lohr Rd |  | ohr Road |  | Lohr Rd |  | Spotts Close |  | Lohr Rd |  | rtland Ave |
| Parameters | Canadian | British Columbia |  | 47 |  | 50 |  | 52 |  | 53 |  | 80 |  | 81 |  | 51 |  | n/a |
|  | DWQ Guidelines ${ }^{(1)}$ | SDWQ Guidelines ${ }^{(2)}$ | Sample Date | 16-Jun-2022 |  | 15-Jun-2022 |  | 5-Jun-2022 |  | 5-Jun-2022 |  | 5-Jun-2022 |  | 14-Jun-2022 |  | 15-Jun-2022 |  | -Jun-2022 |
| Metals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum, total | $0.1 \pm$ | 9.5 | mg/L | 0.0479 |  | 0.0103 |  | 0.0062 |  | 0.00572 | < | 0.0005 |  | 0.00175 |  | 0.00057 |  | 0.00861 |
| Antimony, total | 0.006 | 0.006 | mg/L | 0.00002 | < | 0.00002 | < | 0.00002 | $<$ | 0.00002 | < | 0.00002 | < | 0.00002 | < | 0.00002 | < | 0.00002 |
| Arsenic, total | 0.010 | 0.010 | mg/L | 0.000056 |  | 0.000042 |  | 0.00005 | < | 0.00002 | < | 0.00002 | < | 0.00002 |  | 0.000053 | < | 0.00002 |
| Barium, total | 2 | n/a | mg/L | 0.0043 |  | 0.00776 |  | 0.00817 |  | 0.00257 |  | 0.0056 |  | 0.0022 |  | 0.0172 |  | 0.00214 |
| Beryllium, total | n/a | n/a | mg/L | 0.000023 | < | 0.00001 | < | 0.00001 | < | 0.00001 |  | 0.000011 | < | 0.00001 | < | 0.00001 |  | 0.000012 |
| Bismuth, total | n/a | n/a | mg/L | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 |
| Boron, total | 5 | 5 | mg/L | 0.01 |  | 0.076 |  | 0.071 |  | 0.011 | < | 0.01 |  | 0.93 |  | 0.044 |  | 0.012 |
| Cadmium, total | 0.007 | 0.005 | mg/L | 0.0000372 |  | 0.0000106 | < | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 |  | 0.0000291 |
| Calcium, total | n/a | n/a | mg/L | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |
| Chromium, total | 0.05 | 0.05 | mg/L | 0.0001 |  | 0.00011 | $<$ | 0.0001 | < | 0.0001 | < | 0.0001 | < | 0.0001 | < | 0.0001 | < | 0.0001 |
| Cobalt, total | n/a | 0.001 | mg/L | 0.0000283 |  | 0.000057 |  | 0.0000241 |  | 0.0000084 |  | 0.0000055 | < | 0.000005 |  | 0.0000107 |  | 0.0000334 |
| Copper, total | $2^{\text {(1) }}$ and $1 \dagger$ | $2^{(2)}$ and $1 \dagger$ | mg/L | 0.195 |  | 0.0381 |  | 0.00515 |  | 0.00006 |  | 0.014 |  | 0.000102 |  | 0.00565 |  | 0.00644 |
| Iron, total | $0.3 \dagger$ | $0.3 \dagger$ | mg/L | 0.0292 |  | 0.142 |  | 0.002 |  | 0.0026 |  | 0.0152 |  | 0.0218 |  | 0.0013 |  | 0.119 |
| Lead, total | 0.005 | 0.005 | mg/L | 0.00171 |  | 0.0022 |  | 0.000488 | < | 0.000005 |  | 0.000366 |  | 0.000163 |  | 0.000288 |  | 0.000319 |
| Lithium, total | n/a | n/a | mg/L | 0.0005 | < | 0.0005 | < | 0.0005 | < | 0.0005 | < | 0.0005 |  | 0.0012 | < | 0.0005 | < | 0.0005 |
| Magnesium, total | n/a | n/a | mg/L | 2.17 |  | 2.88 |  | 4.88 |  | 2.62 |  | 2.68 |  | 0.69 |  | 6.24 |  | 3.36 |
| Manganese, total | $0.12{ }^{\text {(1) }}$ and $0.02 \dagger$ | $0.12^{(2)}$ and $0.02 \dagger$ | mg/L | 0.0162 |  | 0.000978 |  | 0.000167 |  | 0.000188 |  | 0.000509 |  | 0.000476 |  | 0.000194 |  | 0.00225 |
| Mercury, total | 0.001 | 0.001 | mg/L | 0.0000032 | < | 0.0000019 | < | 0.0000019 | < | 0.0000019 | < | 0.0000019 | < | 0.0000019 | < | 0.0000019 | < | 0.0000019 |
| Molybdenum, total | n/a | 0.088 | mg/L | 0.00005 |  | 0.000311 |  | 0.000462 |  | 0.000449 |  | 0.00038 |  | 0.00204 |  | 0.00102 |  | 0.000274 |
| Nickel, total | n/a | 0.08 | mg/L | 0.000146 |  | 0.000155 |  | 0.000183 |  | 0.000037 |  | 0.000099 |  | 0.000042 |  | 0.000591 |  | 0.000068 |
| Phosphorus, total | n/a | n/a | mg/L | 0.0086 |  | 0.004 |  | 0.0042 |  | 0.0069 |  | 0.0029 | < | 0.002 | < | 0.002 |  | 0.0035 |
| Potassium, total | n/a | n/a | mg/L | 0.372 |  | 0.338 |  | 0.542 |  | 0.262 |  | 0.411 |  | 0.212 |  | 0.388 |  | 0.307 |
| Selenium, total | 0.05 | 0.01 | mg/L | 0.000086 | < | 0.00004 |  | 0.000067 | < | 0.00004 |  | 0.000052 | < | 0.00004 |  | 0.000194 |  | 0.000092 |
| Silicon, total | n/a | n/a | mg/L | 6.9 |  | 6.31 |  | 7.64 |  | 7.77 |  | 8.04 |  | 11.2 |  | 7.73 |  | 6.91 |
| Silver, total | n/a | n/a | mg/L | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 | < | 0.000005 |
| Sodium, total | 200 † | n/a | mg/L | 7.77 |  | 7 |  | 5.8 |  | 4.48 |  | 10 |  | 42.7 |  | 6.24 |  | 15.4 |
| Strontium, total | 7.0 | 7.0 | mg/L | 0.0317 |  | 0.0436 |  | 0.0797 |  | 0.0379 |  | 0.0614 |  | 0.0488 |  | 0.137 |  | 0.0351 |
| Sulphur, total | n/a | n/a | mg/L | < 3 | < | 3 |  | 3.1 | < | 3 | < | 3 |  | 5 |  | 3 |  | 8 |
| Thallium, total | n/a | n/a | mg/L | 0.000002 | < | 0.000002 | < | 0.000002 | < | 0.000002 | < | 0.000002 | < | 0.000002 | < | 0.000002 | < | 0.000002 |
| Tin, total | n/a | n/a | mg/L | 0.0002 |  | 0.00043 | < | 0.0002 | < | 0.0002 | < | 0.0002 | < | 0.0002 | < | 0.0002 | < | 0.0002 |
| Titanium, total | n/a | n/a | mg/L | 0.00086 | < | 0.0005 | < | 0.0005 | < | 0.0005 | < | 0.0005 | < | 0.0005 | < | 0.0005 | < | 0.0005 |
| Uranium, total | 0.02 | 0.02 | mg/L | 0.0000663 |  | 0.0000191 |  | 0.000337 |  | 0.000113 |  | 0.000702 |  | 0.0000042 |  | 0.00061 |  | 0.0000152 |
| Vanadium, total | n/a | n/a | mg/L | 0.0002 |  | 0.00091 |  | 0.00058 |  | 0.00046 | < | 0.0002 | < | 0.0002 |  | 0.00053 |  | 0.00036 |
| Zinc, total | $5 \dagger$ | $3^{(2)}$ and $5 \dagger$ | mg/L | 0.00761 |  | 0.0129 |  | 0.00275 |  | 0.00088 |  | 0.00742 |  | 0.00644 |  | 0.00289 |  | 0.0263 |
| Zirconium, total | n/a | n/a | mg/L | 0.0001 | < | 0.0001 | < | 0.0001 | < | 0.0001 | < | 0.0001 | < | 0.0001 | < | 0.0001 | < | 0.0001 |
| Conventionals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Hardness (as CaCO3) | n/a | n/a | mg/L | 37.2 |  | 50.3 |  | 113 |  | 56.5 |  | 95.7 |  | 17.3 |  | 154 |  | 49.8 |
| Chloride, dissolved | $250 \dagger$ | $250 \dagger$ | mg/L | 8.7 |  | 11 |  | 10 |  | 5.3 |  | 7.9 |  | 15 |  | 14 |  | 17 |
| Total dissolved solids | 500 † | n/a | mg/L | 100 |  | 92 |  | 160 |  | 88 |  | 130 |  | 140 |  | 190 |  | 120 |
| Specific Conductivity (Lab) | n/a | n/a | $\mu \mathrm{s} / \mathrm{cm}$ | 120 |  | 140 |  | 260 |  | 140 |  | 240 |  | 230 |  | 330 |  | 190 |
| Conductivity (Field) | n/a | n/a | $\mu \mathrm{s} / \mathrm{cm}$ | 84 |  | 106 |  | 193 |  | 138 |  | 171 |  | 165 |  | 236 |  | 139 |
| pH (lab) | 7.0 to 10.5 † | n/a | pH | 6.8 |  | 6.65 |  | 7.17 |  | 7.03 |  | 7.25 |  | 8.48 |  | 7.54 |  | 6.53 |
| pH (field) | 7.0 to 10.5 † | n/a | pH | 7.06 |  | 6.69 |  | 7.03 |  | 6.72 |  | 6.82 |  | 9.11 |  | 7.21 |  | 6.73 |
| Temperature (field) | $15 \dagger$ | 15 † | ${ }^{\circ} \mathrm{C}$ | 10.6 |  | 10.5 |  | 12.1 |  | 10.1 |  | 11.4 |  | 10.7 |  | 10 |  | 10.7 |
| Ammonia | n/a | n/a | mg/L | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |

Notes:

1) Hearin Canada (uppaated in June 2019 version). Guidelines for Canadian Drinking Water Quality-Summary Tabl
(2) British Columbia Approved Source Drinking Water Quality

Limit for this parameter is an operational guideline (OG) onls.

* Limit for this guideline is an iterim value.

BOLD 'Sample concentrations expressed in bold exceed AO drinking water पuality guidelines
Sample concentrations highighted in yellow exceed drinking water quality guidelines.
n/a - no drinking water quality guideline is availabe for this parameter.

| Parameters | Criteria |  | Well Number | Site \#17 | $\begin{gathered} \hline \text { Site \#21 } \\ \hline \text { Mark Lane } \end{gathered}$ | $\begin{gathered} \hline \text { Site \#20 } \\ \hline \text { Mark Lane } \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { BFD of Site \#20 } \\ \hline \text { Mark Lane } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Durrance Close |  |  |  |
|  | Canadian | British Columbia |  | n/a | n/a | n/a | n/a |
|  | DWQ Guidelines ${ }^{\text {(1) }}$ | SDWQ Guidelines ${ }^{(2)}$ |  | Sample Date | 14-Jun-2022 | 16-Jun-2022 | 16-Jun-2022 | 16-Jun-2022 |
|  |  |  |  |  |  |  |  |
| Aluminum, total | $0.1 \pm$ | 9.5 | mg/L | --- | --- | --- | --- |
| Antimony, total | 0.006 | 0.006 | mg/L | -- | -- | --- | --- |
| Arsenic, total | 0.010 | 0.010 | mg/L | -- | --- | --- | --- |
| Barium, total | 2 | n/a | mg/L | --- | --- | --- | --- |
| Beryllium, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Bismuth, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Boron, total | 5 | 5 | mg/L | --- | --- | --- | --- |
| Cadmium, total | 0.007 | 0.005 | mg/L | -- | --- | --- | --- |
| Calcium, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Chromium, total | 0.05 | 0.05 | mg/L | --- | --- | --- | --- |
| Cobalt, total | n/a | 0.001 | mg/L | --- | --- | --- | --- |
| Copper, total | $2^{(1)}$ and $1+$ | $2^{(2)}$ and $1+$ | mg/L | --- | --- | --- | --- |
| Iron, total | 0.3 † | 0.3 † | mg/L | --- | --- | --- | --- |
| Lead, total | 0.005 | 0.005 | mg/L | --- | --- | --- | --- |
| Lithium, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Magnesium, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Manganese, total | $0.12^{(1)}$ and $0.02 \dagger$ | $0.12^{(2)}$ and $0.02 \dagger$ | mg/L | --- | --- | --- | --- |
| Mercury, total | 0.001 | 0.001 | mg/L | --- | --- | --- | --- |
| Molybdenum, total | n/a | 0.088 | mg/L | --- | --- | --- | --- |
| Nickel, total | n/a | 0.08 | mg/L | --- | --- | --- | --- |
| Phosphorus, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Potassium, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Selenium, total | 0.05 | 0.01 | mg/L | --- | --- | --- | --- |
| Silicon, total | n/a | n/a | $\mathrm{mg} / \mathrm{L}$ | --- | --- | --- | --- |
| Silver, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Sodium, total | 200 † | n/a | mg/L | 6.34 | 9.92 | 11.5 | 11.4 |
| Strontium, total | 7.0 | 7.0 | $\mathrm{mg} / \mathrm{L}$ | --- | --- | --- | --- |
| Sulphur, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Thallium, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Tin, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Titanium, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Uranium, total | 0.02 | 0.02 | mg/L | --- | --- | --- | --- |
| Vanadium, total | n/a | n/a | mg/L | --- | --- | --- | --- |
| Zinc, total | $5 \dagger$ | $3^{(2)}$ and $5 \dagger$ | mg/L | --- | --- | --- | --- |
| Zirconium, total | n/a | n/a | mg/L | --- | -- | -- | -- |
| Conventionals |  |  |  |  |  |  |  |
| Hardness (as CaCO3) | n/a | n/a | mg/L | --- | --- | --- | --- |
| Chloride, dissolved | $250 \dagger$ | 250 † | mg/L | 4.4 | 28 | 15 | 15 |
| Total dissolved solids | $500 \dagger$ | n/a | $\mathrm{mg} / \mathrm{L}$ | --- | --- | --- |  |
| Specific Conductivity (Lab) | n/a | n/a | $\mu \mathrm{s} / \mathrm{cm}$ | 390 | 550 | 360 | 360 |
| Conductivity (Field) | n/a | n/a | $\mu \mathrm{s} / \mathrm{cm}$ | 281 | 401 | 259 | 259 |
| pH (lab) | 7.0 to $10.5 \dagger$ | n/a | pH | 8.26 | 7.78 | 8.07 | 8.09 |
| pH (field) | 7.0 to 10.5 † | n/a | pH | 7.08 | 7.2 | 7.79 | 7.79 |
| Temperature (field) | 15 † | $15 \dagger$ | ${ }^{\circ} \mathrm{C}$ | 10.9 | 12.2 | 11.1 | 11.1 |
| Ammonia | n/a | n/a | mg/L | < 0.015 | < 0.015 | < 0.015 | < 0.015 |

Notes:

1) Heath Canada (updated in June 2019 version). Guidelines for Canadian Drinking Water Quality-Summary Table.
(2) British Columbia Approved Source Drinking Water Quality Guideline (SDWQGs) (2020 edition).

Linit tor this parameter is an operational guideline (OG) only
Limit for this parameter is an aesthetic objective ( $A 0$ ), not a human heath objective.
Limit for this guideline is an iterim value.
BOLD Sample concentrations expressed in bold exceed $A O$ drinking water quality guidelines
Sample concentrations highighted in yellow exceed drinking water quality guidelines.
na a -no drinking water quality guideline is available tor this parameler.

B3. Surface Water Quality


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B4. Monthly Leachate Quality Data Hartland Valve Chamber

| State | Parameter | Unit | Sewer <br> min | max |  | Hartland Valve Chamber FR1 26-Apr-2022 |  |  | Hartland Valve Chamber FR2 26-Apr-2022 |  |  | Hartland Valve Chamber FRM 26-Apr-2022 |  |  | Hartland Valve Chamber SS 30-May-2022 |  |  | Hartland Valve Chamber SS 27-Jun-2022 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 27-Jul-2022 |  |  | Hartland Valve Chamber FR1 22-Aug-2022 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CONVENTIONALS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Alkalinity - Total - Ph 4.5 | mg/L | --- | --- |  | --- |  |  | --- |  |  | --- |  |  | --- |  |  | --- |  | --- |  |  | --- |  |
| Total | BOD | $\mathrm{mg} / \mathrm{L}$ | --- | 500 | $<$ | 20 |  | < | 20 |  | < | 20 |  |  | 25. |  |  | 18. |  | 27. |  |  | 27. |  |
| Total | CBOD | mg/L | --- | --- |  | --- |  |  | --- |  |  | --- |  |  | --- |  |  | --- |  | --- |  |  | --- |  |
| Total | Chloride | $\mathrm{mg} / \mathrm{L}$ | --- | 1500 |  | 330 |  |  | 320 |  |  | 325 |  |  | 300. |  |  | 350. |  | 420. |  |  | 410. |  |
| Total | COD | $\mathrm{mg} / \mathrm{L}$ | --- | 1000 |  | 9420 | a |  | 8610 | a |  | 9015 | a |  | 6320. | a |  | 462. |  | 2630. | a |  | 127. |  |
| Total | Conductivity | $\mu \mathrm{S} / \mathrm{cm}$ | --- | --- |  | 3569 |  |  | 3569 |  |  | 3569 |  |  | 3560. |  |  | 4249. |  | 4654. |  |  | 4994. |  |
| Total | Cyanide SAD | mg/L | --- | 1 |  | 0.013 |  |  | 0.013 |  |  | 0.013 |  |  | 0.0236 |  |  | 0.011 |  | 0.014 |  | $<$ | 0.01 |  |
| Total | Cyanide WAD | mg/L | --- | 1 |  | 0.01 |  |  | 0.012 |  |  | 0.011 |  |  | 0.0163 |  |  | --- | < | 0.01 |  | < | 0.01 |  |
| Total | Dissolved Oxygen | mg/L | --- | --- |  | 1.37 |  |  | 1.37 |  |  | 1.37 |  |  | 0.3 |  |  | 0.25 |  | --- |  |  | --- |  |
| Total | Fecal Coliforms | CFU/100 mL | --- | --- |  | --- |  |  | --- |  |  | --- |  |  | --- |  |  | --- |  | --- |  |  | --- |  |
| Total | N - Ammonia (As N) | $\mathrm{mg} / \mathrm{L}$ | --- | --- |  | 250 |  |  | 250 |  |  | 250 |  |  | 260. |  |  | 270. |  | 300. |  |  | 310. |  |
| Total | N - Nitrite (As N) | mg/L | --- | --- |  | 1.17 |  |  | 1.21 |  |  | 1.190 |  |  | 0.615 |  |  | 0.0286 |  | 0.367 |  |  | 0.115 |  |
| Total | N - Nitrate (As N) | mg/L | --- | --- |  | 14 |  |  | 15 |  |  | 15 |  |  | 11.3 |  |  | 0.11 |  | 0.38 |  |  | 0.68 |  |
| Total | N - Nitrite+Nitrite (As N ) | $\mathrm{mg} / \mathrm{L}$ | --- | --- |  | --- |  |  | --- |  |  | --- |  |  | --- |  |  | --- |  | --- |  |  | --- |  |
| Total | N - TKN (As N) | $\mathrm{mg} / \mathrm{L}$ | --- | --- |  | --- |  |  | --- |  |  | --- |  |  | --- |  |  | --- |  | --- |  |  | --- |  |
| Total | N - Total (As N) | $\mathrm{mg} / \mathrm{L}$ | --- | --- |  | --- |  |  | --- |  |  | --- |  |  | --- |  |  | --- |  | --- |  |  | --- |  |
| Total | Oil \& Grease, Mineral | $\mathrm{mg} / \mathrm{L}$ | --- | 15 | $<$ | 2 |  | $<$ | 2 |  | $<$ | 2. |  | $<$ | 2. |  | < | 2. | $<$ | 2. |  | < | 2. |  |
| Total | Oil \& grease, total | $\mathrm{mg} / \mathrm{L}$ | --- | 100 | < | 1 |  | < | 1 |  | < | 1. |  | $<$ | 1. |  | < | 1. | < | 1. |  | $<$ | 1. |  |
| Total | ORP | mV | --- | --- |  | 113.0 |  |  | 113.0 |  |  | 113 |  |  | 146. |  |  | - 12. |  | 64. |  |  | - 43. |  |
| Total | pH | pH | 5.5 | 11 |  | 7.82 |  |  | 7.81 |  |  | 7.82 |  |  | 7.25 |  |  | 7.75 |  | 8.03 |  |  | 8.43 |  |
| Dissolved | Sulphide | mg/L | --- | 1 | < | 0.045 |  | $<$ | 0.045 |  | $<$ | 0.0450 |  |  | 0.053 |  |  | 0.009 |  | 0.0095 |  |  | 0.0057 |  |
| Total | Sulphide | mg/L | --- | 1 | < | 0.045 |  | < | 0.045 |  | $<$ | 0.0450 |  |  | 0.066 |  |  | 0.027 | < | 0.009 |  |  | 0.067 |  |
| Dissolved | Sulphate | $\mathrm{mg} / \mathrm{L}$ | --- | 1500 |  | 130 |  | < | 100 |  |  | 115 |  |  | 44. |  |  | 25. |  | 19. |  | < | 1. |  |
|  | Temperature | ${ }^{\circ} \mathrm{C}$ | --- | --- |  | 16.4 |  |  | 16.4 |  |  | 16.4 |  |  | 17.7 |  |  | 21.4 |  | 24.5 |  |  | 22.5 |  |
| Total | TOC | mg/L | --- | --- |  | --- |  |  | --- |  |  | --- |  |  | --- |  |  | --- |  | --- |  |  | --- |  |
| Total | Total Phenols | $\mathrm{mg} / \mathrm{L}$ | --- | 1 | < | 0.03 |  | < | 0.03 |  | $<$ | 0.03 |  |  | 0.018 |  |  | 0.0068 |  | 0.1 |  | $<$ | 0.015 |  |
| Total | TSS | $\mathrm{mg} / \mathrm{L}$ | --- | 350 |  | 19 |  |  | 19 |  |  | 19 |  |  | 31. |  |  | 17. |  | 23. |  |  | 12. |  |


| State | Parameter | Unit | Sewer | max |  | Hartland Valve Chamber FR1 26-Apr-2022 |  | Hartland <br> Valve <br> Chamber <br> FR2 <br> 26-Apr-2022 |  | Hartland Valve Chamber FRM 26-Apr-2022 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 30-May-2022 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 27-Jun-2022 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 27-Jul-2022 |  | Hartland Valve Chamber FR1 22-Aug-2022 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| METALS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Aluminum | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Antimony | $\mu \mathrm{g} / \mathrm{L}$ | -- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Arsenic | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.4 |  | 6.8 |  | 6.74 |  | 6.77 |  | 5.58 |  | 6.61 |  | 7.72 |  | 8.48 |  |
| Total | Barium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Beryllium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Bismuth | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Boron | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Cadmium | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.3 |  | 0.04 |  | 0.052 |  | 0.0460 |  | 0.038 |  | 0.032 |  | 0.057 |  | 0.034 |  |
| Total | Calcium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | -- |  | --- |  |
| Total | Chromium | $\mu \mathrm{g} / \mathrm{L}$ | --- | 4 |  | 44 |  | 49.8 |  | 46.90 |  | 41.6 |  | 50.8 |  | 59.5 |  | 63.7 |  |
| Total | Chromium III | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Chromium Vi | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Cobalt | $\mu \mathrm{g} / \mathrm{L}$ | --- | 5 |  | 13.3 |  | 13 |  | 13.15 |  | 11.2 |  | 14.2 |  | 15.7 |  | 17.6 |  |
| Total | Copper | $\mu \mathrm{g} / \mathrm{L}$ | --- | 1 |  | 10.2 |  | 10.5 |  | 10.35 |  | 8.34 |  | 11.6 |  | 9.92 |  | 9.08 |  |
| Total | Hardness (As Caco3) | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Iron | $\mu \mathrm{g} / \mathrm{L}$ | --- | 50 |  | 2140 |  | 2120 |  | 2130 |  | 2240 |  | 2180 |  | 3120 |  | 3040 |  |
| Total | Lead | $\mu \mathrm{g} / \mathrm{L}$ | --- | 1 |  | 0.667 |  | 0.664 |  | 0.6655 |  | 0.575 |  | 0.459 |  | 0.66 |  | 0.65 |  |
| Total | Lithium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Magnesium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Manganese | $\mu \mathrm{g} / \mathrm{L}$ | --- | 5 |  | 674 |  | 656 |  | 665 |  | 657 |  | 699 |  | 741 |  | 839 |  |
| Total | Mercury | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.02 | < | 0.019 | < | 0.019 | < | 0.02 | $<$ | 0.038 | < | 0.038 | $<$ | 0.019 | $<$ | 0.019 |  |
| Total | Molybdenum | $\mu \mathrm{g} / \mathrm{L}$ | --- | 5 |  | 1.98 |  | 2.35 |  | 2.165 |  | 2.39 |  | 3.01 |  | 3.5 |  | 3.36 |  |
| Total | Nickel | $\mu \mathrm{g} / \mathrm{L}$ | --- | 3 |  | 39.8 |  | 38.4 |  | 39.1 |  | 34.3 |  | 40.2 |  | 47.1 |  | 51.6 |  |
| Total | Phosphorus | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Potassium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Selenium | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.3 |  | 0.42 |  | 0.46 |  | 0.44 |  | 0.356 |  | 0.48 |  | 0.59 |  | 0.49 |  |
| Total | Silicon | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Silver | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.5 | < | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | < | 0.02 | $<$ | 0.05 | < | 0.05 |  |
| Total | Sodium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Strontium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Sulphur | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | -- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Thallium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Tin | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Titanium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Uranium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Vanadium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Zinc | $\mu \mathrm{g} / \mathrm{L}$ | --- | 3 |  | 16.8 |  | 15.5 |  | 16.15 |  | 11.8 |  | 12.3 |  | 11.3 |  | 13.1 |  |
| Total | Zirconium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | -- |  | --- |  |


| State | Parameter | Unit | Sewer Use Criteria |  | HartlandValveChamberFR126-Apr-2022 |  |  | Hartland Valve Chamber FR2 26-Apr-2022 |  | Hartland <br> Valve <br> Chamber <br> FRM <br> 26-Apr-2022 |  | Hartland Valve Chamber SS 30-May-2022 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 27-Jun-2022 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 27-Jul-2022 |  | Hartland Valve Chamber FR1 22-Aug-2022 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| POLYCYCLIC AROMATIC HYDROCARBONS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Total PAHs | $\mu \mathrm{g} / \mathrm{L}$ |  | 0.05 |  | 6.60 |  | 7.30 |  | 6.95 |  | 7.90 |  | 0.76 |  | 2.10 |  | 1.20 |  |
| LOW WEIGHT |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 2-Chloronaphthalene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | 2-Methylnaphthalene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Acenaphthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 2.9 |  | 3.1 |  | 3.000 |  | 2.5 |  | 0.071 |  | 0.44 |  | 0.3 |  |
| Total | Acenaphthylene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.055 |  | 0.056 |  | 0.056 |  | 0.068 |  | 0.011 |  | 0.025 |  | 0.016 |  |
| Total | Anthracene | $\mu \mathrm{g} / \mathrm{L}$ | --- | -- |  | 0.17 |  | 0.18 |  | 0.175 |  | 0.19 | < | 0.01 |  | 0.17 |  | 0.035 |  |
| Total | Fluorene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 1.8 |  | 2. |  | 1.900 |  | 1.8 |  | 0.15 |  | 0.54 |  | 0.32 |  |
| Total | Naphthalene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.059 |  | 0.076 |  | 0.068 |  | 0.24 |  | 0.04 |  | 0.06 |  | 0.12 |  |
| Total | Phenanthrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.3 |  | 0.32 |  | 0.310 |  | 0.43 |  | 0.011 |  | 0.059 |  | 0.038 |  |
| Total | Total Lmw-Pah'S | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| HIGH WEIGHT |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Benzo(A)Anthracene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.042 |  | 0.047 |  | 0.0445 |  | 0.068 |  | 0.023 |  | 0.048 |  | 0.022 |  |
| Total | Benzo(A)Pyrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.0078 |  | 0.0081 |  | 0.00795 |  | 0.019 | < | 0.005 |  | 0.014 |  | 0.0066 |  |
| Total | Benzo(B)Fluoranthene + Benzo(J)Fluoranthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.011 |  | 0.012 |  | 0.0115 |  | 0.024 | < | 0.01 |  | 0.02 | $<$ | 0.01 |  |
| Total | Benzo(G,H,I)Perylene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | < | 0.02 | $<$ | 0.02 | < | 0.02 |  | 0.26 | < | 0.02 | < | 0.02 | < | 0.02 |  |
| Total | Benzo(K)Fluoranthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | < | 0.01 | < | 0.01 | $<$ | 0.01 |  | 0.012 | $<$ | 0.01 | < | 0.01 | < | 0.01 |  |
| Total | Chrysene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.037 |  | 0.042 |  | 0.0395 |  | 0.054 |  | 0.029 |  | 0.063 |  | 0.021 |  |
| Total | Dibenzo(A,H)Anthracene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.02 | < | 0.02 | < | 0.02 |  | 0.16 | < | 0.02 | < | 0.02 | < | 0.02 |  |
| Total | Fluoranthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.56 |  | 0.62 |  | 0.59 |  | 0.7 |  | 0.2 |  | 0.34 |  | 0.17 |  |
| Total | Indeno(1,2,3-C,D)Pyrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.02 | $<$ | 0.02 | < | 0.02 |  | 0.18 | $<$ | 0.02 | < | 0.02 | $<$ | 0.02 |  |
| Total | Pyrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.31 |  | 0.39 |  | 0.35 |  | 0.45 |  | 0.23 |  | 0.33 |  | 0.14 |  |
| Total | Total Hmw-Pah'S | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| VOLATILE ORGANICS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Dissolved | Benzene | $\mu \mathrm{g} / \mathrm{L}$ | --- | 100 |  | 0.6 |  | 0.51 |  | 0.56 |  | 0.63 | < | 0.4 | < | 0.4 | < | 0.4 |  |
| Dissolved | Ethylbenzene | $\mu \mathrm{g} / \mathrm{L}$ | --- | 200 | $<$ | 0.4 | < | 0.4 | < | 0.40 |  | 0.55 | $<$ | 0.4 | $<$ | 0.4 | < | 0.4 |  |
| Total | M \& P Xylenes | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | -- |  |
| Total | Methyl Tertiary Butyl Ether | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | -- |  | --- |  | -- |  |
| Total | O-Xylene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Styrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Dissolved | Toluene | $\mu \mathrm{g} / \mathrm{L}$ | --- | 200 |  | 1.5 |  | 1.1 |  | 1.30 |  | 1.2 | < | 0.4 | < | 0.4 | < | 0.4 |  |
| Dissolved | Xylenes | $\mu \mathrm{g} / \mathrm{L}$ | --- | 200 |  | 1.2 |  |  |  | 1.10 |  | 1.7 | < | 0.4 | < | 0.4 |  | 0.78 |  |

Notes:

- Exceeded maximum allowable value specified in CRD Sewer Use Bylaw 2922
*     - Exceedances are due to compromised/expired preservatives, and results are not representative
$---=$ Not available.
FR1 - Field replicate 1.
FR2 - Field replicate 2.
FR3 - Field replicate 3.
FRM - Mean of field replicates
SS - Single Sample
NS- Not sampled


| State | Parameter | Unit | Sewe min | riteria <br>  <br> 1 |  | Hartland Valve Chamber FR2 22-Aug-2022 |  | Hartland Valve Chamber FRM 22-Aug-2022 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 27/Sep/2022 |  | Hartland Valve Chamber SS 19/Oct/2022 |  | Hartland <br> Valve <br> Chamber <br> FR1 <br> 30/Nov/2022 |  | Hartland <br> Valve <br> Chamber <br> FR2 <br> 30/Nov/2022 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| metals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Aluminum | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Antimony | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Arsenic | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.4 |  | 8.18 |  | 8.33 |  | 6.38 |  | 8.16 |  | 5.51 |  | 5.48 |  |
| Total | Barium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Beryllium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Bismuth | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Boron | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Cadmium | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.3 |  | 0.048 |  | 0.041 |  | 0.15 |  | 0.089 |  | 0.0848 |  | 0.0761 |  |
| Total | Calcium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Chromium | $\mu \mathrm{g} / \mathrm{L}$ | --- | 4 |  | 61.6 |  | 62.65 |  | 36.80 |  | 72.5 |  | 35.4 |  | 33.8 |  |
| Total | Chromium III | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Chromium Vi | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Cobalt | $\mu \mathrm{g} / \mathrm{L}$ | --- | 5 |  | 16.8 |  | 17.2 |  | 15.00 |  | 24.1 |  | 13.4 |  | 13 |  |
| Total | Copper | $\mu \mathrm{g} / \mathrm{L}$ | --- | 1 |  | 9.18 |  | 9.13 |  | 15.10 |  | 13.8 |  | 23.6 |  | 22.3 |  |
| Total | Hardness (As Caco3) | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Iron | $\mu \mathrm{g} / \mathrm{L}$ | --- | 50 |  | 3070 |  | 3055 |  | 4,600.00 |  | 3650 |  | 3670 |  | 3070 |  |
| Total | Lead | $\mu \mathrm{g} / \mathrm{L}$ | --- | 1 |  | 0.92 |  | 0.785 |  | 1.00 |  | 0.92 |  | 1.41 |  | 1.27 |  |
| Total | Lithium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Magnesium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Manganese | $\mu \mathrm{g} / \mathrm{L}$ | --- | 5 |  | 818 |  | 828.5 |  | 965.00 |  | 856 |  | 650 |  | 640 |  |
| Total | Mercury | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.02 | $<$ | 0.019 | $<$ | 0.019 | < | 0.04 | $<$ | 0.019 | $<$ | 0.019 | $<$ | 0.019 |  |
| Total | Molybdenum | $\mu \mathrm{g} / \mathrm{L}$ | --- | 5 |  | 3.43 |  | 3.395 |  | 2.87 |  | 4.72 |  | 5.23 |  | 5.29 |  |
| Total | Nickel | $\mu \mathrm{g} / \mathrm{L}$ | --- | 3 |  | 50.2 |  | 50.9 |  | 37.40 |  | 76.3 |  | 38.1 |  | 36.8 |  |
| Total | Phosphorus | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Potassium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Selenium | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.3 |  | 0.5 |  | 0.495 |  | 0 |  | 0.58 |  | 0.578 |  | 0.575 |  |
| Total | Silicon | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Silver | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.5 | < | 0.05 | $<$ | 0.05 | $<$ | 0.02 | < | 0.05 |  | 0.018 |  | 0.018 |  |
| Total | Sodium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Strontium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Sulphur | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Thallium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | -- |  | --- |  |
| Total | Tin | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Titanium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Uranium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Vanadium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Zinc | $\mu \mathrm{g} / \mathrm{L}$ | --- | 3 |  | 17.2 |  | 15.15 |  | 78.60 |  | 20.7 |  | 32.9 |  | 26.4 |  |
| Total | Zirconium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |


| State | Parameter | Unit | Sewer Use Criteria |  | HartlandValveChamberFR222-Aug-2022 |  |  | Hartland Valve Chamber FRM 22-Aug-2022 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 27/Sep/2022 |  | ```Hartland Valve Chamber SS 19/Oct/2022``` |  | ```Hartland Valve Chamber FR1 30/Nov/2022``` |  | Hartland <br> Valve <br> Chamber <br> FR2 <br> 30/Nov/2022 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| POLYCYCLIC AROMATIC HYDROCARBONS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Total PAHs | $\mu \mathrm{g} / \mathrm{L}$ |  | 0.05 |  | 1.20 |  | 1.20 |  | 3.50 |  | 34.00 |  | 2.80 |  | 2.30 |  |
| LOW WEIGHT |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 2-Chloronaphthalene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | 2-Methylnaphthalene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Acenaphthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.33 |  | 0.315 |  | 1.40 |  | 6.1 |  | 0.89 |  | 0.63 |  |
| Total | Acenaphthylene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.015 |  | 0.0155 | < | 0.05 |  | 0.068 |  | 0.015 |  | 0.02 |  |
| Total | Anthracene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.032 |  | 0.0335 |  | 0.09 |  | 0.55 |  | 0.072 |  | 0.075 |  |
| Total | Fluorene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.32 |  | 0.32 |  | 1.10 |  | 3.9 |  | 0.61 |  | 0.48 |  |
| Total | Naphthalene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.12 |  | 0.12 | < | 0.05 |  | 15 |  | 0.25 |  | 0.086 |  |
| Total | Phenanthrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.041 |  | 0.0395 |  | 0.21 |  | 1.7 |  | 0.12 |  | 0.072 |  |
| Total | Total Lmw-Pah'S | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| HIGH WEIGHT |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Benzo(A)Anthracene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.026 |  | 0.024 | < | 0.05 |  | 0.084 |  | 0.022 |  | 0.043 |  |
| Total | Benzo(A)Pyrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.005 |  | 0.0058 | < | 0.03 |  | 0.019 | $<$ | 0.005 |  | 0.0062 |  |
| Total | Benzo(B)Fluoranthene + Benzo(J)Fluoranthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.01 | $<$ | 0.01 | $<$ | 0.05 |  | 0.022 | - | 0.01 | $<$ | 0.01 |  |
| Total | Benzo(G,H,I)Perylene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.10 | $<$ | 0.02 | $<$ | 0.02 | < | 0.02 |  |
| Total | Benzo(K)Fluoranthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.01 | $<$ | 0.01 | $<$ | 0.05 | $<$ | 0.01 | $<$ | 0.01 | $<$ | 0.01 |  |
| Total | Chrysene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.01 |  | 0.0155 | $<$ | 0.05 |  | 0.064 |  | 0.024 |  | 0.046 |  |
| Total | Dibenzo(A,H)Anthracene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | < | 0.02 | < | 0.02 | $<$ | 0.10 | $<$ | 0.02 | $<$ | 0.02 | < | 0.02 |  |
| Total | Fluoranthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.13 |  | 0.15 |  | 0.29 |  | 0.8 |  | 0.36 |  | 0.45 |  |
| Total | Indeno(1,2,3-C,D)Pyrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.02 | $<$ | 0.02 | < | 0.10 | $<$ | 0.02 | $<$ | 0.02 | < | 0.02 |  |
| Total | Pyrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.11 |  | 0.125 |  | 0.20 |  | 0.53 |  | 0.25 |  | 0.33 |  |
| Total | Total Hmw-Pah'S | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| VOLATILE ORGANICS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Dissolved | Benzene | $\mu \mathrm{g} / \mathrm{L}$ | --- | 100 | < | 0.4 | < | 0.4 | < | 0.40 |  | 0.64 | < | 0.4 | < | 0.4 |  |
| Dissolved | Ethylbenzene | $\mu \mathrm{g} / \mathrm{L}$ | --- | 200 | < | 0.4 | $<$ | 0.4 | < | 0.40 |  | 0.88 | $<$ | 0.4 | $<$ | 0.4 |  |
| Total | M \& P Xylenes | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Methyl Tertiary Butyl Ether | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | O-Xylene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | -- |  | --- |  |
| Total | Styrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Dissolved | Toluene | $\mu \mathrm{g} / \mathrm{L}$ | --- | 200 | $<$ | 0.4 | $<$ | 0.4 |  | 0.53 |  | 0.58 |  | 0.61 |  | 0.52 |  |
| Dissolved | Xylenes | $\mu \mathrm{g} / \mathrm{L}$ | --- | 200 |  | 0.71 |  | 0.745 |  | 0.54 |  | 2.7 | $<$ | 0.4 | $<$ | 0.4 |  |

Notes

- Exceeded maximum allowable value specified in CRD Sewer Use Bylaw 2922
*     - Exceedances are due to compromised/expired preservatives, and results are not representati
$---=$ Not available.
FR1 - Field replicate 1.
FR2 - Field replicate 2.
FR3 - Field replicate 3.
FRM - Mean of field replicates
SS - Single Sample
NS- Not sampled

\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline State \& Parameter \& Unit \& \begin{tabular}{c} 
Sewer \\
\hline min
\end{tabular} \& riteria

max \& \& Hartland Valve Chamber FRM 30/Nov/2022 \& \& Hartland
Valve
Chamber
SS

28/Dec/2022 \& \& \begin{tabular}{|c|}
\hline Hartland <br>
Valve <br>
Chamber <br>
SS <br>
17/Jan/2023 <br>
\hline

 \& \& 

\hline Hartland <br>
Valve <br>
Chamber <br>
FR1 <br>
14/Feb/2023 <br>
\hline

 \& \& 

\hline Hartland <br>
Valve <br>
Chamber <br>
FR2 <br>
14/Feb/2023 <br>
\hline

 \& \& 

\hline Hartland <br>
Valve <br>
Chamber <br>
FRM <br>
14/Feb/2023 <br>
\hline

 \& \& 

\hline Hartland <br>
Valve <br>
Chamber <br>
SS <br>
<br>
23/Mar/2023
\end{tabular} \& <br>

\hline \multicolumn{20}{|l|}{CONVENTIONALS} <br>
\hline Total \& Alkalinity - Total - Ph 4.5 \& mg/L \& --- \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& <br>
\hline Total \& BOD \& mg/L \& --- \& 500 \& \& 17.5 \& \& 18. \& \& 24. \& \& 17. \& \& 18. \& \& 17.5 \& \& 15. \& <br>
\hline Total \& CBOD \& mg/L \& --- \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& <br>
\hline Total \& Chloride \& $\mathrm{mg} / \mathrm{L}$ \& --- \& 1500 \& \& 270. \& \& 200. \& \& 270. \& \& 250. \& \& 250. \& \& 250. \& \& 260. \& <br>
\hline Total \& COD \& mg/L \& --- \& 1000 \& \& 351.5 \& \& 315. \& \& 415. \& \& 331. \& \& 346. \& \& 338.5 \& \& 398. \& <br>
\hline Total \& Conductivity \& $\mu \mathrm{S} / \mathrm{cm}$ \& --- \& --- \& \& 2614. \& \& 2303. \& \& 3113. \& \& 2783. \& \& 2783. \& \& 2783. \& \& 2942. \& <br>
\hline Total \& Cyanide SAD \& mg/L \& --- \& 1 \& \& 0.0149 \& \& 0.014 \& \& 0.0222 \& \& 0.0133 \& \& 0.0084 \& \& 0.01085 \& \& 0.0112 \& <br>
\hline Total \& Cyanide WAD \& mg/L \& --- \& 1 \& \& 0.0053 \& < \& 0.0025 \& \& 0.0077 \& \& 0.0071 \& < \& 0.005 \& \& 0.00605 \& \& 0.0028 \& <br>
\hline Total \& Dissolved Oxygen \& mg/L \& --- \& --- \& \& 1.09 \& \& 0.16 \& \& 2.7 \& \& 0.81 \& \& 0.81 \& \& 0.81 \& \& 1.02 \& <br>
\hline Total \& Fecal Coliforms \& CFU/100 mL \& --- \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& <br>
\hline Total \& N - Ammonia (As N) \& mg/L \& --- \& --- \& \& 205. \& \& 160. \& \& 270. \& \& 210. \& \& 210. \& \& 210. \& \& 240. \& <br>
\hline Total \& N - Nitrite (As N) \& $\mathrm{mg} / \mathrm{L}$ \& --- \& --- \& \& 3.225 \& \& 1.47 \& \& 1.68 \& \& 1.84 \& \& 1.84 \& \& 1.84 \& \& 1.34 \& <br>
\hline Total \& N - Nitrate (As N) \& mg/L \& --- \& --- \& \& 24.7 \& \& 15.9 \& \& 9.96 \& \& 13.7 \& \& 13.7 \& \& 13.7 \& \& 17.1 \& <br>
\hline Total \& N - Nitrite+Nitrite (As N) \& mg/L \& --- \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& <br>
\hline Total \& N-TKN (As N) \& mg/L \& --- \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& <br>
\hline Total \& N - Total (As N) \& mg/L \& --- \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& <br>
\hline Total \& Oil \& Grease, Mineral \& $\mathrm{mg} / \mathrm{L}$ \& --- \& 15 \& $<$ \& 2. \& $<$ \& 2. \& < \& 2. \& < \& 2. \& $<$ \& 2. \& < \& 2. \& < \& 2. \& <br>
\hline Total \& Oil \& grease, total \& $\mathrm{mg} / \mathrm{L}$ \& --- \& 100 \& \& 1.1 \& < \& 1. \& \& 1.3 \& \& 3.8 \& < \& 1. \& \& 2.4 \& \& 1.4 \& <br>
\hline Total \& ORP \& mV \& --- \& --- \& \& - 14. \& \& - 298. \& \& - 150.8 \& \& 80.8 \& \& 80.8 \& \& 80.8 \& \& 132.4 \& <br>
\hline Total \& pH \& pH \& 5.5 \& 11 \& \& 7.45 \& \& 7.36 \& \& 8.1 \& \& 7.82 \& \& 7.82 \& \& 7.82 \& \& 8. \& <br>
\hline Dissolved \& Sulphide \& mg/L \& --- \& 1 \& \& 0.019 \& \& --- \& \& 0.023 \& \& 0.042 \& \& 0.034 \& \& 0.038 \& \& 0.031 \& <br>
\hline Total \& Sulphide \& mg/L \& --- \& 1 \& $<$ \& 0.036 \& < \& 0.0036 \& \& 0.032 \& \& 0.045 \& \& 0.03 \& \& 0.0375 \& \& 0.029 \& <br>
\hline \multirow[t]{2}{*}{Dissolved} \& Sulphate \& $\mathrm{mg} / \mathrm{L}$ \& --- \& 1500 \& \& 220. \& \& 120. \& \& 110. \& \& 100. \& \& 110. \& \& 105. \& \& 100. \& <br>
\hline \& Temperature \& ${ }^{\circ} \mathrm{C}$ \& --- \& --- \& \& 9.1 \& \& 12.6 \& \& 14.3 \& \& 13.4 \& \& 13.4 \& \& 13.4 \& \& 13.3 \& <br>
\hline Total \& TOC \& mg/L \& --- \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& \& --- \& <br>
\hline Total \& Total Phenols \& mg/L \& --- \& 1 \& $<$ \& 0.0075 \& $<$ \& 0.0075 \& \& 0.0066 \& $<$ \& 0.0015 \& < \& 0.0015 \& \& 0.0015 \& $<$ \& 0.038 \& <br>
\hline Total \& TSS \& mg/L \& --- \& 350 \& \& 54.5 \& \& 12. \& \& 15. \& \& 13. \& \& 10. \& \& 11.5 \& \& 20. \& <br>
\hline
\end{tabular}

| State | Parameter | Unit | Sewer min | max |  | Hartland Valve Chamber FRM $30 /$ Nov/2022 | Hartland <br> Valve <br> Chamber <br> SS <br> 28/Dec/2022 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 17/Jan/2023 |  | Hartland Valve Chamber FR1 14/Feb/2023 |  | Hartland <br> Valve <br> Chamber <br> FR2 <br> 14/Feb/2023 |  | Hartland <br> Valve <br> Chamber <br> FRM <br> 14/Feb/2023 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 23/Mar/2023 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| metals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Aluminum | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Antimony | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Arsenic | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.4 |  | 5.495 | 5.65 |  | 7.04 |  | 7.6 |  | 4.99 |  | 6.295 |  | 5.37 |  |
| Total | Barium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Beryllium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Bismuth | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Boron | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Cadmium | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.3 |  | 0.08045 | 0.0781 |  | 0.0781 |  | 0.079 |  | 0.05 |  | 0.0645 |  | 0.073 |  |
| Total | Calcium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Chromium | $\mu \mathrm{g} / \mathrm{L}$ | --- | 4 |  | 34.6 | 32.2 |  | 43 |  | 55.3 |  | 37 |  | 46.15 |  | 51.9 |  |
| Total | Chromium III | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Chromium Vi | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Cobalt | $\mu \mathrm{g} / \mathrm{L}$ | --- | 5 |  | 13.2 | 10.6 |  | 12.9 |  | 15.5 |  | 10.5 |  | 13 |  | 12 |  |
| Total | Copper | $\mu \mathrm{g} / \mathrm{L}$ | --- | 1 |  | 22.95 | 29.4 |  | 19.3 |  | 21.4 |  | 14 |  | 17.7 |  | 24 |  |
| Total | Hardness (As Caco3) | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Iron | $\mu \mathrm{g} / \mathrm{L}$ | --- | 50 |  | 3370 | 1580 |  | 2220 |  | 2610 |  | 1690 |  | 2150 |  | 1650 |  |
| Total | Lead | $\mu \mathrm{g} / \mathrm{L}$ | --- | 1 |  | 1.34 | 0.703 |  | 0.751 |  | 0.765 |  | 1.03 |  | 0.8975 |  | 0.529 |  |
| Total | Lithium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Magnesium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Manganese | $\mu \mathrm{g} / \mathrm{L}$ | --- | 5 |  | 645 | 480 |  | 640 |  | 855 |  | 569 |  | 712 |  | 591 |  |
| Total | Mercury | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.02 | $<$ | 0.019 | 0.005 | $<$ | 0.038 | $<$ | 0.038 | $<$ | 0.038 | $<$ | 0.038 | $<$ | 0.038 |  |
| Total | Molybdenum | $\mu \mathrm{g} / \mathrm{L}$ | --- | 5 |  | 5.26 | 4.54 |  | 2.95 |  | 3.86 |  | 2.52 |  | 3.19 |  | 16.1 |  |
| Total | Nickel | $\mu \mathrm{g} / \mathrm{L}$ | --- | 3 |  | 37.45 | 29.2 |  | 37.5 |  | 45.7 |  | 30.3 |  | 38 |  | 89.7 |  |
| Total | Phosphorus | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Potassium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Selenium | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.3 |  | 0.5765 | 0.616 |  | 0.558 |  | 0.74 |  | 0.495 |  | 0.6175 |  | 0.569 |  |
| Total | Silicon | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Silver | $\mu \mathrm{g} / \mathrm{L}$ | --- | 0.5 |  | 0.018 | 0.024 |  | 0.015 | $<$ | 0.02 | < | 0.02 | $<$ | 0.02 | $<$ | 0.02 |  |
| Total | Sodium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Strontium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Sulphur | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | -- |  | --- |  | --- |  |
| Total | Thallium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | -- |  | --- |  | --- |  |
| Total | Tin | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Titanium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Uranium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Vanadium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Zinc | $\mu \mathrm{g} / \mathrm{L}$ | --- | 3 |  | 29.65 | 21.7 |  | 15.5 |  | 19.6 |  | 11.8 |  | 15.7 |  | 15.7 |  |
| Total | Zirconium | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |


| State | Parameter | Unit | Sewer Use Criteria |  | Hartland <br> Valve <br> Chamber <br> FRM <br> 30/Nov/2022 |  |  | Hartland <br> Valve <br> Chamber <br> SS <br> 28/Dec/2022 |  | Hartland <br> Valve <br> Chamber <br> SS <br> 17/Jan/2023 |  | Hartland <br> Valve <br> Chamber <br> FR1 <br> 14/Feb/2023 |  | Hartland Valve Chamber FR2 14/Feb/2023 |  | Hartland Valve Chamber FRM 14/Feb/2023 |  | Hartland Valve Chamber SS 23/Mar/2023 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Total PAHs | $\mu \mathrm{g} / \mathrm{L}$ |  | 0.05 |  | 2.55 |  | 1.80 |  | 5.80 |  | 8.30 |  | 8.00 |  | 8.15 |  | 5.40 |  |
| LOW WEIGHT |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 2-Chloronaphthalene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | 2-Methylnaphthalene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Acenaphthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.76 |  | 0.27 |  | 1.9 |  | 3.3 |  | 3.2 |  | 3.25 |  | 2.3 |  |
| Total | Acenaphthylene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.0175 |  | 0.04 |  | 0.062 |  | 0.086 |  | 0.08 |  | 0.083 | < | 0.01 |  |
| Total | Anthracene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.0735 |  | 0.019 |  | 0.28 |  | 0.16 |  | 0.17 |  | 0.165 |  | 0.09 |  |
| Total | Fluorene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.545 |  | 0.2 |  | 1.1 |  | 2 |  | 2 |  | 2 |  | 1 |  |
| Total | Naphthalene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.168 |  | 0.13 |  | 0.033 |  | 0.46 |  | 0.32 |  | 0.39 |  | 0.1 |  |
| Total | Phenanthrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.096 |  | 0.023 |  | 0.083 |  | 0.35 |  | 0.38 |  | 0.365 |  | 0.23 |  |
| Total | Total Lmw-Pah'S | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| HIGH WEIGHT |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Benzo(A)Anthracene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.0325 |  | 0.055 |  | 0.12 |  | 0.057 |  | 0.055 |  | 0.056 |  | 0.019 |  |
| Total | Benzo(A)Pyrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.0056 |  | 0.01 |  | 0.036 |  | 0.012 |  | 0.013 |  | 0.0125 | $<$ | 0.005 |  |
| Total | Benzo(B)Fluoranthene + Benzo(J)Fluoranthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.01 |  | 0.013 |  | 0.052 |  | 0.013 |  | 0.012 |  | 0.0125 | $<$ | 0.01 |  |
| Total | Benzo(G,H,I)Perylene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 |  |
| Total | Benzo(K)Fluoranthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | < | 0.01 | $<$ | 0.01 |  | 0.018 | $<$ | 0.01 | $<$ | 0.01 | - | 0.01 | $<$ | 0.01 |  |
| Total | Chrysene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.035 |  | 0.053 |  | 0.11 |  | 0.055 |  | 0.052 |  | 0.0535 |  | 0.019 |  |
| Total | Dibenzo(A,H)Anthracene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 |  |
| Total | Fluoranthene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.405 |  | 0.52 |  | 1.1 |  | 0.66 |  | 0.62 |  | 0.64 |  | 0.19 |  |
| Total | Indeno(1,2,3-C,D)Pyrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 | $<$ | 0.02 |  |
| Total | Pyrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | 0.29 |  | 0.39 |  | 0.79 |  | 0.45 |  | 0.42 |  | 0.435 |  | 0.11 |  |
| Total | Total Hmw-Pah'S | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| VOLATILE ORGANICS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \| Dissolved | Benzene | $\mu \mathrm{g} / \mathrm{L}$ | --- | 100 | < | 0.4 | < | 0.4 | < | 0.4 | < | 0.4 |  | 0.44 |  | 0.42 | < | 0.4 |  |
| Dissolved | Ethylbenzene | $\mu \mathrm{g} / \mathrm{L}$ | --- | 200 | < | 0.4 | $<$ | 0.4 | $<$ | 0.4 | $<$ | 0.4 | < | 0.4 | $<$ | 0.4 | $<$ | 0.4 |  |
| Total | M \& P Xylenes | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Methyl Tertiary Butyl Ether | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | O-Xylene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Total | Styrene | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  | --- |  |
| Dissolved | Toluene | $\mu \mathrm{g} / \mathrm{L}$ | --- | 200 |  | 0.565 |  | 1.3 | $<$ | 0.4 | $<$ | 0.4 |  | 1 |  | 0.7 | $<$ | 0.4 |  |
| Dissolved | Xylenes | $\mu \mathrm{g} / \mathrm{L}$ | --- | 200 | $<$ | 0.4 |  | 0.47 | $<$ | 0.4 |  | 0.41 |  | 0.55 |  | 0.48 | $<$ | 0.4 |  |

Notes:

- Exceeded maximum allowable value specified in CRD Sewer Use Bylaw 2922
*     - Exceedances are due to compromised/expired preservatives, and results are not representati
$---=$ Not available.
FR1 - Field replicate 1
FR2 - Field replicate 2.
FR3 - Field replicate 3.
FRM - Mean of field replicates
SS - Single Sample
NS- Not sampled

B5. Quarterly Leachate Quality Trace Organics


B6. Monthly Leachate Quality Phase 2 Cleanout

| State | Parameter | Units | Sewer Use Criteria |  | Phase 2Cleanout Cleanout$\qquad$ | Phase 2 <br> Cleanou $\qquad$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | CONVENTIONALS |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Alkalinity - Total - Ph 4.5 | mgl | - |  |  |  | - |  |  |
| Total | B00 | mgl | - | 500 |  |  |  |
| Total | CBOD | mgl | - |  |  |  |  |
| Dis | Choride | mgl | - | 1500 | - | - |  |
| Total | cod | mgh | - | 1000 | - | - |  |
| Total | Conducivivy | uscm | - |  |  |  |  |
| Total | Cyande SAD | mgl |  | 1 |  |  |  |
| Tolal | Cyanide MAD | mgl | - | 1 |  |  |  |
| Tolal | Dissolved Oxygen | mgh |  |  |  |  |  |
|  | 速 |  |  |  |  |  |  |
| Tolal | N-NHS (Ass) | mgit |  |  |  |  |  |
| Total | N- NO3 Ass N | mgl |  | - |  |  |  |
| Total | N- $\mathrm{NO} 3+\mathrm{NO2}(\mathrm{As} \mathrm{N})$ | mgl |  |  |  |  |  |
| Total | $\mathrm{N}-\mathrm{Tkn}(\mathrm{AsN})$ | mgl |  | - |  |  |  |
| Total | N - Total (As N) | mgl | $\cdots$ | - |  |  |  |
| Total | Oil \& Grease, Mineral | mgl | - | 15 |  |  |  |
| Total | Oil \& grease, total | mgl | - | 100 |  |  |  |
| Toal |  | mv |  |  |  |  |  |
| Total | pH | pH | 5.5 | 11 | - | - |  |
| Dissolved | Suphide | mgl |  | 1 | - | - |  |
| Total | Suphide | mgl | - | 1 |  |  |  |
| Dissolved | suphate | mgl |  | ${ }^{1500}$ | - | - |  |
| Total | Temperature |  | - |  |  |  |  |
| Toal | ${ }^{\text {Tocas }}$ | ${ }_{\text {mgl }}^{\text {mat }}$ | - | - | - | - |  |
| Toal | ${ }_{\text {Tss }}$ | mal | - | ${ }_{350}$ | - | - |  |








[^2]B7. Monthly Leachate Quality North Purge Wells



*     - Exxeedances are due to compromisedlexpired presenatives, and results are not representative
s . Single
anmole

| SS- Single Sample |
| :---: |
| NS - Not Sampled |

B8. Monthly Leachate Quality Controlled Waste Drainage

| States | Parameter | Units | $\begin{array}{\|c\|} \hline \text { Sewer Us } \\ \hline \text { min } \end{array}$ | ${ }_{\text {max }}$ |  | $\begin{gathered} \text { Controled } \\ \text { Waste Ditch } \\ \text { SS } \\ \text { 31-May-2022 } \end{gathered}$ | $\begin{aligned} & \text { Controled } \\ & \text { Watied ith } \\ & \text { ss } \\ & \text { 27-Jun-2022 } \end{aligned}$ | $\begin{gathered} \text { Contolelen } \\ \text { Waste itach } \\ \text { Ns } \\ \text { 27-Jul-2022 } \end{gathered}$ | $\begin{array}{\|c} \hline \begin{array}{c} \text { Controlled } \\ \text { Waste Ditch } \\ \text { Ns } \\ \text { 22-Aug-2022 } \end{array} \\ \hline \end{array}$ | $\begin{array}{\|l} \hline \text { Controlled } \\ \text { Waste Ditch } \\ \text { Ss } \\ \text { 27-Sep-20222 } \end{array}$ | $\begin{array}{\|c} \hline \text { Controled } \\ \text { Waste Ditch } \\ \text { Ss } \\ \text { 29-Oct-2021 } \end{array}$ | Controlled Waste Ditch Waste Ditch 30-Nov-2022 | $\underset{\substack{\text { Controled } \\ \text { Waste Dich }}}{ }$ <br> 28-DeC-2022 |  | $\begin{array}{\|c} \hline \text { Controlled } \\ \text { Waste Ditch } \\ \text { ss } \\ \text { 14-Feb-2023 } \end{array}$ | $\begin{array}{\|c} \begin{array}{c} \text { Controlled } \\ \text { Waste Ditch } \end{array} \\ \text { ss } \\ \text { 23-Mar-2023 } \\ \hline \end{array}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{\text {Total }}$ Total |  | ${ }_{\text {mgh }}^{\text {mal }}$ | $\cdots$ | $\stackrel{-}{500}$ | $\stackrel{-}{60}$ | $\stackrel{-}{83}$ | $\stackrel{-}{72}$ | $\stackrel{-}{37}$ | $\frac{-170}{170}$ | 3 | $\stackrel{-1}{19}$ | $\cdots$ | - | $\stackrel{-}{34}$ | $\stackrel{-}{22}$ | $\stackrel{-}{65}$ |  |
| Toal | ${ }_{\text {cbob }}$ | mgh |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Dissolved | Chloide | mgh | - | 1500 | 390 | 480 | 500 | 600 | 620 | 670 | 690 |  |  | 280 | 260 | 450 |  |
| Total | ${ }^{\text {cod }}$ | mgh | - | 1000 | ${ }_{396}^{390}$ | ${ }_{5196}^{498}$ | ${ }_{503}^{565}$ | ${ }^{1390}$ | ${ }_{7733}$ | ${ }_{681}^{6513}$ | 778 | - | - | 312 | ${ }^{275}$ | 545 |  |
| Total | Conductivity | ${ }_{\text {uscm }}$ | $\cdots$ | 1 | ${ }_{\substack{3800 \\ 0.0086}}$ | 5158 | ${ }^{6256}$ | 6610 | 7433 0.011 | ${ }_{0}^{6543}$ | ${ }_{6}^{654}$ |  |  | ${ }^{3230}$ | ${ }^{2572}$ | 473 |  |
| Total | Cyanide SAD | mgl | - | + | ${ }^{0.0086}$ |  |  | 0.021 |  |  |  |  |  |  |  | 0.069 |  |
| Toal | cyande WAO | mgh | - | 1 | 0.0025 | 0.0132 |  | 0.01 | 0.01 | 0.0005 | ${ }^{0.0064}$ |  |  | 0.0029 | 0.0042 | 0.0025 |  |
| Total | solved Oxygen | mgh | - | -- | 5.23 | 4.26 | 6.1 |  |  |  | 52 | - | - | 7.77 | 7.14 | 4.4 |  |
| Total | Feal Colifoms | 100 mL |  | $\cdots$ |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Toal | N-NH3 (As N) | mgl | - | $\cdots$ | 260 | 380 | 450 | 480 | 510 | 610 | 580 |  |  | ${ }^{230}$ | 190 | 410 |  |
| Toal | N- No2 (As N) | mgl | - | $\cdots$ | ${ }^{3.73}$ | ${ }^{9.75}$ | 0.311 | ${ }^{9.55}$ | 11.9 | 11 | 12.1 | - | - | 2.82 | 2.72 | 3.71 |  |
| Toal | N- No3 (As N) | mgl | -- | -- | 13 | ${ }^{3.22}$ | 0.212 | 1.9 | 1.19 | 0.25 | 6.04 | - | - | 14.5 | 20.4 | 4.33 |  |
| Toal |  | mgl | $\cdots$ | $\cdots$ |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | ${ }^{N-T \mathrm{Tk}}(\mathrm{Ass}$ N) | mgh | - | - | - | - | - | - | - | - | - |  |  | - |  |  |  |
| Toal | $N$ - Toal (As N) | mgl | - | $\cdots$ | - |  | $-$ |  |  |  |  |  |  |  |  |  |  |
| $\xrightarrow{\text { Toial }}$ Toal |  | mgh | $\cdots$ | 15 100 | ${ }_{1}$ | $\stackrel{2}{1}$ | ${ }_{1}^{2}$ | ${ }_{1}$ | ${ }_{1}$ | ${ }_{1}$ | 0 |  |  | $\stackrel{2}{1}$ | ${ }_{1}$ | ${ }^{2}$ |  |
| ${ }_{\text {Toal }}^{\text {Toal }}$ | ${ }^{\text {orp }}$ Oing grease, olaial | ${ }_{\text {mg }}^{\text {mi }}$ | -- |  | ${ }_{-1.0}$ | ${ }_{35.0}$ | ${ }_{23.0}$ | ${ }^{-1.0}$ | 21.0 | 109.0 | ${ }_{132.0}$ |  | - | ${ }_{126.7}$ | 75.0 | 119.3 |  |
| Total | pH | pH | ${ }^{5.5}$ | 11 | ${ }^{8.23}$ | 8.26 | 8.26 | 8.17 | 8.59 | 8.55 | 8.15 |  |  | 8.22 | ${ }^{8.4}$ | 8.69 |  |
| Dissolved |  | mgl |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Suphide | mgh | $\cdots$ | 15 | ${ }^{0.032}$ | 0.019 | 0.009 | 0.009 | 0.075 | 0.009 | 0.036 |  |  | 0.036 | 0.036 | 0.036 |  |
| Dissoved | Suphate | mgh |  |  | ${ }_{10}^{10}$ | -187 | ${ }^{10}$ |  | 22 | ${ }^{25}$ | 16 | - | - |  |  |  |  |
| Toal |  |  | - | , |  | 10. |  | 2.1 |  |  | 6.6 |  |  | 16 | 14.4 | 15.9 |  |
| Toial | Total Phenols | mgh | - | 1 | 0 | 0.015 | 0.015 | 0.031 | 0.017 | ${ }_{0}^{0.0083}$ | 0.0075 | - | - | ${ }^{0.0057}$ | 0.015 | 0.015 |  |
| Toal | Tss | mgh | -- | 350 | ${ }_{5} 5$ | ${ }^{8.8}$ | 24 | 11 | 280 | 15 | 6 | $\cdots$ | - | 2.4 | 6.8 | 31 |  |
| metals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Aluminum | Hgh | -- | - | -- | -- | - | - | - | - | -- |  |  | - | - | - |  |
| Total | Anituony | gigh | $\underline{-}$ | 400 | 4.11 | 4.99 | 5.95 | 5.81 | 8.71 | 6.93 | 6.23 | - | - | 3.68 | 3.12 | 5.89 |  |
| Total | Barium | ugh | - | - | $\cdots$ | $\cdots$ | $\cdots$ | $\cdots$ | $\cdots$ | $\cdots$ | $\cdots$ | - | - | $\cdots$ | $\cdots$ | $\cdots$ |  |
| Total |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Tooal | ${ }^{\text {Broron }}$ | ${ }_{\text {H992 }}$ | $\cdots$ | $\cdots$ | -- | - | - | - | -- | - | - | - | - | - | -- | - |  |
| Total | Casmium | H9\% | - | 300 | 0.061 | 0.025 | 0.025 | 0.025 | 0.102 | 0.072 | 0.025 | - | - | 0.0524 | 0.061 | 0.035 |  |
| Total | Calcium | нgh |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Toal | chromium | mgh | $\cdots$ | 4000 | 32.9 | 34.8 | 39.8 | 41.7 | 66.7 | 46.9 | 46.3 |  | - | 21.6 | 20.9 | 36.2 |  |
| Toal | Chromium III | нgh | $\cdots$ | - |  | - |  |  |  |  |  | - | - |  |  |  |  |
| Total | Chromium Vi | mgh | - | - | - | - | - | - | $\stackrel{-}{275}$ | - | $\cdots$ | - |  | - | -1 | $\stackrel{-}{187}$ |  |
| Total | Cobat | нgh | $\cdots$ | 5000 | 16.4 | 18.2 <br> 2.1 | ${ }^{21.1 .1}$ | ${ }^{22.3}$ | $\stackrel{27.5}{1.6}$ | ${ }^{29.2}$ | ${ }^{28.2}$ |  |  | ${ }_{12}^{12.3}$ | 13 | ${ }_{5}^{18.7}$ |  |
| Total | ${ }_{\text {copper }}^{\text {Hardoss (As Caco3) }}$ | \%gh | $\cdots$ |  | 10.6 | 2.19 | 1.42 | 1.53 | 16.6 | 1.91 | 0.92 |  |  | 12.3 | ${ }^{13}$ | 5.4 |  |
| Total | ron | $\mu \mathrm{g}$ / | - | 5000 | 1990 | 3100 | 3970 | 2680 | 22200 | 2710 | 2580 | - | - | ${ }^{1440}$ | 1470 | 7020 |  |
| Toal | Lead | wgh | -- | 1000 | 0.232 | 0.18 | 0.19 | 0.28 | 2 | 0.19 | 0.13 |  | - | 0.218 | 0.227 | 0.497 |  |
| Toal | Lihum | rgh | $\cdots$ | - |  | - |  |  |  |  |  |  |  |  |  |  |  |
| Toal | Manesame | ${ }_{\text {eqg }}$ | - | 5000 | 1700 | 443 | 469 |  | 1150 |  | 113 |  | - | 155 | 320 |  |  |
| Toal | Mergury |  | -- | ${ }_{20} 20$ | 0.019 | ${ }_{0}^{4.038}$ | $\stackrel{4098}{0.038}$ | ${ }_{0}^{0.019}$ | 0.019 | 0.019 | 0.038 | - | - | 0.038 | ${ }_{0}^{0.038}$ | ${ }_{0} 0.038$ |  |
| Total | Molvodenum | g92 | - | 5000 | ${ }_{1}^{1.24}$ | 1.39 | 1.69 |  |  | 1.87 | ${ }_{1.57}$ |  |  |  | 1.05 | ${ }_{3,36}$ |  |
| Total | Nickel | rgh | - | 3000 | 36.7 | ${ }^{37.7}$ | 42.9 | 46.9 | 55.7 | 57.4 | 55.1 | - | - | 32.4 | 28.2 | 46.8 |  |
| Total | Phosshorus | нgl | - |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Pooassium | mgh |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Selenium | rgh | $\cdots$ | 300 | 0.43 | 0.53 | 0.59 | 0.56 | 0.76 | 0.75 | 0.62 | - | - | 0.378 | 0.336 | 0.57 |  |
| Toal | Silicon | нgh | $\cdots$ |  |  |  |  |  |  |  |  | - | - |  |  |  |  |
| Total | siver | нgh | - | 500 | 0.02 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | - | - | 0.012 | 0.02 | 0.02 |  |
| $\xrightarrow{\text { Toal }}$ Toal | Stationm | $\underset{\text { Mgh }}{\text { Mgh }}$ | $\cdots$ | - |  |  |  |  |  |  |  |  |  |  |  | - |  |
| Total | Suphur | mgl | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |
| Total | Thalium | mgl | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |
| Toal | Tin | HgL | -- |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Toal | thanum | Hgh | - | - | - | - |  |  |  |  |  |  |  |  |  |  |  |
| ${ }_{\text {Tolal }}^{\text {Toal }}$ | Vanadium |  | $\cdots$ | $\cdots$ | $\cdots$ | - | - | - | -- | -- | $\cdots$ | -- | - | - | $\cdots$ | - |  |
| Total | Zinc | нgh | - | 3000 | 49.9 | 56.6 | 51.4 | 52.7 | 395 | 39.5 | 39.7 | - | - | 22.2 | 29.7 | 80.3 |  |
| voLatue organics |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Dissoved | benzene | Mg/ | $\cdots$ | ${ }^{100}$ | ${ }^{0.4}$ | ${ }_{0}^{0.4}$ | ${ }^{0.4}$ | ${ }_{0}^{0.4}$ | ${ }_{0}^{0.4}$ | ${ }_{0}^{0.4}$ | ${ }_{0}^{0.4}$ | - | - | ${ }^{0.4}$ | ${ }_{0}^{0.4}$ | ${ }_{0}^{0.4}$ |  |
| Dissoved | $\frac{\text { Ethybenzene }}{\text { M\&PXVeneses }}$ |  | $\cdots$ | 200 | $\stackrel{0.4}{ }$ |  |  | $\stackrel{0.4}{+}$ | $\stackrel{0.4}{ }$ |  | $\stackrel{0}{0}$ | - | - |  | 0.4 |  |  |
| Tooal | Methy Teritiar Buty Eltee | wgh | $\cdots$ | $\cdots$ | - | - | - | - | - | -- | - | $\cdots$ | $\cdots$ | - | - | - |  |
| Total | O-xylene | wg/ | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |
| ${ }_{\text {L }}^{\text {Tisalal }}$ | Styene | $\xrightarrow{\text { Hgh }}$ | $\cdots$ | $\stackrel{-}{200}$ | $\stackrel{-}{0 .}$ | $\stackrel{-}{0 .}$ | 0.4 | 0.64 | $\stackrel{-}{0 .}$ | $\stackrel{-}{0 .}$ | $\stackrel{-}{0 .}$ | - | - | $\stackrel{-}{0 .}$ | $\stackrel{-}{0.53}$ | $\stackrel{-}{0 .}$ |  |
| Dissolved | Xxvenes | Mg | - | 200 | 0.4 | 0.4 | 0.4 | 0.94 | 0.4 | 0.4 | 0.4 | - |  | 0.4 | 1.7 | 0.4 |  |
|  | d maximum allowable valu dun to comprom Sample | specified in C sed/expired p $\qquad$ | $R D$ Sewe $R D$ Sewer preservativ | Use Bylaw Use Bylaw ves, and | 2922 <br> sults are not rep | Entative |  |  |  |  |  |  |  |  |  |  |  |

B9. Monthly Leachate Quality Markham Valve Chamber

| State | Parameter | Units | Sewer U <br> min | $\underset{\text { max }}{ }$ | $\begin{gathered} \text { Sout Purge } \\ \text { Wells } \\ \text { ss } \\ 26 \text {-Ap-2022 } \end{gathered}$ |  | $\left.\begin{array}{\|c} \text { Sout Purge } \\ \text { Wells } \\ \text { ss } \\ \text { 31-May-2022 } \end{array} \right\rvert\,$ | South Purge Wells SS 27-Jun-2022 |  | South Purge wels ss $27-$-ul-2022 |  | $\begin{array}{\|c} \text { South Purge } \\ \text { Wels } \\ \text { ss } \\ \text { 22-Aug-2022 } \end{array}$ |  | $\begin{array}{\|c} \begin{array}{c} \text { South Purge } \\ \text { WWIs } \\ \text { ss } \\ \text { 27-Sep-2022 } \end{array} \\ \hline \end{array}$ | South Purge Wells SS 28-Oct-2022 | South Purge Wells <br> 30-Nov-2022 | South Purge Wells <br> 28-Dec-2022 | $\begin{aligned} & \text { South Purge } \\ & \text { Wells } \\ & \text { SS } \\ & \text { 17-Jan-2023 } \end{aligned}$ |  |  |  | $\begin{array}{\|c} \hline \text { South purge } \\ \text { well } \\ \text { ss } \\ \text { 23-Mar-2033 } \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CONVENTIONALS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | ${ }_{\text {CBOD }}^{\text {BOD }}$ | mgh | $\cdots$ | 500 | 54 |  | ${ }^{3.6}$ | 3 |  | 9.3 |  | ${ }^{6.8}$ |  | 6 | 5.8 | - | - | 5.9 |  | 4.6 |  | 2.5 |
| $\xrightarrow{\text { Total }}$ Total | ${ }_{\text {cteo }}^{\text {cboo }}$ | mgh | $\cdots$ |  |  |  |  |  |  |  |  |  |  |  |  | $\cdots$ | $\cdots$ |  |  |  |  |  |
| $\xrightarrow{\text { Total }}$ Total | ${ }_{\text {choride }}$ | mgh | -- | 1500 <br>  <br> 1000 <br> 1 | 490 <br> 9000 <br> 0 |  | ${ }_{75}^{110}$ | ${ }_{84}^{120}$ |  | $\begin{array}{r}150 \\ 1010 \\ \hline 101\end{array}$ |  | 150 5510 |  | 160 145 | 160 <br> 5710 | - | - | ${ }_{68}^{110}$ |  | ${ }_{62}^{88}$ |  | 78 59 |
| ${ }_{\text {Total }}^{\text {Toala }}$ | ${ }^{\text {cood }}$ Conductivity | ${ }_{\text {mgla }}$ | - | 1000 | ${ }_{19090}^{1298}$ |  | 75 1331 | 84 1059 |  | 1010 1849 |  | ${ }^{5510}$ |  | ${ }_{185}^{1873}$ | 19710 | $\cdots$ | $\cdots$ | $\stackrel{68}{1431}$ |  | $\frac{62}{1127}$ |  | ${ }_{1}^{1145}$ |
| Total | Cyanide SAD | mgl | -- | 1 | 0.047 |  | ${ }^{0.00163}$ | 0.00189 |  | 0.00284 |  | ${ }^{0.00181}$ |  | ${ }^{0.00372}$ | ${ }^{0.00155}$ | - | - | ${ }^{0.00059}$ |  | ${ }^{0.00108}$ |  | ${ }^{0.00118}$ |
| Total | Cyanide WAD | mgl | $\cdots$ | 1 | 0.016 |  | 0.00078 |  |  | 0.00087 | < | 0.0005 |  | 0.00163 | 0.00169 | - | - | 0.0005 |  | 0.0005 | - |  |
| Total | Dissolved Oxygen | mgh | - | $\cdots$ | 4.38 |  | 3.64 | 4.5 |  |  |  |  |  | -- | 42 | - | - | 6.95 |  | 38.4 |  | 3.87 |
| $\xrightarrow{\text { Total }}$ Total | Fecal Colitoms $N-N+3$ ( As $N$ N |  |  |  | 250 |  | 44 | 49 |  | 61 |  | 64 |  | 69 | 64 |  |  | 57 |  | 45 |  | 45 |
| $\underset{\text { Tolal }}{\text { Toal }}$ |  | mgil | $\cdots$ | - | ${ }_{427}^{250}$ |  | 0.005 | 0.005 |  | 0.0053 |  | 0.005 |  | 0.005 | 0.0475 |  |  | ${ }_{0}^{0.0084}$ |  | 0.0058 | < | ${ }_{0} 0.005$ |
| Total | N-NO3 (As N) | mgh | - | - | 65 | - | 0.02 | 0.02 | < | 0.02 | - | 0.02 | < | 0.02 | 0.02 | - | - | 0.026 |  | 0.02 |  | 0.034 |
| Total | $\mathrm{N}-\mathrm{NO}^{+}+\mathrm{No2}$ ( As N) | mgl | - | - | -- |  | -- |  |  |  |  | -- |  |  |  | - | - | $\cdots$ |  | $\cdots$ |  | - |
| Total | N-Tkn (As N$)$ | mgl | -- | - | -- |  | - | - |  | - |  | - |  | - | - |  |  | - |  | - |  | - |
| $\xrightarrow{\text { Total }}$ Total | N- Total (As N) | mgh | $\cdots$ |  |  |  | $\cdots$ |  |  | $\cdots$ |  | $\stackrel{\square}{2}$ |  | $\cdots$ | - |  |  | $\cdots$ |  | 2 |  |  |
| $\underset{\substack{\text { Total } \\ \text { Toala }}}{ }$ | Oil | mgal | - | ${ }^{15}$ | ${ }^{2}$ |  | ${ }^{2}$ | - ${ }^{1}$ |  | ${ }^{2}$ | < | 1 | < | 2 | $\times \quad 1$ |  |  | ${ }^{2}$ |  | ${ }^{2}$ |  | 2 |
| Total | ORP | mv |  |  | -126 |  | -107 | -124 |  | -105 |  | 19 |  | 106 | 91 | - | - | $\stackrel{.67 .1}{ }$ |  | 49.2 |  | 69.7 |
| Total | pH | pH | 5.5 | 11 | 7.28 |  | 7.15 | 7.16 |  | ${ }^{7} 43$ |  | 7.96 |  | 7.4 | 7.28 |  | - | 7.3 |  | 7.89 |  | 8.71 |
| Dissolved | Suphide | mgl |  | 1 | 0 |  | 0 |  |  |  |  |  |  |  | 0.024 |  |  |  |  | 0 |  |  |
| ${ }_{\text {Dissolved }}$ | Suphide | ${ }_{\text {mght }}^{\text {mol }}$ | $\cdots$ | $\stackrel{1}{150}$ | 0.018 100 |  | ${ }^{0.016}$ | 0.021 |  | ${ }^{0.013}$ |  | 0.017 |  | $\stackrel{0.03}{ }$ | 0.036 | $\underline{-}$ | - |  |  | ${ }_{14}$ |  | 0 |
| Dissolved | ${ }^{\text {Suphate }}$ Temoeraure | ${ }_{\text {mgh }}^{\text {che }}$ |  | $\stackrel{1500}{ }$ | 100 158 |  | $\stackrel{10}{168}$ | $\stackrel{1}{179}$ |  | $\stackrel{1}{226}$ | < | $\stackrel{1}{196}$ | < | 178 | 182 |  |  | ${ }_{14}^{3.2}$ |  | ${ }_{14}^{1.4}$ | $\underline{L}$ | $\stackrel{1}{15}$ |
| Total |  | mgl | $\cdots$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 5 |
| Total | TTotal henols | mgl | $\cdots$ | 1 | 0 |  | 0.0015 | ${ }^{0.0033}$ |  | ${ }^{0.0083}$ |  | ${ }^{0.0051}$ | < | ${ }^{0.0015}$ | ${ }^{0.0072}$ | - | - | ${ }^{0.0027}$ |  | 0.002 | < | 0.0015 |
| Total |  | mgl | - | 350 | 6 |  | 35 | 41 |  | 33 |  | 19 |  | 31 | 9.6 | - | - | 10 |  | 74 |  | 24 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\pm{ }^{\text {Total }}$ | \|Auminum | rgh | - | $\cdots$ | $\cdots$ |  | $\cdots$ | - |  | - |  | - |  | - | - | - | - | $\cdots$ |  | $\cdots$ |  | - |
| $\underset{\text { Tolal }}{\text { Toal }}$ | ${ }^{\text {Anitmony }}$ Asenic | ${ }_{\text {Hght }}^{\text {Hght }}$ | $\cdots$ | $\stackrel{-}{400}$ | $\stackrel{-}{24.7}$ |  | 9.13 | 8.94 |  | 9.73 |  | 9.04 |  | 5.42 | 5.09 | $\cdots$ | -- | 4.67 |  | 7.54 |  | 3.15 |
| Total | Baium | нğ | - |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Berylium | нgh | -- | -- | - |  | - | -- |  | -- |  | - |  | - | - | - |  | -- |  | - |  | - |
| Total | Bismuth | rgh | $\cdots$ | -- | - |  | - | -- |  | - |  | - |  | - | - | - |  | -- |  | - |  |  |
| Total | Boron | rgh | - |  |  |  |  |  |  |  |  |  |  |  |  | - |  |  |  |  |  |  |
| Total | Caamium | нgh | - | 300 | 0.212 |  | 0.01 | 0.005 |  | 0.019 | < | 0.01 | < | 0.01 | 0.15 | - | - | ${ }^{0.0128}$ |  | 0.01 |  | 0.0394 |
| $\xrightarrow{\text { Total }}$ Total | Calcium | нgh | -- |  | 125 |  | 1 | $\cdots$ |  | $\cdots$ |  |  |  |  |  | $\cdots$ | - |  |  | 0 |  |  |
| $\xrightarrow{\text { Tolal }}$ Toal | ${ }_{\text {chen }}^{\text {chromium }}$ | ${ }_{\text {Lggh }}^{\text {Hgh }}$ | $\cdots$ | 4000 | 125 |  | $\underline{-}$ | 0.92 |  | 1.22 |  | 1.17 |  | $\stackrel{0}{0}$ | 1.57 | - | - | 0.97 |  | 0.9 |  | 0.7 |
| Total | Chromium Vi | rgh | -- | -- | - |  | - | - |  | -- |  | - |  | - | -- | - |  | -- |  | -- |  |  |
| Total | Cobat | ugh | - | 5000 | 29.2 |  | 2.23 | 2.34 |  | 3.19 |  | 3.46 |  | 2.84 | 3.59 | - | - | 2.03 |  | 2.21 |  | 1.85 |
| Total | Copper | нgh | - | 1000 | ${ }^{73.6}$ |  | 0.24 | 0.19 |  | 0.47 | < | 0.2 | < | 0.2 | 0.63 | - | - | 0.58 |  | 0.4 |  | 0.57 |
| $\xrightarrow{\text { Total }}$ Total | Harchess (As Caco3) | нgh | - |  |  |  | 18200 |  |  | $\stackrel{1200}{120}$ |  | $\stackrel{-}{1060}$ |  | $\cdots$ | $\cdots$ | - | - |  |  | - |  | 850 |
| $\underset{\text { Tooal }}{\text { Toal }}$ | Lead | $\stackrel{\text { Hgt }}{\text { get }}$ | $\cdots$ | ${ }_{10000}$ | ${ }_{2}^{19.17}$ |  | ${ }^{18200}$ | ${ }_{0}^{16.638}$ |  | ${ }_{0}^{122005}$ |  | ${ }_{0}^{10.373}$ |  | ${ }_{0}^{10.120}$ | ${ }_{0}^{10.28}$ | -- | $\cdots$ | ${ }_{0}^{10.792}$ |  | ${ }_{0}^{32100}$ |  | ${ }_{0600}^{80.69}$ |
| Total | Lithium | Hgh | -- |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Magnesium | нgh | -- |  | - |  | -- | -- |  | -- |  | - |  |  |  |  |  | - |  | - |  | - |
| Total | Manganese | нgh | - | 5000 | ${ }^{263}$ |  | 188 | 149 |  | 129 |  | 120 |  | 88.5 | 145 | - | - | 260 |  | 256 |  | 259 |
| ${ }_{\text {Total }}^{\text {Toal }}$ | Mercury | rgh | $\cdots$ | $\stackrel{20}{500}$ | ${ }^{0.029}$ |  | 0.0019 | ${ }^{0.038}$ | < | 0.0019 | < | 0.019 | < | 0.019 | 0.0019 | -- | - | 0.0019 |  | 0.0019 |  | ${ }^{0.0019}$ |
| $\underset{\substack{\text { Total } \\ \text { Total }}}{ }$ | - Morbbenum | ${ }_{\text {Hght }}^{\text {qg }}$ | - | 5000 3000 | $\stackrel{2.69}{69.6}$ |  | ${ }_{6}^{14.9}$ | 㐌.44 |  | ${ }_{7}^{20.8}$ |  | 21.4 8.14 |  | ${ }^{13.3} 6$ | ${ }_{7}^{15.1}$ | $\cdots$ | $\cdots$ | 7.79 6.94 |  | 7.38 <br> 6.52 |  |  |
| Total | Phosphorus | ggh | -- |  |  |  |  |  |  |  |  |  |  |  | $\cdots$ | - | - |  |  |  |  |  |
| Total | Potassum | rgh | - | -- | -- |  |  | -- |  |  |  | - |  |  |  |  |  |  |  | -- |  |  |
| Total | Selenium | ugh | - | 300 | 1.1 |  | 0.083 | 0.099 |  | 0.085 |  | 0.104 |  | 0.094 | 0.096 | - | - | 0.082 |  | 0.1 |  | 0.072 |
| Total | silicon | ugh | - |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Siver | rgh | - | 500 | 0.05 |  | 0.02 | 0.01 |  | 0.02 |  | 0.02 | < | 0.02 | 0.02 | $\cdots$ | - | 0.01 |  | 0.02 |  | 0.01 |
| $\xrightarrow{\text { Total }}$ Toal | ${ }_{\text {Sodium }}^{\text {Storen }}$ | нgh | - | $\cdots$ | - |  | $\cdots$ | - |  |  |  | - |  |  | - | $\cdots$ | - | - |  | - |  | - |
| Total | Sulphur | ${ }_{\text {gqL }}$ | -- | - | - |  | - | - |  | - |  | - |  | - | - | - | - | - |  | - |  | - |
| Total | Thalium | Hgh | -- | $\cdots$ | -- |  | - | -- |  | - |  | - |  | - | - | - |  | -- |  | -- |  | - |
| Total | Tin | ugh | - | $\cdots$ | $\cdots$ |  | - | - |  | - |  | - |  | - | - | - |  | $\cdots$ |  | - |  | $\cdots$ |
| Total | Ttanium | ugh | - | - | - |  | $\cdots$ | $\cdots$ |  | - |  | - |  | - | - | - | - | $\cdots$ |  | $\cdots$ |  | - |
| $\frac{\text { Total }}{\text { Total }}$ | Uraium | Mgh | - | $\cdots$ | $\cdots$ |  | $\cdots$ | - |  | - |  | - |  | - | - | - | - | $\cdots$ |  | $\cdots$ |  | $\cdots$ |
| Total | Znc | нgh | -- | 3000 | 44.9 |  | 6.2 | 3.5 |  | 12.5 |  | 3.1 | < | 2 | 33.5 | - | -- | 10.2 |  | 19.4 |  | 9.8 |
| Total | ZZironium | нgl |  |  |  |  |  |  |  |  |  |  |  |  |  | - | - |  |  |  |  |  |
| VoLatle organics |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\xrightarrow{\text { Total }}$ Toal |  | $\frac{\text { Hgh }}{\text { gol }}$ | $\cdots$ | 100 200 | 0.4 0.4 |  | 0.4 0.4 | 0.4 0.4 |  | 0.45 0.4 |  | 0.4 0.4 | < | ${ }^{0.4}$ | 0.4 0.4 | $\cdots$ | - | ${ }^{0.4}$ |  | ${ }^{0.4}$ |  | ${ }^{0.4}$ |
| Total | Etayybenzene | ${ }_{\text {Hght }}^{\text {Hg }}$ | $\cdots$ | 200 | 0.4 |  |  | 0.4 |  |  |  | 0.4 | < | 0.4 | 0.4 | - | $\cdots$ | 0.4 |  | $\stackrel{0.4}{-}$ |  |  |
| Total | Meethy Tertiary Suty Eher | нgh | $\cdots$ | $\cdots$ | - |  | - | $\cdots$ |  | - |  | - |  | - | - | - | - | - |  | - |  | - |
| Toal | -xylene | Hgh | -- | -- | - |  | - | - |  | - |  | - |  | -- | - | -- | - | - |  | - |  | -- |
| Total | Toluene | ${ }_{\text {g } 9 \text { ght }}$ | $\cdots$ | 200 | 0.41 |  | 0.4 | 0.4 |  | 0.61 |  | 0.4 | - | 0.4 | 0.4 | - | - | 0 |  | 0 |  | 04 |
| Total | xylenes | yg | -- | 200 | 0.4 |  | 0.4 | < 0.4 |  | 0.48 |  | 0.45 | $<$ | 0.4 | 0.4 | -- | - | 0.4 | $<$ | 0.4 | < | 0.4 |



- Exceedances are due to compromiseedelexpired preservatives, and results are not representative

B10. Monthly Leachate Quality West Face Drainage

| State | Parameter | unis |  | $\frac{\max }{}$ | $\begin{gathered} \text { West Face } \\ \text { Drainage } \\ \text { Ss } \\ \text { 26-Apr-2022 } \end{gathered}$ | West Face Dranag SS 31-May-2022 | West Face Drainage <br> SS 27-Jun-2022 <br> 27-Jun-2022 | West Face Drainage <br> SS 27-Jul-2022 <br> 27-Jul-2022 | West Face Drainage $\begin{gathered} \text { SS } \\ \text { 22-Aug-2022 } \\ \hline \end{gathered}$ | $\begin{gathered} \text { West face } \\ \text { Dranaese } \\ \text { ss } \\ \text { 27--se-2022 } \\ \hline \end{gathered}$ | West Face Drainage 19-Oct-2022 | West Face Drainage <br> --- 30-Nov-2022 | West Face Drainage <br> 28-Dec-2022 | West Face Drainage $\qquad$ 17-Jan-2023 | $\begin{gathered} \text { West face } \\ \text { Dranae } \\ \text { Dise } \\ \text { 14-Feb-2023 } \end{gathered}$ | $\begin{gathered} \text { West Face } \\ \text { Drainage } \\ \text { Ss } \\ \text { 23-Mar-2023 } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CONVENTIONALS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total |  | ${ }_{\text {mgh }}^{\text {mgh }}$ | $\cdots$ | $\stackrel{-}{500}$ | $\stackrel{-}{50}$ | $\stackrel{-}{59}$ | $\stackrel{-110}{ }$ | $\stackrel{-170}{17}$ | $\stackrel{17}{170}$ | $\stackrel{-150}{ }$ | $\cdots$ | $\cdots$ | $\cdots$ | $\stackrel{-7}{67}$ | 7 | $\stackrel{-}{82}$ |
| Total |  | mgh |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Dissolved | Choride | mgh | $\cdots$ | 1500 | 490 | 700 | 890 | 1000 | 900 | 1000 |  |  | - | 760 | 790 | 920 |
| Total | ${ }^{\text {Cood }}$ Condutiviy | mgh | - | 1000 | 7690 | 1060 | 5370 | 2820 | ${ }^{9330}$ | 2590 |  |  |  | ${ }^{1830}$ | 1550 | ${ }^{1800}$ |
| Total | Conductivit | ${ }_{\text {Lsocm }}^{\text {molt }}$ | $\cdots$ | 1 | ${ }_{0}^{6264}$ | (9401 | ${ }^{133388}{ }_{0}^{10040}$ | 14066 <br> 0041 <br> 0 |  | ${ }^{149953} \begin{aligned} & 10604 \\ & 0.0\end{aligned}$ | - | - | - | ${ }^{110066}$ | ${ }^{10063}$ | 12673 <br> 0.039 |
| Toal | Cranaide e WAD | ${ }_{\text {mgh }}^{\text {mol }}$ | -- | I | ${ }_{0}^{0.0013}$ | ${ }_{0}^{0.0308}$0.0177 |  | ${ }_{0}^{0.044}$ | 0.05 0.05 | ${ }_{0}^{0.0064}$ | - | - | $\cdots$ | 0.05 <br> 0.05 | ${ }_{0}^{0.0037}$ | 0.039 <br> 0.025 |
| $\xrightarrow{\text { Tooal }}$ Toal | Cissolved OXxygen | ${ }_{\text {mght }}^{\text {mgh }}$ | - | - | ${ }_{0.24}$ | ${ }_{1}^{1.26}$ | 0.79 |  |  |  | - | - | - | ${ }_{3.04}$ | ${ }^{0.04}$ | ${ }_{1}^{0.2025}$ |
| Toal | Fecal Coliform | CFUV100 mL |  | - |  |  |  | - | -- |  |  |  |  |  |  |  |
| Total | $\mathrm{N}-\mathrm{Nh} 3(\mathrm{AsN})$ | mgl | -- | - | 470 | 750 | 1300 | 1400 | 1300 | 1600 | -- | - | -- | 900 | 880 | 1100 |
| Total | N - $\mathrm{No} 2(\mathrm{As} \mathrm{N}$ ) | mgl | - | - | 0.5 | 0.5 | 0.0177 | 0.67 | 0.5 | 0.5 | - | - | - | 0.5 | 0.5 | ${ }_{0} 0.5$ |
| Total | $N$ - $\mathrm{No3}(\mathrm{As} \mathrm{N}$ ) | mgl | - | - | 2 | 2 | 0.02 | 2 | 2 | 2 | - | - | -- | ${ }^{8.7}$ | 31.3 | 2 |
| Total | N- $\mathrm{No3} 3+\mathrm{Noz}$ (As N) | mgh | - | - | - | - | - | - | - | - | - | - | - | - |  | - |
| Toial |  | mgh | - | -- | - | - |  |  | - |  |  |  |  |  |  |  |
| Toal |  | ${ }_{\text {mgh }}^{\text {mal }}$ | $\cdots$ | $\stackrel{-}{15}$ | 2 | 2 | 2 | 2 | 2 | 2 | - | - | - | $\overline{9.9}$ | 2 | $\stackrel{-}{2}$ |
| Total | Oil grease, total | mgh | -- | 100 |  |  |  | 1 | 2.4 | ${ }^{1.3}$ |  |  |  |  |  | 3 |
| Total |  | mV |  |  | 116 | ${ }^{-14}$ | -93 | ${ }^{130}$ | -52 | ${ }^{36}$ |  |  |  | 158.5 | 151.0 | 141.0 |
| Total | pH | pH | 5.5 | 11 | 7.47 | 7.58 | 7.7 | 7.72 | 8.41 | 7.87 |  |  |  | 7.66 | 8.10 | 8.08 |
| Dissolved | Iohide | mgh |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Suphide | mg/ | -- | 1 | 0.023 | 0.018 | 0.041 | 0.023 | 0.018 | 0.036 | - | - | - | 0.036 | 0.036 | 0.036 |
| Dissolved | Suphate | mgh | $\cdots$ | 1500 | 100 |  | 100 | 100 |  |  |  |  |  |  | ${ }_{48}^{48}$ |  |
| Total | Temperature |  |  |  | 23.4 | 25.5 | 26.6 | 24.1 | 25.4 | 22.4 | - | - |  | 27.7 | ${ }^{27}$ | 25.5 |
| Total |  | mgl | - | $\cdots$ |  |  |  |  |  |  | - | - | - |  |  |  |
| Total <br> Total | ${ }_{\text {T }}$ Total Phenols | mgh | $\cdots$ | $\stackrel{1}{350}$ | ${ }_{3.3}^{0}$ | ${ }^{0.065}$ | ${ }^{0.099}$ | ${ }_{43}^{1.3}$ | 0.08 110 | 0.0041 | $\cdots$ | - | $\cdots$ | ${ }^{0.011}$ | ${ }^{0.067}$ | 0.1 19 |
| METALS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Aluminum |  |  |  |  |  |  |  |  |  |  |  |  |  |  | -- |
| Total | Antimony | Hgh | $\cdots$ | $\stackrel{-}{40}$ | $\stackrel{\square}{14}$ | $\stackrel{7}{192}$ | 447 | $\stackrel{-7}{46.5}$ | $\cdots$ | 658 | - | - | - | 35 | 28 | 368 |
| Total | Barium | Hgh | -- |  |  |  |  |  |  |  |  | - | - |  |  |  |
| Total | Berylium | pgh | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Total | Bismuth | нgh | - | - | - | - | - | - | - | - | - | - | - | - | -- | - |
| Tolal | Bron | ngh | - | $\cdots$ |  |  |  |  |  |  | - | - | - |  |  |  |
| Toal | ${ }^{\text {caamium }}$ | Hgl | - | 300 | 0.081 | 0.13 | 0.284 | 0.262 | 0.268 | ${ }_{0}^{0.329}$ | - | - |  | 0.204 | 0.182 | 0.256 |
| Tooal | Chromium | $\underset{\text { Lght }}{\text { Lght }}$ | - | 4000 | ${ }^{72.2}$ | 97.5 | 209 | ${ }^{273}$ | 242 | ${ }^{317}$ | - | - | - | 153 | 89.2 | 194 |
| Total | Chromium III | pgh | - |  |  |  |  |  |  |  | - | - | - | - |  | - |
| Iotal | Chromium Vi | Hgh | - | $\cdots$ | $\stackrel{-}{225}$ | -- | - | $\stackrel{-}{39}$ | - | -- |  |  |  | - | $\underline{723}$ | , |
| Toal | cooalt | pgh | - | ${ }_{1000}$ | 12. | 179 | ${ }_{5}$ | 72 | 18.5 | 88.5 |  |  |  | 2 | 2 | ${ }^{36.6}$ |
| $\xrightarrow{\text { Tolal }}$ | Copper Harness (As Caco3) | ${ }_{\text {Lgh }}$ | $\cdots$ | 1000 | $\stackrel{19.1}{ }$ | \% | 51.3 |  | 18.6 | 8.15 | - |  |  | 33.8 | 35.2 | 43.4 |
| Total | ron | pgh | -- | 50000 | 2070 | 2740 | 4270 | 3670 | 4390 | 4430 | - | - | - | 2140 | 2000 | 2990 |
| Total | Lead | нgl | -- | 1000 | 1.64 | 2.01 | 4.88 | 4.29 | 4.24 | 4.85 |  | - | -- | 2.96 | 1.85 | 3.56 |
| Total | Litium | ugh | -- |  | - |  |  |  |  | - | - |  |  |  |  |  |
| Total | Magnesium | $\xrightarrow{\text { Mgh }}$ | $\cdots$ | 5000 | 354 | 9414 | $\stackrel{-}{431}$ | 380 | $\stackrel{-116}{ }$ | 45 |  | - | - | $\stackrel{-}{305}$ | 251 |  |
| Total | Mercury | Lgh | $\cdots$ | ${ }_{20}$ | 0.019 | 0.038 | 0.038 | 0.019 | 0.019 | ${ }_{0} 0.038$ | $\cdots$ | - | - | ${ }_{0.038}$ | ${ }_{0}^{0.038}$ | ${ }_{0}^{0.038}$ |
| Total | Moybdenum | pgh | -- | 5000 | 1.32 | 1.02 | 2.85 | 1.51 | 1.84 | 1.49 | - | - | - | 2.65 | 1.27 | 5.7 |
| ${ }_{\text {Tolal }}^{\text {Toalal }}$ | ${ }^{\text {Nickel }}$ Phoshorus | $\underset{\text { ugh }}{\text { ugh }}$ | $\cdots$ | 3000 | 66.9 | 91.2 | 126 | 115 | 106 | 134 | $\cdots$ | $\cdots$ | - | 87.1 | 83.3 | 117 |
| Total | Potassium | pgh | $\cdots$ |  |  | - |  | - |  |  | - | - | - |  |  |  |
| Tolal | Selenium | pgr | - | 300 | 0.64 | 0.71 | 1.16 | 0.98 | 1.14 | 1.29 | - | - | - | 0.879 | 0.76 | 1.09 |
| Total | siver | +g | - | 500 | 0.05 | 0.05 | 0.05 | 0.1 | 0.05 | 0.05 |  |  |  | 0.019 | 0.05 | 0.05 |
| Total |  | pgh | $\cdots$ |  |  |  |  | - |  | - | - | - | - |  |  |  |
| Total | Strontum | Mgh | $\cdots$ | - | - | - | - | - | - | - | - | - |  |  |  |  |
| Toal | Thalium | ygh | - | - | - | - | - | - | - | - | - | - | - | - |  |  |
| Total | Tin | pgh | $\cdots$ | - | - | - | - | - | - | - |  |  |  | - | - | - |
| Total | TTanium | нgh | - | - | - | - | - | - | - | - |  |  | - | - |  |  |
| Total | Uranium | ngh | - | - | - | - | $\cdots$ | $\cdots$ | $\cdots$ | - | - | - | - | - | - | -- |
| Toal | Vanadium | pgr | - |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Znio | п92 | $\cdots$ | 3000 | ${ }^{25.2}$ | ${ }^{35}$ | ${ }^{89.3}$ | 103 | 99.7 | ${ }^{120}$ | $\cdots$ | - | $\cdots$ | 49 | 41.7 | 70 |
| Total |  | нgh | - | $\cdots$ |  | $\cdots$ | - | $\cdots$ | $\cdots$ | $\cdots$ | $\cdots$ | - | - | - |  | $\cdots$ |
| volatle organics |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Tolal | ${ }^{\text {bentene }}$ Etylberzene | $\xrightarrow{\text { pgl }}$ | $\cdots$ | ${ }^{100}$ | ${ }_{6}^{1.3}$ | ${ }_{3.9}^{1.6}$ | ${ }_{0}^{1.7}$ | ${ }_{3.4}^{1.1}$ | ${ }_{0}^{0.4}$ | ${ }_{0}^{0.4}$ | -- | -- | - | ${ }_{5.6}^{2.2}$ | ${ }_{3.2}^{1.3}$ | ${ }_{2.9}^{2.1}$ |
| Total | M P P X Xlenes | M9/ | -- |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Toal | Methy Teriary Buty Eher | pgr | $\cdots$ | $\cdots$ | - | - | - | - | - | - | - | - | - | - | - | - |
| Toalal | O-xylene | pgr |  | - |  | - |  |  | - | - |  |  |  |  |  |  |
| Toal | Streene | нgh |  |  |  |  |  |  |  |  | $\cdots$ | - | - |  |  |  |
| Total | $\pm$ | $\underset{\text { pgh }}{\text { pgh }}$ | $\cdots$ | 200 200 | ${ }^{3.3}$ | ${ }_{4}^{4.2}$ | ${ }^{3.2} 18$ | $\stackrel{8}{8.9}$ | ${ }_{4}^{1.6}$ | $<\begin{gathered}0.4 \\ 3.6\end{gathered}$ | $\cdots$ | - | $\cdots$ | 16 | ${ }_{11}^{3.1}$ | ${ }_{17}^{5.6}$ |


*. Exxeedances may be due to compromisedlexppiried peeserevatives, and results may not be representative
ss. Singles Sample
SS- Single Sample
Ns - Not Sampled

## B11. Monthly Leachate Quality - Cell 3 Pipe Outlet



## B12. Monthly Leachate Quality Emerging Contaminant

Appendix B-12. Monthly Leachate Quality - Emerging Contaminant 2022-2023

| State | Parameter | Units | Sewer Use Criteria |  | $\begin{gathered} \text { Hartland Valve } \\ \text { Chamber } \\ --- \\ 26-\text { Apr-2022 } \\ \hline \end{gathered}$ | Hartland Valve Chamber SS 31-May-2022 |  | Hartland Valve Chamber27-Sep-2022 | Hartland Valve Chamber27-Sep-2022 | Hartland Valve Chamber27-Sep-2022 | Hartland Valve Chamber <br> --- <br> 19-Oct-2022 | Hartland Valve Chamber $\qquad$ <br> 30-Nov-2022 | Hartland Valve Chamber --- <br> 30-Nov-2022 | Hartland Valve Chamber $\qquad$ <br> 30-Nov-2022 | Hartland Valve Chamber14-Feb-2023 | Hartland Valve Chamber <br> --- <br> 14-Feb-2023 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  | min | max |  |  |  |  |  |  |  |  |  |  |  |  |
| EMERGING CONTAMINANTS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | Perfluorobutanesulfonic acid | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 0.47 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | Perfluorodecanesulfonic acid (PFDS) | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 0.48 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | Perfluoroheptanesulfonic acid | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- | < | 0.02 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | Perfluorohexanesulfonic acid | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 0.047 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | Perfluorononanesulfonic acid | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- | < | 0.02 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | Perfluorooctanesulfonic acid | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 0.19 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | Perfluorooctanoic acid (PFOA) | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 1. | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | Perfluoropentanesulfonic acid | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- | $<$ | 0.02 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PF3A | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 0.037 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PF4A | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- | < | 0.02 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PFBoA | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 0.18 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PFDA | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 0.6 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PFDoA | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 0.026 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PFHpA | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 0.43 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PFHxA | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- | < | 0.02 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PFNA | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- | < | 0.02 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PFOSA | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- | $<$ | 0.02 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PFPeA | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | 0.51 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | PFUA | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- | $<$ | 0.02 | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Total | 1,4-Dioxane | $\mu \mathrm{g} / \mathrm{L}$ | --- | --- | --- |  | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | Notes:

Notes:

- Exceeded minimum allowable value specified in CRD Sewer Use Bylaw 2922.


## Appendix C

## Climate Data

- C1. Daily Rainfall Data -

Hartland Landfill Weather Station

- C2. Monthly Rainfall Data -

Hartland Landfill Weather Station

C1. Daily Rainfall Data Hartland Landfill Weather Station

| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| January | 1 | 2.4 | 11.6 | 0 | 7.6 | 0 | 8 | 5.4 | 0 | 11 | 5 | 35 | 0 | 22.2 | 22 | 0 | 0 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 10.5 | 2.25 | 0 |
| January | 2 | 0 | 0 | 0 | 0 | 0 | 2.6 | 32.6 | 0 | 0 | 0.8 | 64.4 | 11.2 | 0 | 1.3 | 0.0 | 0.8 | 0.0 | 1.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0 | 0 | 40.5 | 0 | 0 |
| January | 3 | 0 | 0.2 | 0 | 16.8 | 4.6 | 2.6 | 6.8 | 5.8 | 0 | 1.6 | 7.8 | 1.6 | 0.2 | 0.0 | 0.0 | 5.5 | 0.0 | 10.5 | 4.3 | 0.0 | 0.0 | 0.0 | 17.75 | 0 | 7.5 | 16 | 0.75 |
| January | 4 | 0 | 16.2 | 0 | 11.2 | 13 | 0.2 | 15 | 0 | 0 | 1.4 | 9.6 | 0.4 | 20.4 | 40.0 | 4.3 | 8.3 | 5.3 | 0.0 | 3.5 | 0.5 | 0.0 | 0.8 | 74 | 0 | 13.75 | 5.75 | 0 |
| January | 5 | 0 | 20.8 | 0 | 0 | 30.2 | 0 | 0 | 0 | 0 | 23.8 | 57.4 | 11 | 2.8 | 5.8 | 21.8 | 22.8 | 1.8 | 0.0 | 31.5 | 2.0 | 0.0 | 4.0 | 1.5 | 0 | 16.5 | 8.75 | 0.25 |
| January | 6 | 0 | 20.4 | 0.4 | 6.4 | 0 | 10.2 | 0 | 9.8 | 3 | 11.4 | 2.6 | 0 | 21 | 0.0 | 23.8 | 8.5 | 3.8 | 0.0 | 26.0 | 2.5 | 0.0 | 0.0 | 0.5 | 0 | 0.75 | 0 | 0.5 |
| January | 7 | 0 | 1 | 0 | 1.8 | 0 | 44.4 | 0 | 16.4 | 14.2 | 6 | 55.8 | 0 | 28.6 | 0 | 11 | 1 | 9 | 0 | 0 | 0 | 0 | 14 | 7.5 | 0 | 0.25 | 33.75 | 0 |
| January | 8 | 0 | 0 | 0 | 7.6 | 0.2 | 21.8 | 0 | 2.2 | 12.8 | 1.8 | 4.2 | 8.2 | 2.2 | 15.0 | 0.0 | 1.8 | 6.0 | 9.0 | 0.3 | 0.0 | 0.0 | 6.5 | 0 | 0 | 5 | 0.75 | 0 |
| January | 9 | 0 | 0 | 19.6 | 19.8 | 0 | 0 | 0 | 2 | 0 | 16.4 | 10.8 | 0.6 | 0.2 | 2.0 | 0.3 | 0.5 | 6.8 | 7.5 | 0.3 | 0.0 | 0.3 | 1.5 | 1.5 | 0 | 1.5 | 1.25 | 0 |
| January | 10 | 0 | 0 | 21 | 3 | 0 | 0.4 | 0 | 1.2 | 0 | 42.4 | 12 | 36.6 | 17.2 | 1.5 | 0.0 | 0.3 | 2.3 | 5.0 | 2.3 | 0.0 | 0.0 | 2.5 | 1.5 | 23 | 3.25 | 0 | 0 |
| January | 11 | 0 | 0 |  | 1.2 | 0 | 0.4 | 0 | 1.6 | 0.6 | 14.2 | 0.6 | 3.4 | 0.8 | 27.8 | 0.0 | 0.0 | 0.0 | 2.5 | 3.0 | 0.0 | 0.0 | 4.0 | 1.5 | 2 | 23.5 | 7 | 1.5 |
| January | 12 | 0 | 0 | 0.2 | 4.2 | 0 | 7.8 | 7 | 0 | 0 | 17 | 0 | 1.8 | 0 | 4 | 3 | 0 | 1 | 4 | 1 | 6 | 0 | 5 | 0 | 12.5 | 34 | 25 | 17 |
| January | 13 | 0 | 17.2 | 0 | 0.6 | 3.2 | 0 | 0.4 | 2.2 | 0 | 8.4 | 0 | 0 | 0 | 2.5 | 16.3 | 0.0 | 0.0 | 2.3 | 0.3 | 12.5 | 0.0 | 3.8 | 0 | 0 | 0.75 | 13.25 | 3.25 |
| January | 14 | 0 | 20.4 | 65.6 | 6 | 0 | n/a | 2.2 | 2.2 | 0 | 0 | 0 | 12 | 0 | 22 | 14 | 1 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 2.25 | 0 |
| January | 15 | 0 | 22.8 | 0.2 | 10.8 | 0 | 0 | 0 | 2 | 3.6 | 0 | 0 | 0 | 0 | 21 | 15 | 6 |  | 0 | 0 | 0 | 0 | 0 | 0 | 3.75 | 0 | 0.75 | 5.5 |
| January | 16 | 0 | 13.4 | 10.2 | 6.4 | 0 | 0.8 | 0 | 0 | 8.6 | 36 | 5.6 | 0 | 0 | 0 | 11 | 0 | 0 | 0 | 3 | 2 |  | 0 | 0 | 0 | 0.75 | 0.25 | 0 |
| January | 17 | 0 | 16.4 | 8.6 | 2.2 | 0.4 | 0.8 | 0 | 0 | 54.4 | 1.4 | 0 | 0 | 0 | 1.3 | 0.3 | 0.5 | 0.0 | 0.0 | 0.0 | 12.0 | 14.5 | 12.5 | 0 | 0 | 1.5 | 0 | 2.25 |
| January | 18 | 0 | 0.2 | 20.6 | 0 | 10.2 | 1.8 | 0 | 5.4 | 18.4 | 0 | 12.4 | 0.4 | 0 | 2 | 0 | 0 | 0 | 0 | 21 | 6 | 6 | 4 | 7.75 | 1.75 | 0.25 | 6.5 | 1.5 |
| January | 19 | 0 | 0.4 | 3.2 | 0 | 2.6 | 8.6 | 0 | 2.6 | 15.4 | 3.6 | 5.4 | 2.8 | 0 | 0.0 | 0.0 | 0.0 | 0.3 | 0.3 | 37.0 | 0.5 | 2.3 | 4.0 | 5.25 | 1.5 | 0 | 3.75 | 0 |
| January | 20 | 0 | 0 | 2.8 | 5.4 | 0.4 | 11.6 | 0.6 | 0 | 10 | 6 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 0 | 1 | 0 | 0 | , | 4.5 | 1.75 | 0 | 5 | 0 |
| January | 21 | 0 | 0.4 | 1 | 10 | 21.4 | 10.8 | 9.4 | 0 | 0.4 | 2.2 | 2 | 0 | 0 | 0.0 | 8.8 | 19.5 | 0.0 | 0.0 | 0.0 | 7.8 | 0.5 | 12.0 | 0.25 | 18 | 0.75 | 6.5 | 2.25 |
| January | 22 | 0 | 1 | 1.6 | 0 | 0 | 4 | 27.6 | 0 | 42.2 | 0 | 18.2 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 1 | 34 | 0 | 12 | 0 | 21.25 | 0 | 0 | 0.25 |
| January | 23 | 0 | 16 | 0.4 | 0.4 | 0 | 4.4 | 10 | 8.6 | 0.2 | 0.2 | 8.8 | 0 | 0 | 0.0 | 5.0 | 4.3 | 0.0 | 0.0 | 8.8 | 2.5 | 0.0 | 23.0 | 31.5 | 19.5 | 0.75 | 0 | 2 |
| January | 24 | 0 | 6.2 | 0 | 0 | 1.2 | 22.8 | 9.4 | 0.6 | 0 |  | 0 | 0 | 0 | 4 | 10 | 1 | 4 | 0 | 14 | 0 | 0 | 1 | 3.5 | 3.25 | 13.25 | 0.25 | 2 |
| January | 25 | 0 | 0.4 | 0.4 | 0.8 |  | 20.4 | 0.6 | 0 | 0 | 0 | 0.2 | 0 | 0 | 1 | 0 | 20 | 6 | 0 | 4 | 0 | 0 | 4 | 0.25 | 3.75 | 3.5 | 0 | 0.25 |
| January | 26 | 0 | 6.6 | 2 | 0 | 0 | 3 | 24.4 | 3.6 | 0 | 18.6 | 0 | 1 | 0 | 0.0 | 0.0 | 21.5 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 6.5 | 0 | 3 | 0.5 | 0 | 1.5 |
| January | 27 | 0 | 4 | 7.4 | 0 | 0 | 6.2 | 6.2 | 4.4 | 0 | 3 | 0 | 0.2 | 6 | 0 | 0 | 0 | 3 | 0 | 0 | 10 | 0 | 6 | 0 | 12 | 2 | 0 | 0.5 |
| January | 28 | 0 | 1.2 | 19.2 | 0 | 0 | 0 | 0 | 16.8 | 0.2 | 11.2 | 0 | 0.4 | 0 | 0.0 | 5.0 | 0.0 | 0.0 | 1.5 | 0.3 | 0.5 | 0.0 | 3.0 | 0 | 8.5 | 3.25 | 0.75 | 0 |
| January | 29 | 0 | 0 | 88.2 | 0 | 5.8 | 0 | 7.4 | 13.4 | 0 | 33.2 | 0 | 8.2 | 0 | 0.5 | 6.0 | 1.5 | 0.0 | 1.3 | 0.0 | 1.3 | 2.0 | 2.8 | 0 | 15.25 | 0 | 0 | 0 |
| January | 30 | 0 | 3.2 | 2.2 | 0 | 5.2 | 16 | 2.8 | 18.6 | 11.4 | 10.8 | 0 | 4.2 | 0 | 6.5 | 0.0 | 21.8 | 0.5 | 21.3 | 0.0 | 8.8 | 0.0 | 13.0 | 0 | 9.5 | 0 | 2.25 | 0 |
| January | 31 | 0 | 0 | 5 | 1 | 0 | 3.6 | 38 | 0.4 | 3.8 | 1 | 0 | 2.8 | 0 | 0 | 0 | 1 | 1 | 2 | 0 | 4 | 0 | 5 | 0 | 6.25 | 10.25 | 7.5 | 4 |
| Total Monthly Rainfall |  | 2.4 | 200 | 28.8 | 123.2 | 99.4 | 213.2 | 205.8 | 119.8 | 210.2 | 277.4 | 312.8 | 106.8 | 121.6 | 180.5 | 161.75 | 150.75 | 53.75 | 70.25 | 160.75 | 114.5 | 26 | 153.5 | 158.75 | 166.5 | 194.5 | 149.5 | 45.25 |
| Maximum Da | Rainfall | 2.4 | 22.8 | 88.2 | 19.8 | 30.2 | 44.4 | 38 | 18.6 | 54.4 | 42.4 | 64.4 | 36.6 | 28.6 | 40 | 23.75 | 22.75 | 8.5 | 21.25 | 37 | 34 | 14.5 | 23 | 74 | 23 | 40.5 | 33.75 | 17 |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| February | 1 | 0 | 0 | 28.8 | 26.4 | 0 | 0.8 | 0.2 | 6 | 0 | 10.2 | 0 | , | 1.4 | 3.0 | 0.0 | 4.8 | 0.5 | 0.0 | 1.0 | 7.5 | 0.0 | 14.50 | 1.0 | 24.25 | 22 | 0 | 1.5 |
| February | 2 | 0 | 0 | 45.4 | 0.4 | 9.4 | 0 | 0.2 | 8 | 0 | 11 | 0 | 0.2 | 4.6 | 2 | 0 | 5 | 0 | 0 | 3 | 1 | 0 | 5.50 | 11.3 | 0.25 | 7 | 0 | 0 |
| February | 3 | 0 | 0.4 | 3.2 | 0 | 3.2 | 0.4 | 0 | 1.2 | 1.6 | 5.6 | 1.6 | 0 | 0 | 1.0 | 2.0 | 0.0 | 0.0 | 0.0 | 3.5 | 0.3 | 0.0 | 1.50 | 0.0 | 0 | 0 | 2 | 4.25 |
| February | 4 | 0 | 0 | 6.2 | 0 | 9.2 | 0 | 0 | 2.2 | 10.6 | 32.6 | 6.6 | 0.2 | 0 | 0.5 | 9.8 | 0.0 | 1.8 | 0.0 | 1.0 | 1.8 | 14.0 | 1.3 | 0.3 | 13 | 0.75 | 3.75 | 2.5 |
| February | 5 | 0 | 0 | 0.6 | 0 | 4 | 2.6 | 0 | 0 | 0 | 0 | 0 | 12.8 | 0 | 1 | 0 | 0 | 0 | 0 | 4 | 0 | 3 | 0.0 | 0.0 | 18.5 | 0.25 | 2.25 | 2 |
| February | 6 | 0 | 0 | 17.6 | 0.4 | 0.4 | 11.8 | 0 | 6.4 | 25.4 | 0 | 0 | 14 | 4.4 | 0 | 9 | 0 | 10 | 0 | 28 | 0 | 0 | 0.8 | 1.3 | 17.75 | 0 | 0 | 11.5 |
| February | 7 | 0 | 0 | 5.4 | 0.4 | 0 | 8.2 | 0 | 0 | 0 | 0.2 | 4 | 13 | 0 | 2.8 | 0.3 | 0.0 | 6.8 | 0.0 | 17.3 | 0.0 | 0.3 | 0.3 | 0.3 | 18.25 | 0 | 0 | 27.25 |
| February | 8 | 0 | 1 | 11.4 | 5.4 | 13.4 | 0.2 | 0 | 0 | 0 | 0.4 | 1 | 1.8 | 0 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 | 10.8 | 0.0 | 0.0 | 0.3 | 0.0 | 0.25 | 0 | 0 | 0.5 |
| February | 9 | 0 | 1.6 | 1.6 | 1.8 | 0 | 0 | 0 | 0 | 0 | 0 | 1.4 | 3.4 | 0 | 0 | 0 | 6 | 0 | 0 | 5 | 0 | 15 | 0.0 | 0.0 | 0 | 0 | 1 | 5.75 |
| February | 10 | 0 | 0.8 | 2.2 | 0 | 0 | 0.6 | 0 | 0 | 0 | 0 | 0.4 | 5.8 | 12.2 | 0 | 0 | 15 | 0 | 0 | 11 | 0 | 4 | 0.0 | 0.0 | 0 | 0 | 0.5 | 1.25 |
| February | 11 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 1.2 | 3.4 | 1.4 | 3 | 1 | 2 | 0 | 17 | 2 | 4 | 1 | 0.0 | 0.0 | 0 | 0 | 0 | 0 |
| February | 12 | 0 | 13.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0.6 | 0 | 0.8 | 0 | 0 | 4 | 23 | 1 | 0 | 7 | 0 | 15 | 0 | 0.0 | 0.0 | 0 | 0 | 0 | 2 |
| February | 13 | 0 | 11.8 | 2.4 | 0 | 0 | 0 | 0 | 0 | 0 | 5.8 | 1 | 0 | 0 | 2.0 | 1.8 | 4.0 | 1.0 | 5.8 | 5.8 | 10.5 | 0.0 | 1.3 | 0.0 | 2.5 | 0 | 0 | 8 |
| February | 14 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 14.6 | 0 | 0 | 4 | 12 | 1 |  | 2 | 4 | 12 | 5 | 1.0 | 0.0 | 0.25 | 1.5 | 0 | 0 |
| February | 15 | 0 | 0.6 | 0 | 0.6 | 0.6 | 0 | 1.2 | 3.8 | 0 | 0 | 10.8 | 3.4 | 0 | 1 | 0 | 0 | 9 | 10 | 0 | 8 | 5 | 1.0 | 0.0 | 4.5 | 11.75 | 0.5 | 0 |
| February | 16 | 0 | 0 | 4.8 | 0.2 | 10.8 | 7.8 | 2.4 | 2.8 | 0 | 0 | 0.8 | 0 | 0 | 5 | 7 | 0 | 1 | 18 | 0 | 74 | 0 | 6.0 | 10.5 | 0 | 11.25 | 0 | 2.25 |
| February | 17 | 0 | 0 | 12.8 | 0 | 0.4 | 0.6 | 5.4 | 3 | 0 | 0 | 2.8 | 0 | 0 | 0 | 0 | 7 | 0 | 3 | 0 | 0 | 0 | 20.8 | 0.5 | 0 | 0 | 0 | 1.75 |
| February | 18 | 0 | 0.6 | 10.6 | 0 | 3.4 | 0 | 1.6 | 8 | 0 | 0 | 1.6 | 0 | 0 | 0 | 0 | 9 | 0 | 13 | 0 | 15 | 0 | 0.5 | 0.0 | 0 | 5.25 | 0 | 0.5 |
| February | 19 | 0 | 3.2 | 9.6 | 0 | 0 | 4.8 | 2.4 | 0 | 0 | 0 | 22.2 | 0 | 0 | 0 | 0 | 28 | 6 | 18 | 0 | 6 | 0 | 0.0 | 0.0 | 0 | 3.25 | 0 | 0 |
| February | 20 | 0 | 0 | 0 | 0 | 0 | 0.4 | 10.2 | 0 | 0 | 0 | 16 | 2.2 | 0 | 0 | 0 | 0 | 0 | 11 | 2 | 1 | 0 | 0.0 | 0.8 | 0 | 4 | 0 | 1.25 |
| February | 21 | 0 | 3.2 | 3.8 | 1.4 | 0 | 65.8 | 2.4 | 0 | 0 | 0 | 1.2 | 0 | 0 | 0 | 1 | 3 | 1 | 3 | 0 | 3 | 0 | 0.0 | 0.3 | 0 | 19.25 | 0.5 | 2.25 |
| February | 22 | 0 | 3 | 24.4 | 1.4 | 4.4 | 38.2 | 0.8 | 0 | 0 | 0 | 0.6 | 0 | 2.4 | 0 | 0 | 2 | 1 | 0 | 0 | 1 | 0 | 2.8 | 0.0 | 2.5 | 1.25 | 5.25 | 0 |
| February | 23 | 0 | 2.6 | 10.6 | 4 | 0 | 14.4 | 0.2 | 0 | 0 | 1.2 | 0 | 0 | 1.6 | 2.3 | 0.0 | 0.0 | 21.3 | 4.0 | 0 | 11 | 0 | 3.3 | 4.5 | 15.25 | 5 | 0 | 0 |
| February | 24 | 0 | 0 | 54.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5.4 | 0 | 1.8 | 10 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 3.5 | 5.3 | 0 | 0 | 0 | 0 |
| February | 25 | 0 | 1.8 | 1.6 | 0 | 0 | 0 | 0 | 4.2 | 0 | 0 | 0 | 0 | 13.4 | 1 | 0 | 28 | 2 | 4 | 0 | 0 | 0 | 5.0 | 3.5 | 0.25 | 10 | 1.25 | 0 |
| February | 26 | 0 | 0 | 0.2 | 5.8 | 0 | 0 | 0 | 3.8 | 0 | 12 | 0 | 0 | 0 | 5 | 0 | 6 | 6 | 16 | 1 | 0 | 0 | 0.0 | 0.0 | 0.25 | 1 | 0 | 9 |
| February | 27 | 0 | 1 | 21.6 | 16.8 | 0 | 0 | 0 | 1.8 | 0 | 0.2 | 0 | 1.2 | 0 | 3.0 | 2.5 | 0.0 | 1.3 | 0.0 | 15.0 | 0.3 | 0.0 | 0.8 | 0.0 | 0 | 0.75 | 7.25 | 1 |
| February | 28 | 0 | 13 | 7.4 | 1 | 0 | 0 | 2 | 0 | 1 | 0 | 0 | , | 0.6 | 0 | 4 | 2 | 0 | - | 20 | 1 |  | 0.3 | 0.0 | 10.25 | 0.5 | 31.75 | 14.5 |
| February | 29 | n/a | n/a | n/a | 13.6 | n/a | n/a | n/a | 0.8 | n/a | n/a | n/a | 1.6 | n/a | n/a | n/a | 0 |  |  |  | 3.25 |  |  |  | 1 |  |  |  |
| Total Monthly Rainfall Maximum Daily Rainfall |  | 0 | 58.6 | 286.4 | 79.6 | 59.2 | 160.6 | 29 | 53 | 39.2 | 79.2 | 94 | 63 | 43.8 | 50.25 | 72.5 | 122.5 | 67.3 | 130.0 | 133.0 | 172.8 | 50.0 | 70.0 | 39.3 | 129.0 | 104.8 | 56.0 | 99.0 |
|  |  | 0 | 13.8 | 54.2 | 26.4 | 13.4 | 65.8 | 10.2 | 8 | 25.4 | 32.6 | 22.2 | 14 | 13.4 | 9.75 | 22.5 | 27.5 | 21.25 | 18.25 | 27.75 | 74 | 14.75 | 20.75 | 11.25 | 24.25 | 22 | 31.75 | 27.25 |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| March | 1 | 0 | 11.8 | 9.8 | 0.4 | 6.6 | 0 | 0 | 0 | 6.6 | 0 | 0 | 2.8 | 0 | 0 | 4.25 | 5.00 | 18.00 | 0.00 | 0.00 | 3.00 | 1.75 | 0.25 | 0.00 | 0.75 | 0.25 | 23.25 | 0.25 |
| March | 2 | 0 | 1 | 3.6 | 2.6 | 11.2 | 0 | 1.2 | 0 | 0.2 | 0 | 5.4 | 1.6 | 0.2 | 2 | 6.5 | 0.0 | 3.3 | 0.0 | 4.0 | 22.5 | 7.0 | 14.0 | 0.0 | 5.75 | 0 | 5.75 | 10 |
| March |  | 0 | 2 | 32 | 6.2 | 2.2 | 0 | 1.4 | 2.2 | 0 | 0 | 2.4 | 4.6 | 0 | 0 | 1.75 | 5.75 | 6.25 | 13.00 | 0.25 | 5.50 | 11.75 | 0.75 | 0.00 | 0 | 0 | 5 | 3.5 |
| March |  | 2.2 | 0 | 9 | 5.8 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0.2 | 0.2 | 0 | 0 | 7.25 | 0.00 | 0.00 | 4.00 | 0.00 | 7.50 | 0.00 | 0.00 | 0.00 | 0.75 | 2.5 | 0 | 3.5 |
| March | 5 | 0.8 | 0 | 0 | 0 | 0 | 0 | 0.4 | 11 | 0 | 0 | 1.2 | 0 | 0 | 0 | 5.5 | 4.8 | 0.0 | 14.0 | 0.0 | 0.0 | 4.0 | 0.0 | 0.0 | 1.25 | 0 | 0 | 0.25 |
| March | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0.4 | 0 | 0 | 2.2 | 0 | 0 | 0 | 0 | 0.25 | 12.25 | 3.75 | 2.5 | 0 | 0 | 1.75 | 0 | 0 | 0 | 0.75 | 0 | 0.25 |
| March | 7 | 0.2 | 0 | 0 | 0 | 0 | 0.2 | 4.8 | 43.8 | 0.6 | 0.4 | 2.8 | 2.2 | 0 | 1.5 | 0 | 0 | 6.5 | 14.75 | 0 | 7.75 | 10.25 | 0 | 0 | 0 | 2.75 | 0 | 0 |
| March | 8 | 0 | 2.6 | 0 | 0 | 5.2 | 2.8 | 0.8 | 0 | 1.6 | 15.8 | 8.8 | 0 | 1 | 0 | 0 | 0 | 1.5 | 0 | 0 | 11.75 | 7.75 | 5.25 | 0.25 | 0 | 0 | 0 | 0 |
| March | 9 | 8.8 | 8.2 | 0 | 1 | 2.8 | 1.2 | 20.4 | 0.2 | 3.4 | 5.4 | 1.6 | 4.4 | 0 | 0.25 | 7 | 0 | 0 | 2.5 | 0 | 1.75 | 11.75 | 0.25 | 0 | 0 | 0 | 1 | 0 |
| March | 10 | 0 | 5.8 | 0.6 | 0.2 | 0 | 3.4 | 3 | 0 | 0 | 0.8 | 12 | 0 | 0 | 2.25 | 15.75 | 15.5 | 0 | 0 | 0 | 6.25 | 0.25 | 0 | 0 | 0.75 | 0 | 0 | 1.25 |
| March | 11 | 0 | 0.4 | 0 | 4.6 | 0 | 44.8 | 13.6 | 0 | 0 | 3.8 | 65.8 | 0 | 0 | 11 | 2.75 | 5.5 | 0.25 | 2.75 | 0.25 | 38.5 | 7.5 | 0 | 0 | 0 | 0 | 1 | 0.75 |
| March | 12 | 0 | 1.2 | 2 | 0.6 | 0 | 18 | 17.4 | 0 | 0 | 0 | 3.4 | 0 | 0 | 13.25 | 1 | 0.75 | 6.25 | 0 | 3 | 1.5 | 2.25 | 0 | 9.5 | 0.5 | 0 | 1.75 | 15.25 |
| March | 13 | 0 | 2.8 | 3.6 | 1.8 | 3 | 16.6 | 17.6 | 0 | 0 | 0 | 1.2 | 0 | 0 | 1 | 1 | 7.25 | 2.5 | 0 | 0.75 | 0.25 | 9 |  | 1.25 | 0.25 | 0 | 0.25 | 3.25 |
| March | 14 | 0.6 | 0 | 7 | 14.4 | 0.2 | 0 | 6.8 | 0 | 0 | 0.4 | 3.2 | 10 | 0 | 5.75 | 2.75 | 0.75 | 13.25 | 3 | 1 | 8.5 | 11.25 | 0.25 | 0 | 0 | 0.25 | 0.25 | 0.5 |
| March | 15 | 0 | 0.6 | 0.4 | 0 | 4.6 | 1.6 | 0 | 0 | 0 | 0 | 6.4 | 24 | 0 | 2.5 | 19.25 | 3.5 | 9.5 | 6.25 | 16 | 2.5 | 1.75 | 0 | 1 | 0 | 0 | 20.5 | 0 |
| March | 16 | 0 | 0 | 8 | 3 | 0.6 | 15 | 0 | 0 | 4.6 | 1 | 6 | 15 | 0 | 1.5 | 8 | 19 | 2.5 | 22.25 | 36.25 | 0 | 2.5 | 0 | 0.25 | 0 | 0 | 4.5 | 0 |
| March | 17 | 0 | 0 | 0 | 5.8 | 0 | 0 | 1.2 | 0 | 1 | 0.8 | 9.8 | 3.8 | 2.2 | 0.25 | 4 | 1.75 | 0 | 0.75 | 0 | 0 | 13.75 | 2.75 | 0 | 0 | 0 | 0.5 | 0 |
| March | 18 | 1 | 0 | 0 | 4 | 5.4 | 8.6 | 0.8 | 6.6 | 0 | 0 | 0 | 2 | 0 | 0 | 0.25 | 1.25 | 0 | 0 | 0.75 | 0 | 9 | 0 | 0 | 0 | 0 | 19.5 | 0 |
| March | 19 | 0 | 0 | 0.6 | 5.2 | 7.2 | 2.2 | 3 | 15.8 | 11 | 0 | 6 | 4.6 | 0 | 0 | 1.25 | 0.25 | 8.5 | 0.75 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0.75 | 0 |
| March | 20 | 0 | 0 | 0 | 1 | 0.8 | 13.4 | 0 | 0 | 4 | 0 | 1.4 | 2.2 | 0 | 0 | 0.5 | 0 | 3.25 | 1.25 | 17.25 | 0 | 0.25 | 0 | 0 | 0 | 0.5 | 0.75 | 4.75 |
| March | 21 | 0.2 | 2.2 | 4.6 | 0 | 0 | 1.6 | 8.6 | 0 | 0.4 | 0 | 0.4 | 0 | 0 | 0.75 | 0 | 15 | 5.75 | 0 | 7.25 | 2.25 | 2.75 | 0.5 | 0 | 0 | 8.75 | 4.75 | 0.25 |
| March | 22 | 0 | 7 | 0 | 7.6 | 0 | 0 | 27.4 | 0 | 0 | 1.4 | 4.4 | 0 | 0 | 0 | 4.75 | 0.25 | 0 | 0 | 13 | 1.25 | 0.75 | 2.5 | 0 | 0 | 0.75 | 9 | 0 |
| March | 23 | 0 | 8.8 | 1 | 0 | 0 | 0 | 6.2 | 0.2 | 0 | 0 | 9.6 | 0.8 | 0 | 0 | 0.25 | 0 | 0 | 1 | 0.5 | 0 | 1.5 | 0 | 0 | 0 | 0 | 5 | 0.5 |
| March | 24 | 0.8 | 3.4 | 0 | 0 | 0 | 0 | 1 | 7.8 | 0 | 2.6 | 8.4 | 0 | 0 | 0 | 0 | 0 |  | 0 | 1.5 | 4.5 | 2 | 0.25 | 0 | 1.25 | 8.25 | 0.25 | 19.5 |
| March | 25 | 0 | 1.4 | 0 | 0 | 8.4 | 0 | 0.4 | 2.4 | 0 | 0 | 0 | 0.2 | 0 | 5.25 | 0.5 | 0 | 0 | 0 | 0.5 | 1.5 | 1.25 | 0 | 0 | 0 | 0.75 | 0 | 1.5 |
| March | 26 | 1.6 | 0.4 | 15.2 | 0 | 5 | 2.2 | 6.4 | 1.6 | 34.8 | 1 | 0 | 0 | 0 | 0.5 | 1.5 | 0 | O | 8 | 6.75 | 0 | 2 | 3.25 | 2 | 0.75 | 0 | 0 | 0.5 |
| March | 27 | 7.6 | 0 | 0 | 0.2 | 5.6 | 0.6 | 3.6 | 0.2 | 4 | 0 | 0 | 0.6 | 0 | 0 | 0 | 1.75 | 0 | 2.75 | 0 | 3.25 | 1.75 | 0.25 | 0.5 | 5 | 0 | 4.5 | 0 |
| March | 28 | 0 | 4 | 0.4 | 0 | 7.6 | 3.6 | 0.4 | 0 | 4.4 | 0.6 | 0 | 4.6 | 0 | 2.75 | 0 | 0 | 0 | 0.5 | 1.25 | 7 | 20.75 | 0 | 0 | 9.25 | 14.75 | 0.75 | 0 |
| March | 29 | 0.2 | 0 | 5.4 | 0 | 1.8 | 0 | 0 | 0 | 3.8 | 0 | 0 | 0 | 0.2 | 40 | 0 | 2 | 0 | 7 | 0.25 | 0 | 11.75 | 0.75 | 0 | 1.25 | 0 | 7 | 0 |
| March | 30 | 0 | 3 | 6.8 | 0 | 0.8 | 0 | 0.6 | 2.2 | 0 | 0.4 | 0.4 | 0.6 | 0 | 1 | 2 | 16.5 |  | 5 | 11.75 | 0 | 0 | 0 | 0 | 12 | 0 | 0 | 0 |
| March | 31 | 0.6 | 6.2 | 0 | 0 | 2 | 0 | 2.2 | 2.8 | 5.6 | 0 | 1.4 | 15 | 0 | 0 | 10.25 | 10.75 | - | 0.5 | 0.25 | 0 | 0 | 0 | 0 | 0.25 | 0 | 0.5 | 4.25 |
| Total Monthly Rainfall |  | 24.6 | 72.8 | 110 | 64.4 | 81 | 135.8 | 149.6 | 97 | 86 | 36.6 | 162.2 | 99.2 | 3.6 | 92 | 108 | 130 | 91 | 113 | 123 | 137 | 158 | 33 | 15 | 40 | 41 | 117 | 70 |
| Maximum Daily Rainfall |  | 8.8 | 11.8 | 32 | 14.4 | 11.2 | 44.8 | 27.4 | 43.8 | 34.8 | 15.8 | 65.8 | 24 | 2.2 | 40 | 19.25 | 19 | 18 | 22.25 | 36.25 | 38.5 | 20.75 | 14 | 9.5 | 12 | 14.75 | 23.25 | 19.5 |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| April | 1 | 9 | 0 | 0 | 0 | 1.2 | 0 | 0 | 0 | 7.6 | 13.4 | 0.4 | 0 | 18.6 | 0 | 4.25 | 2.25 | 0.00 | 0.00 | 0.25 | 0.00 | 4.50 | 0.0 | 0 | 4.75 | 0.00 | 0.25 |  |
| April | 2 | 0 | 0 | 1.4 | 0 | 0 | 0 | 0 | 0 | 0.6 | 0 | 0 | 0 | 18.2 | 16 | 16 | 3 | 0 | 0 | 1 | 0 | 0 | 0.0 | 0.5 | 0 | 0 | 0 |  |
| April | 3 | 0 | 0 | 1.2 | 0 | 0 | 0 | 0.4 | 0 | 2.2 | 5.6 | 0 | 0 | 0 | 3.5 | 16.5 | 0.0 | 0.0 | 0.0 | 1.0 | 4.8 | 0.0 | 0.0 | 3.5 | 0.8 | 0.0 | 18.3 |  |
| April | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 4.6 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0.25 | 1.00 | 0.00 | 0.50 | 0.00 | 0.25 | 4.00 | 0.0 | 0 | 0.00 | 0.00 | 30.00 |  |
| April | 5 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 9.2 | 0 | 0.2 | 0 | 0 | 0 | 15 | 0 | 11 | 0 | 0 | 1 | 12 | 18.8 | 1 | 0 | 0 | 3 |  |
| April | 6 | 0 | 0 | 0 | 3.4 | 2.4 | 0 | 1.8 | 0 | 1.6 | 0 | 0 | 0.4 | 0 | 0.5 | 10.75 | 0.75 | 2.50 | 3.50 | 0.00 | 0.00 | 3.00 | 17.8 | 2 | 0.00 | 0.00 | 0.00 |  |
| April | 7 | 0 | 0 | 2.2 | 0 | 6.4 | 2 | 3.8 | 0 | 4.6 | 0 | 4.6 | 4.6 | 0 | 8 | 2.75 | 6.75 | 3.75 | 0.00 | 0.00 | 0.00 | 10.50 | 1.3 | 0 | 0.00 | 0.00 | 0.00 |  |
| April | 8 | 0 | 0.6 | 0.6 | 0 | 0 | 0 | 2.6 | 0 | 0 | 3.6 | 6 | 1.4 | 0 | 5.75 | 0 | 0 | 12 | 0 | 0 | 0 | 3 | 11.8 | 11.5 | 0 | 5 | 2 |  |
| April | 9 | 0 | 0 | 0.2 | 0 | 0 | 2.4 | 1.8 | 0 | 0 | 5.6 | 2 | 0 | n/a | 0.5 | 0 | 0 | 0 | 12 | 0 | 0 | 0 | 3.0 | 3.5 | 0 | 0 | 2 |  |
| April | 10 | 0 | 1.6 | 0 | 0 | 4 | 1 | 0 | 0 | 0.4 | 4.2 | 1.2 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0.0 | 2 | 0 | 9 | 26 |  |
| April | 11 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 4.5 | 0.0 | 14.8 | 0.0 | 4.5 | 0.0 | 0.0 | 0.5 | 8 | 0.0 | 0.5 | 0.5 |  |
| April | 12 | 0 | 0.8 | 0.4 | 0 | 1.2 | 0.2 | 1 | 0 | 0.4 | 0.2 | 0 | 0 | 17.4 | 0 | 0.75 | 4.50 | 0.25 | 0.00 | 8.50 | 4.25 | 20.50 | 1.0 | 0.25 | 0.00 | 0.00 | 2.50 |  |
| April | 13 | 0 | 0 | 0 | 1.6 | 0.6 | 11.2 | 13.6 | 0 | 0.6 | 4.8 | 1.4 | 0.2 | 3.6 | 0 | 1.75 | 2.25 | 5.50 | 0.00 | 0.25 | 0.00 | 6.25 | 0.5 | 6.75 | 0.00 | 0.00 | 0.50 |  |
| April | 14 | 0 | 0 | 0 | 0.8 | 0 | 32 | 0.2 | 0 | 0 | 8.6 | 0.4 | 10.8 | 31.4 | 0 | 1.25 | 0.00 | 6.50 | 0.00 | 1.00 | 2.50 | 0.00 | 24.5 | 0 | 0.00 | 0.00 | 0.00 |  |
| April | 15 | 0 | 0 | 0 | 0.4 | 0 | 9 | 0.8 | 2.2 | 8 | 0.4 | 0 | 0 | n/a | 0.5 | 7.5 | 0.0 | 0.8 | 0.0 | 1.5 | 0.0 | 10.0 | 3.5 | 0 | 0.0 | 0.0 | 0.0 |  |
| April | 16 | 0.8 | 0 | 0 | 0 | 0 | 3.2 | 0.8 | 0 | 5 | 0 | 3.4 | 0 | 0 | 0 | 1.25 | 0.00 | 3.00 | 4.25 | 0.00 | 0.00 | 0.00 | 1.3 | 0 | 0.00 | 0.00 | 0.00 |  |
| April | 17 | 1.2 | 0 | 0 | 0 | 12.2 | 0.6 | 0.6 | 0 | 0.2 | 2.4 | 0.8 | 0 | 4.4 | 3.25 | 0 | 9 | 1 | 7 | 0 | 0 | 1 | 12.3 | 0 | 0 | 0 | 0 |  |
| April | 18 | 3.2 | 0.6 | 0 | 0 | 2.2 | 0 | 0.2 | 0 | 1.6 | 0.2 | 1 | 6.6 | 0 | 0 | 0 | 1 | 0 | 11 | , | 0 | 1 | 0.5 | 11.25 | 0 | 0 | 1 |  |
| April | 19 | 2.8 | 0 | 4.4 | 0 | 0 | 0 | 2.8 | 0 | 0 | 0 | 0 | 22.6 | 0 | 3 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0.0 | 0.75 | 0 | 0 | 1 |  |
| April | 20 | 21.8 | 0 | 2.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0.4 | 0 | 0 | 0 | 11.25 | 1.25 | 12.25 | 6.50 | 2.00 | 0.00 | 0.00 | 0.50 | 0.0 | 0 | 0.00 | 0.00 | 1.75 |  |
| April | 21 | 0 | 1 | 0 | 0 | 0 | 0.2 | 1.4 | 0 | 0 | 1.2 | 0 | 8 | 0 | 0.5 | 1.5 | 0.3 | 0.8 | 1.0 | 0.0 | 0.0 | 0.0 | 3.8 | 0 | 0.0 | 0.0 | 1.3 |  |
| April | 22 | 0 | 0 | 4.4 | 4.4 | 3 | 0 | 0 | 0 | 0 | 0 | 0.8 | 0 | 1 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 3 | 1.5 | 3.25 | 17 | 0 | 3 |  |
| April | 23 | 4 | 0 | 0 | 1 | 3.6 | 0 | 1 | 0 | 0 | 0 | 0 | 1.4 | 0 | 7.25 | 0 | 0 | 0 | 2 | 0 | 1 | 1 | 0.0 | 0 | 0 | 0 | 0 |  |
| April | 24 | 1.6 | 0.2 | 0 | 0 | 0 | 0 | 7.6 | 0 | 0.6 | 0 | 3.8 | 0 | 0 | 3.75 | 0 | 0 | 0 | 7 | 2 | 0 | 0 | 0.0 | 0 | 0 | 0 | 3 |  |
| April | 25 | 0.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.8 | 0 | 0 | 0 | 2 | 5 | 0 | 11 | 12 | 0 | 0 | 0.0 | 0 | 3 | 13 | 4 |  |
| April | 26 | 0.6 | 0 | 1.2 | 0 | 0 | 9.4 | 0 | 0 | 0 | 0 | 2.6 | 0 | 0 | 0 | 8.25 | 3.25 | 0.00 | 0.25 | 0.75 | 0.00 | 0.25 | 0.0 | 0 | 4.50 | 10.00 | 1.00 |  |
| April | 27 | 1 | 0 | 0 | 0.4 | 0 | 0.6 | 1.2 | 2.2 | 0 | 0 | 11.4 | 3 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 5 | 0.0 | 0 | 1 | 1 | 6 |  |
| April | 28 | 4.2 | 0 | 0 | 0 | 2 | 0 | 0 | 0.4 | 0 | 0 | 0 | 6.4 | 0 | 2.25 | 10.75 | 0.00 | 0.00 | 2.25 | 2.25 | 0.00 | 0.00 | 0.0 | 0 | 1.25 | 0.00 | 0.00 |  |
| April | 29 | 1.8 | 0 | 0 | 0.2 | 1.4 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 3.5 | 0.0 | 0.8 | 0.0 | 0.5 | 0.0 | 4.8 | 6.0 | 0 | 0.8 | 0.0 | 0.8 |  |
| April | 30 | 2 | 0 | 0 | 0.2 | 4.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.75 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0.3 | 0 | 0 | 0 | 12 |  |
| Total Monthly Rainfall |  | 54.8 | 5.8 | 18.4 | 12.4 | 47.6 | 71.8 | 46.2 | 4.8 | 43.6 | 55.6 | 40.8 | 70.4 | 94.6 | 67.75 | 109.75 | 53.5 | 70.5 | 66 | 35 | 14 | 90.75 | 108 | 54.25 | 32.75 | 37.25 | 118 |  |
| Maximum Daily Rainfall |  | 21.8 | 1.6 | 4.4 | 4.4 | 12.2 | 32 | 13.6 | 2.2 | 9.2 | 13.4 | 11.4 | 22.6 | 31.4 | 16 | 16.5 | 12.25 | 14.75 | 11.75 | 11.50 | 4.75 | 20.50 | 24.50 | 11.50 | 16.75 | 12.75 | 30 |  |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| May | 1 | 0.8 | 0 | 0.4 | 2.2 | 3.6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.75 | 0 | 4.25 | 0 | 0 | 0 | 0 | 2.5 | 1.5 | 0 | 0 | 3.25 | 0 |  |
| May | 2 | 0 | 0 | 5.2 | 0 | 0 | 0 | 0 | 0.2 | 2 | 0 | 2.8 | 0.4 | 4 | 15.75 | 0 | 0.75 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 1.75 | 0 | 1.5 |  |
| May | 3 | 0 | 0 | 0.2 | 2 | 0 | 0 | 1.4 | 1 | 0 | 0 | 2.2 | 2.8 | 0 | 8.25 | 9.75 | 0 | 0 | 0 | 0 | 0.75 | 2 | 0 | 0 | 2 | 0 | 0.25 |  |
| May | 4 | 0.6 | 0 | 2 | 3 | 2.2 | 0 | 11.4 | 0.6 | 0.4 | 0 | 0.2 | 0 | 3.2 | 6.75 | 0 | 10.75 | 0 | 1.75 | 0 | 0.25 | 0 | 0 | 0 | 0 | 1.5 | 1 |  |
| May | 5 | 15.8 | 0 | 0 | 2 | 4 | 1 | 4.4 | 0.6 | 0 | 0 | 0.2 | 0 | 8.6 | 0 | 0 | 0 | 0 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9.25 |  |
| May | 6 | 3.6 | 0 | 0 | 1.6 | 1 | 0 | 0 | 0 | 0 | 0.8 | 1 | 0 | 11.8 | 2.75 | 2 | 0 | 0 | 5 | 3.5 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0 |  |
| May | 7 | 0 | 0 | 0.4 | 0 | 0 | 0 | 0 | 1 | 0 | 3.6 | 0 | 0 | 0 | 0 | 4.75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| May | 8 | 0 | 0 | 3.4 | 1.6 | 0 | 0 | O | 0.2 | 0.8 | 0 | 0 | 0 | 0 | 0 | 1.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 1.25 |  |
| May | 9 | 0 | 0 | 1.4 | 0 | 0 | 0 | 0 | 0 | 0.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 |  |
| May | 10 | 0 | 0 | 0.4 | 11 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0.2 | 3.6 | 2.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.75 | 0 | 0 | 0.5 | 0 |  |
| May | 11 | 0 | 0.2 | 1.4 | 0.4 | 0 | 0 | 0 | 0.6 | 0 | 0 | 0 | 0.2 | 2.8 | 0 | 0.5 | 0 | 0 | 0 | 0 | 0 | 10.25 | 8.25 | 0 | 1.75 | 0 | 0 |  |
| May | 12 | 0 | 0.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.4 | 0 | 5.75 | 0 | 2.5 | 0 | 0 | 0 | 1.25 | 0 | 0 | 7.25 | 0 | 3 |  |
| May | 13 | 0 | 0 | 0.4 | 0 | 0 | 1 | 0 | 0 | 0.2 | 0 | 0 | 5.2 | 10.2 | 0 | 0 | 0 | 5.5 | 0 | 0 | 0 | 6.75 | 0 | 0 | 5.75 | 0 | 0 |  |
| May | 14 | 0 | 3.6 | 0.2 | 0 | 6 | 0.2 | 1.4 | 0 | 0.2 | 0 | 0 | 0.2 | 9 | 0 | 0 | 0 | 3.25 | 0 | 0 | 0 | 0 | 0 | 4.75 | 0 | 0 | 2.5 |  |
| May | 15 | 0 | 1 | 0 | 0 | 1 | 0 | 0.2 | 0 | 0.8 | 0 | 0 | 0 | 0 | 0 | 9.75 | 0 | 0 | 0 | 0.5 | 0 | 8.25 | 0 | 1.75 | 0 | 0 | 14.25 |  |
| May | 16 | 0 | 0 | 0.2 | 0 | 5.2 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 20.75 | 0 | 1.75 | 0 | 0 | 0 | 2.75 | 0 | 0 | 10 | 0 | 1 |  |
| May | 17 | 0 | 0 | 9.4 | 0 | 0 | 6.2 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.75 | 0 | 0 |  |
| May | 18 | 0 | 0 | 1.8 | 0 | 0 | 0 | 0 | 0 | 0.4 | 0 | 0 | 0 | 7.8 | 4.5 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3.5 | 5 |  |
| May | 19 | 0 | 0 | 0 | 7.2 | 0 | 0 | 2.4 | 0 | 1.6 | 0 | 1.6 | 1.8 | 0 | 1.5 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| May | 20 | 0 | 0 | 0 | 0 | 0 | 1.8 | 0.6 | 0 | 0 | 0 | 10.4 | 0.2 | 0 | 7.25 | 0 | 0 | 0 | 0 | 0 | 0.75 | 0 | 0 | 9.75 | 0 | 0 | 0 |  |
| May | 21 | 3 | 0.6 | 0 | 4.4 | 0 | 0.4 | 0.4 | 0 | 0.4 | 2.6 | 0 | 4.6 | n/a | 0 | 0 | 0.5 | 4.75 | 0 | 0 | 0 | 0 | 0 | 1.25 | 8 | 0 | 0 |  |
| May | 22 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 9.4 | 0.4 | 17.4 | 0 | 0 | 0 | 1 | 1.5 | 5.75 | 6.75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| May | 23 | 3.6 | 0 | 0 | 0 | 0 | 0 | 1 | 1.2 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 8 | 0 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| May | 24 | 0 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 |  | 0 | 1.25 | 3.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| May | 25 | 0 | 6.8 | 0 | 0 | 0 | 2.6 | 0 | 1.2 | 0 | 3 | 0 | 0 | 0 | 1.5 | 0 | 1.25 | 0 | 0 | 0 | 0 | 0 | 0 | 13.5 | 7.75 | 4.75 | 0.5 |  |
| May | 26 | 0 | 1.8 | 0 | 9.6 | 0 | 2.4 | 0 | 0.6 | 0 | 6.2 | 0 | 0.2 | 1.2 | 7 | 2.25 | 0 | 0.25 | 2 | 0 | 0 | 0 | 0 | 0.25 | 0 | 4 | 6.25 |  |
| May | 27 | 0.4 | 22.8 | 0 | 7.2 | 0 | 0 | 0.2 | 9 | 0 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 3.5 | 0.75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3.5 |  |
| May | 28 | 4.8 | 2.4 | 0 | 0.8 | 3.2 | 2.4 | 0 | 3.2 | 0 | 0 | 0 | 0 | n/a | 6.25 | 2.5 | 0 | 5.25 | 3 | 0 | 9.75 | 0 | 0 | 0 | 0 | 4.25 | 0.5 |  |
| May | 29 | 6 | 0 | 0 | 0 | 0 | 1.4 | 0 | 7.6 | 0 | 0 | 0 | 0 | 0 | 0.75 | 0.5 | 0.25 | 1.5 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |  |
| May | 30 | 5.2 | 0 | 0 | 6.2 | 1.2 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 3.75 | 0 | 0 | 4.25 | 0 | 0 | 0 | 5 | 0 | 0 | 2.5 | 0 | 0.25 |  |
| May | 31 | 14 | 0 | 0 | 0.8 | 0 | 0 | 0 | 0 | 0.4 | 0 | 0 | 0 | 0 | 6.5 | 0 | 8.5 | 1 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0 |  |
| Total Monthly Rainfall |  | 58.8 | 40.4 | 26.8 | 60 | 27.4 | 19.4 | 25.4 | 38.4 | 9.2 | 34.2 | 18.4 | 15.8 | 62.6 | 76.5 | 61.75 | 32 | 53.5 | 40.25 | 4 | 11.75 | 41.25 | 10.5 | 31.25 | 47.5 | 22 | 50.25 |  |
| Maximum Daily Rainfall |  | 15.8 | 22.8 | 9.4 | 11 | 6 | 6.2 | 11.4 | 9.4 | 2 | 17.4 | 10.4 | 5.2 | 11.8 | 15.75 | 20.75 | 10.75 | 8 | 15 | 3.5 | 9.75 | 10.25 | 8.25 | 13.5 | 10 | 4.75 | 14.25 |  |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| June | 1 | 0.2 | 0 | 0.6 | 0 | 8.4 | 0 | 0 | 0 | 0.4 | 5.4 | , | 0 | , | 0 | 0 | 1.75 | 0.5 | , | 0 | 7.5 | 0 | 0 | 0 | , | , | 0 |  |
| June | 2 | 0.4 | 0 | 0 | 0 | 2.8 | 0 | 0 | 0 | 0 | 0.6 | 0 | 0.4 | 0 | 0 | 0 | 5.25 | 0.25 | 0 | 0.25 | 0.25 | 0.75 | 0.25 | 0 | 0.75 | 0 | 0 |  |
| June | 3 | 8.2 | 0 | 0 | 0 | 0.8 | 0 | 0 | 0 | 0 | 0 | 1.4 | 4.4 | 0 | 0 | 1 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0.25 | 0 | 0 | 0 | 4.5 |  |
| June | 4 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.4 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.75 | 0 | 6 |  |
| June | 5 | 0 | 0 | 4 | 1.6 | 0 | 2.4 | 0 | 4.8 | 0 | 0 | 0.8 | 2.8 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 1 | 0.25 |  |
| June | 6 | 0 | 0 | 0.4 | 3.6 | 0.6 | 0 | 0 | 1 | 0 | 0 | 0.8 | 5.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 3.75 | 1 | 0 |  |
| June | 7 | 0.6 | 0 | 1.6 | 0.4 | 0 | 0 | 0 | 0.2 | 0 | 6.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.75 | 0.25 | 0.5 | 0.25 |  |
| June | 8 | 0 | 0 | 0.6 | 0 | 0 | 0 | 0 | 0 | 0.4 | 16.8 | 0 | 0 | 0 | 0 | 0 | 15 | 0.25 | 0 | 0 | 0 | 11.25 | 0 | 0 | 0 | 4 | 0 |  |
| June | 9 | 0 | 0 | 0 | 0 | 10 | 0 |  | 5 | 0 | 0 | 9 | 10.2 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0 |  | 0 | 8 | 0 | 22.5 | 0 | 24.75 |  |
| June | 10 | 0 | 10.8 | 0 | 2.4 | 1.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5.5 | 0.5 | 0 | 0.75 | 1 | 0.25 |  |
| June | 11 | 3 | 0 | 0 | 2.6 | 1.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| June | 12 | 0.2 | 0 | 0 | 15.8 | 1.2 | 0 | 0 | 0.4 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8.75 | 2.25 | 1 |  |
| June | 13 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0.6 | 8.4 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 1.25 | 0.25 |  |
| June | 14 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 1.5 | 1 | 0 | 3 | 0 | 2 | 0 | 0 | 3.75 | 0 |  |
| June | 15 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4.25 | 14 | 0 | 0 | 3.75 | 0 | 0.5 |  |
| June | 16 | 1.2 | 3.6 | 2.4 | 0 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 11.75 | 0 | 0 | 0 | 0.25 | 11.5 | 7.25 |  |
| June | 17 | 13.4 | 0 | 0 | 0 | 0.4 | 2.6 |  | , |  | 0.6 | 0 | 0 | 1.4 | 0.35 | 0 | 1 | 0 | 6.25 | 0 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0.25 |  |
| June | 18 | 0 | 0 | 0 | 0 | 0 | 0.4 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0.25 | 2.25 | , | 0 | 0 | 0 | 1.25 |  |
| June | 19 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.25 | 1.75 | 0.25 | 0 | 2.75 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |  |
| June | 20 | 0.4 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3.25 | 0 | 0.25 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |  |
| June | 21 | 11.8 | 0 | 1.6 | 0 | 0 | 0 | 0 | 0 | 0.4 | 0 | 5.8 | 0 | 1.4 | 0.35 | 0 | 0 | 2.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| June | 22 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8.5 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| June | 23 | 2.2 | 0 | 6.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 0 | 0 | 0 | 4.75 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| June | 24 | 0 | 7.8 | 7.6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3.2 | 0 | 2.8 | 0.7 | 0 | 4.25 | 1.25 | 0.75 | 0 | 1.25 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| June | 25 | 0.8 | 0 | 0.4 | 0 | 0 | 0 | 0 | 0.4 | 0 | 0 | 0 | 0 | 0.2 | 0.05 | 3 | 0 | 2 | 0 | 0 | 0 | 0 | 7.75 | 0 | 0 | 0 | 0 |  |
| June | 26 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| June | 27 | 0.6 | 0 | 0 | 0 | 0.8 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1.4 | 0.35 | 0 | 0 | 1.5 | 0 | 0 | 0 | 0 | 0 | 16.75 | 1.75 | 0 | 0 |  |
| June | 28 | 0 | 0 | 7.4 | 0 | 0.2 | 6.6 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0.25 | 0 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 0 |  |
| June | 29 | 27.4 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 4.4 | 0 | 0 | 0 | 0 | 2.5 | 0 | 1.75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| June | 30 | 0 | 0 | 3.6 | 0.8 | 0 | 0 | 1.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | 0 | 0 |  |
| Total Monthly Rainfall Maximum Daily Rainfall |  | 73.8 | 23.2 | 37 | 27.6 | 28 | 18 | 2 | 20.2 | 1.6 | 30.4 | 34.4 | 24.2 | 7.2 | 1.8 | 7.5 | 44.75 | 30.25 | 12 | 3.5 | 41.5 | 34 | 19.25 | 19 | 45.25 | 26.25 | 46.5 |  |
|  |  | 27.4 | 10.8 | 7.6 | 15.8 | 10 | 6.6 | 1.4 | 8.4 | 0.4 | 16.8 | 9 | 10.2 | 2.8 | 0.7 | 3 | 15 | 12 | 6.25 | 2.75 | 11.75 | 14 | 8 | 16.75 | 22.5 | 11.5 | 24.75 |  |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| July | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0.5 | 0 | 3 | 0 | 0 | 0 | 0 |  |
| July | 2 | 0 | 0 | 3.2 | 0.2 | 0 | 0 | 0 | 3.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8.25 | 0 | 0 | 0 |  |
| July | 3 | 0 | 20.6 | 3.2 | 1.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.2 | 0 | 0 | 0.25 | 5.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 0 | 16.5 |  |
| July | 4 | 0 | 1.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4.25 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.75 |  |
| July | 5 | 2.4 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.4 | 0 | 0 | 0 | 0 | O | 0.25 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0 | 0.25 |  |
| July | 6 | 4.8 | 0 | 0 | 0 | 0 | 0 | 0 | 3.4 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.6 | 0.4 | 0 | 0 | 0 | 1 | 0 | 3.5 | 0 | 0 | 1.75 | 0 | 0 | 2.25 |  |
| July | 8 | 23 | 0 | 0 | 0 | 0 | 7.4 | 0 | 0 | 0.6 | 0 | 0 | 0 | 6 | 1.5 | 0 | 0 | 0 | 0 | 0 | 0.75 | 0 | 0 | 0 | 0.25 | 0 | 0 |  |
| July | 9 | 2.4 | 0 | 0 | 0 | 0 | 0 | 0 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11.5 | 0 | 0 | 4 | 3 | 0 | 0 |  |
| July | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.8 | 0 | 1.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 5.25 | 0 | 0 | 0 |  |
| July | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.6 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0.5 | 0 | 0 | 0 |  |
| July | 12 | 0 | 0.2 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 13 | 0 | 0 | 0 | 0 | n/a | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.25 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 14 | 0 | 1.8 | 0.6 | 0 | n/a | 0 | 4.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 1.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 15 | 0 | 9.2 | 0 | 0 | n/a | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9.75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 |  |
| July | 16 | 0 | 0 | 0.8 | 0 | n/a | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.75 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 17 | 0 | 0 | 0.6 | 0 | n/a | 0 | 0 | 0 | 0 | 0 | 5.6 | 0 | 0 | 0 | 12.75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.5 | 0 | 0 | 0.25 |  |
| July | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14.45 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 |  |
| July | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0.4 | 0 | 0 | 0 | 3.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 21 | 2.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9.2 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.5 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 24 | 0 | 0 | 0.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0.75 | 0 | 0 | 0 | 3 | 7.75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 26 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 27 | 0 | 0 | 0 | 3.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 28 | 0 | 0 | 0 | 10.2 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 29 | 0 | 0 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 30 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| July | 31 | 0 | 0.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.4 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| Total Monthly Rainfall |  | 35 | 34.4 | 9.6 | 16.6 | 10 | 7.4 | 11.8 | 31 | 1.2 | 2 | 41.65 | 7 | 10.6 | 2.65 | 26.25 | 21 | 0 | 6.5 | 17.5 | 18.5 | 0 | 3.25 | 22.25 | 10.25 | 0 | 21.5 |  |
| Maximum | Rainfall | 23 | 20.6 | 3.2 | 10.2 | 10 | 7.4 | 4.2 | 21 | 0.6 | 1.8 | 14.45 | 3.8 | 6 | 1.5 | 12.75 | 5.5 | 0 | 3 | 9.5 | 11.5 | 0 | 3 | 8.25 | 7 | 0 | 16.5 |  |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| August | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | 0 |  |
| August | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 | 0 | 4.25 | 0 | 0 | 0 |  |
| August | 3 | 0 | 0 | 0.2 | 0 | 13.6 | 0 | 0 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.75 | 0 | 0 | 0 | 0 | 0 | 0 | 1.25 | 0 | 0 |  |
| August | 4 | 0 | 0 | 0.2 | 0 | 5.4 | 0 | 0 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 5 | 0 | 0 | 7.8 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 6 | 5.4 | 0 | 0 | 0 | 2.4 | 0 | 0 | 25.6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5.5 | 0 | 0 |  |
| August | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 12.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.25 | 1.25 | 0.31 |  |
| August | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0.6 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 4.75 | 1.19 |  |
| August | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4.4 | 2 | 1.25 | 0 | 0.25 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.8 | 0 | 0 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | 0.75 | 0 | 0 | 0 |  |
| August | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.2 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 12 | 0 | 0 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0 | 1.75 | 0.5 | 0 | 0 | 0 | 0 |  |
| August | 13 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.5 | 0 | 0 | 0 | 0 | 0 |  |
| August | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10.75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 15 | 0 | 0 | 8.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.4 | 0 | 0 | 0 | 0 | 0 | 0 | 1.75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 16 | 0 | 5.4 | 3.6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |  |
| August | 17 | 0 | 0 | 0 | 0 | 0 | 0.8 | 0 | 0 | 2.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 18 | 0 | 0.2 | 0 | 3.6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12.2 | 5.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 20 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.8 | 4.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4.75 | 0 | 0 |  |
| August | 21 | 5.4 | 0 | 0 | 0 | 2.2 | 0 | 0 | 11.2 | 0 | 0 | 7.6 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4.75 | 2.5 | 0 | 0 |  |
| August | 22 | 0 | 0 | 0 | 0 | 11.2 | 0 | 0 | 3.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 23 | 0.2 | 0 | 0 | 0 | 6.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 16.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.25 | 0 | 0 | 0 |  |
| August | 24 | 0.2 | 0 | 0 | 0 | 1.8 | 0 | 0 | 30 | 0 | 0 | 0 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0.25 | 0 | 0 |  |
| August | 25 | 3.4 | 0 | 0 | 0 | 0 | 0 | 0 | 31.6 | 0 | 0 | 0 | 0.6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 26 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0.8 | 0 | 0 | 0 | 4.6 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 27 | 7.2 | 0 | 0 | 0 | 0 | 0 | 0 | 2.6 | 0 | 0 | 0 | 5.2 | 0 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.75 | 0.44 |  |
| August | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.4 | 0 | 1.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 29 | 0 | 0 | 3.8 | 4.8 | 0 | 0 | 0 | 0 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 6.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| August | 30 | 0 | 0 | 0 | 1.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6.75 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |  |
| August | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.2 | 0 | 0 | 12 | 0 | 0 | 0 | 0 | 1.25 | 1.75 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| Total Monthly RainfallMaximum Daily Rainfall |  | 27 | 5.6 | 24.4 | 9.6 | 43.2 | 0.8 | 0.6 | 106.2 | 2.8 | 3.4 | 23.2 | 43.6 | 22.2 | 26.25 | 16.5 | 2.25 | 3 | 12.75 | 15.5 | 2.75 | 3.25 | 0.5 | 12.75 | 18.75 | 7.75 | 1.94 |  |
|  |  | 27 | 5.4 | 8.2 | 4.8 | 13.6 | 0.8 | 0.6 | 31.6 | 2.4 | 2.4 | 12.2 | 15 | 12 | 12.25 | 16.5 | 2 | 2.75 | 10.75 | 6.75 | 1.75 | 1.75 | 0.5 | 4.75 | 5.5 | 4.75 | 1.19 |  |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| September | 1 | 0 | 0 | 0 | 0 | 5.2 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 8.5 | 3.5 | 0 | 0 | 0 | 0 | 0.25 | 0 |  |
| September | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 9 | 11.25 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| September | 3 | 0 | 0 | 0 | 0 | 1.6 | 0 | 0 | 0 | 0 | 0 | 5.6 | 0 | 1.6 | 0 | 0 | 0 | 0 | 19 | 12.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| September | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8.5 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |  |
| September | 5 | 0.4 | 0 | 2 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.4 | 0 | 0 | 0 | 0 | 0 | 0 | 2.5 | 0 | 0 | 0 | 0 | 5.25 | 0 |  |
| September | 6 | 0 | 0 | 0.4 | 0 | 0 | 0 | 1.2 | 0 | 0 | 0 | 0 | 0 | 13.8 | 10.25 | 0 | 0 | 0 | 0 | 4.5 | 2 | 0 | 0 | 0 | 0 | 0.25 | 0 |  |
| September | 7 | 0 | 0 | 0 | 0.4 | 0 | 0 | 7.6 | 0 | 0 | 0 | 0 | 0 | 5.4 | 7 | 0 | 0 | 0 | 0 | 15.25 | 2.5 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| September | 8 | 0 | 1.2 | 0 | 6 | 0 | 0 | 0 | 2.2 | 0 | 0 | 0 | 0 | 0 | 2.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0.5 | 6 | 0 | 0 | 0 | 0 |  |
| September | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11.6 | 0 | 0 | 12.2 | 0.25 | 0 | 0 | 0 | 0 | 0 | 0 | 7.75 | 1.5 | 1 | 0 | 0.75 | 0 |  |
| September | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 1.6 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 0.5 |  | 1 | 0 | 0 | 0 | 8 | 0 | 0 | 0 | 0 |  |
| September | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0.8 | 21.4 | 0 | 0 | 0 | 0 | 0 | 1.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| September | 12 | 0.2 | 0 | 0 | 0 | 0 | 0 | 0 | 0.2 | 0 | 0 | 0 | 0 | 0 | 14.25 | 0 | 0 | 3.75 | 0 | 0 | 0 | 0 | 2.5 | 2.5 | 0 | 6.75 | 0 |  |
| September | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11.4 | 0 | 0.6 | 0 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | 0.5 | 0 |  |
| September | 14 | 6.4 | 0 | 0 | 0 | 0 | 0 | 0.8 | 0.8 | 0 | 0.6 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8.25 | 4.75 | 0 | 0 |  |
| September | 15 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 17.2 | 0 | 0.2 | 0 | 0 | 0 | 1.25 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 13.5 | 2 | 3.5 | 0 |  |
| September | 16 | 9.6 | 0 | 0 | 0 | 0 | 0 | 10 | 2.2 | 0 | 0 | 5.2 | 0 | 5.4 | 2.75 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 4.5 | 0.75 | 0.25 | 0 | 0 |  |
| September | 17 | 9.6 | 0 | 0 | 0 | 0 | 0 | 0 | 6.8 | 0 | 3.2 | 0 | 0 | 0 | 13.25 | 0 | 0 | 0 | 0 | 1.25 | 13.25 | 1 | 4 | 20.75 | 0 | 0 | 0 |  |
| September | 18 | 5.4 | 0.2 | 0 | 0 | 0 | n/a | 8 | 5.8 | 0 | 4.4 | 2.4 | 0 | 0 | 18 | 0.25 | 0 | 0 | 0.25 | 1.25 | 0 | 1.25 | 0 | 1.25 | 0 | 43.5 | 0 |  |
| September | 19 | 0 | 0 | 0 | 0 | 0 | n/a | 4.6 | 7.6 | 0 | 1.2 | 0 | 0 | 6.2 | 10.75 | 2.25 | 0 | 0 | 0.75 | 0.25 | 5 | 2.75 | 0 | 0.5 | 0 | 8.5 | 0 |  |
| September | 20 | 0 | 0 | 0 | 3.2 | 0 | n/a | 0 | 0 | 0 | 10 | 1 | 0.2 | 0 | 0.5 | 0 | 0 | 0 | 0 | 0.75 | 0 | 0.25 | 0 | 0 | 0 | 5.75 | 0 |  |
| September | 21 | 0 | 0 | 0 | 0 | 5.8 | n/a | 0 | 0 | 0 | 0 | 0.6 | 0 | 0 | 0 | 0 | 0 | 8.75 | 0 | 4.5 | 0 | 0 | 5.25 | 0 | 2.25 | 0 | 0 |  |
| September | 22 | 0 | 0 | 0 | 0 | 0 | n/a | 0 | 1.4 | 0 | 0 | 0.6 | 0 | 0 | 0 | 1.75 | 0 | 0.25 | 0 | 0 | 0 | 0 | 6.25 | 6.25 | 0 | 0 | 0 |  |
| September | 23 | 0 | 0 | 0.4 | 0 | 0 | n/a | 0 | 4.2 | 0 | 0 | 0 | 0 | 0 | 10.5 | 11.5 | 0 | 12.5 | 0 | 0 | 2 | 0 | 2 | 12.5 | 27.5 | 0 | 0 |  |
| September | 24 | 0 | 0 | 0 | 0 | 0 | n/a | 0 | 0 | 0 | 0 | 1 | 11.6 | 0 | 1.5 | 0 | 0 | 3.5 | 12.5 | 0 | 0.5 | 0 | 0 | 0.25 | 20 | 0 | 0 |  |
| September | 25 | 0 | 2.2 | 6.2 | 0 | 0 | n/a | 0 | 0 | 0 | 0 | 0.4 | 8.4 | 0 | 7.75 | 0 | 0 | 9.25 | 3 | 4.5 | 0.25 | 3.25 | 0 | 0 | 26 | 0 | 0 |  |
| September | 26 | 23 | 0 | 0.8 | 0 | 14.6 | n/a | 0 | 0 | 0 | 0 | 0 | 1.2 | 0 | 17.5 | 3 | 0 | 4.75 | 1.25 | 2.75 | 0 | 0 | 0 | 0.5 | 11.5 | 0 | 0 |  |
| September | 27 | 12.4 | 0 | 0 | 0 | 0 | n/a | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 24 | 0 | 0 | 12 | 0.25 | 0 | 0 | 0 | 12 | 0 | 36 | 0 |  |
| September | 28 | 8.6 | 0 | 0 | 0 | 0.6 | n/a | 0 | 0 | 0 | 0 | 0.6 | 0 | 0.8 | 0 | 1 | 0 | 6.5 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 4.25 | 0.25 |  |
| September | 29 | 0 | 0 | 0 | 15.6 | 0.8 | n/a | 0 | 0 | 0.4 | 0 | 1 | 0 | 8 | 0 | 0 | 0 | 4.75 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 3 | 0 |  |
| September | 30 | 11 | 0 | 0 | 1.4 | 0 | n/a | 0 | 0 | 0 | 0 | 8.8 | 0 | 0 | 0 | 0 | 0 | 3.25 | 1.5 | 0 | 0 | 1.5 | 0 | 0 | 0 | 5.5 | 0 |  |
| Total Monthly Rainfall Maximum Daily Rainfall |  | 89.6 | 3.6 | 9.8 | 26.8 | 28.6 | 0 | 34.6 | 83.2 | 0.4 | 31.8 | 27.2 | 21.4 | 54.8 | 133.75 | 43.75 | 0.5 | 60.25 | 51.25 | 76.5 | 42.75 | 19.25 | 40.5 | 80.25 | 94.25 | 123.75 | 1.25 |  |
|  |  | 23 | 2.2 | 6.2 | 15.6 | 14.6 | 0 | 10 | 21.4 | 0.4 | 11.6 | 8.8 | 11.6 | 13.8 | 18 | 24 | 0.5 | 12.5 | 19 | 15.25 | 13.25 | 7.75 | 8 | 20.75 | 27.5 | 43.5 | 1 |  |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| October | 1 | 3 | 0 | 0 | 0 | 0 | n/a | 0 | 0 | 0 | 0 | 0 | 0 | 3.4 | 0 | 0 | 0 | 9.25 | 0 | 0 | 0.5 | 0 | 11 | 0 | 0 | 16.25 | 0 |  |
| October | 2 | 0.4 | 9.6 | 0 | 0 | 0 | n/a | 0 | 0 | 0 | 0 | 10.2 | 2 | 0 |  | 0 | 0 | 7 | 0 | 0 | 0.5 | 0 | 3 | 0.75 | 0 | 0 | 0 |  |
| October | 3 | 12 | 2.6 | 0 | 0 | 0 | n/a | 0 | 0 | 0 | 0 | 1 | 13.4 | 0 | 0 | 0.5 | 0 | 5.75 | 0 | 2.25 | 4.25 | 0 | 0.5 | 0.25 | 0.25 | 0 | 0 |  |
| October | 4 | 15 | 0.4 | 0 | 0 | 0 | n/a | 0 | 0 | 0 | 0 | 1 | 6.4 | 0 | 0 | 0.25 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 9.25 | 0.25 | 0 | 0 |  |
| October |  | 1 | 0 | 21 | 0 | 0 | 0 | 0 | 4.8 | 0.4 | 0 | 0 | 1.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 0 | 2.5 | 0.5 | 0 | 0 |  |
| October | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 4.6 | 5.2 | 0 | 1.6 | 9 | 1.8 | 0 | 0 | 2 | 0 |  | 0 | 0 | 8 | 2 | 8.75 | 0 | 0 | 1.25 | 0 |  |
| October | 7 | 8 | 0 | 9.2 | 0 | 0 | 0 | 8.2 | 3.6 | 0.4 | 0 | 10.4 | 2.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12.25 | 0 | 0 | 9 | 0.25 | 5.5 | 0 |  |
| October | 8 | 13.6 | 3.4 | 23.6 | 1.4 | 2 | 0 | 7.4 | 19.2 | 0 | 5.2 | 0 | 0 | 0 | 2.75 | 0.5 | 0 | 1.75 | 0 | 11.75 | 5 | 0 | 8.5 | 0.75 | 0.25 | 0 | 0 |  |
| October | 9 | 0.2 | 1.8 | 1 | 5.2 | 0 | 0 | 7.6 | 3.4 | 0 | 0 | 0.4 | 0.6 | 0 | 23 | 0.5 | 0 | 0 | 0 | 0 | 0.25 | 0 | 5.25 | 0 | 16.5 | 0 | 0 |  |
| October | 10 | 12.2 | 0 | 0 | 3 | 16.2 | 0 | 2.8 | 0 | 0.4 | 0 | 0 | 0 | 0 | 3.5 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0.25 | 0 | 6.5 | 5.25 | 0 |  |
| October | 11 | 0.8 | 0 | 0 | 0 | 0 | 0 | 1.6 | 0 | 0 | 0 | 0 | 0 | 0 | 2.25 | 8 | 0.25 | 0.25 | 1.25 | 12.75 | 0 | 0.25 | 0 | 0 | 34.25 | 4.25 | 0 |  |
| October | 12 | 2.2 | 13.2 | 3.2 | 0 | 7 | 0 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0.5 | 15.75 | 0 | 0 | 4.25 | 0 | 4.5 | 3.25 | 0 | 5 | 2 | 0.25 | 5 |  |
| October | 13 | 2.4 | 3 | 7.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | n/a | 8.2 | 0.8 | 0 | 2.5 | 3.5 | 4 | 1 | 6.75 | 3.25 | 1.75 | 0 | 0 | 28.5 | 2 | 0 |  |
| October | 14 | 6.6 | 14.2 | 0 | 0 | 7.4 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 3.2 | 0 | 0 | 17.75 | 0 | 7 | 0 | 2.5 | 0 | 0 | 0 | 0 | 0.75 | 19.75 |  |
| October | 15 | 2 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 10 | 0 | 7.6 | 3.2 | 0 | 0 | 3 | 0 | 0 | 0.25 | 9.25 | 0 | 0 | 3.25 | 0 | 2.5 | 0 |  |
| October | 16 | 0 | 0 | 0 | 8 | 0.4 | 0 | 100 | 0 | 0 | 1.6 | 0.2 | 7 | 14.2 | 0 | 0 | 14.25 | 0 | 3 | 0 | 2.75 | 0.5 | 0 | 10.5 | 0 | 7.25 | 0 |  |
| October | 17 | 0.8 | 7.6 | 0 | 24.8 | 0 | 0 | 104.4 | 27.2 | 0 | 0 | 1.2 | 0 | 48.6 | 0 | 0 | 0 | 0 | 0.75 | 0 | 2.25 | 13 | 0 | 11.25 | 0.75 | 8.25 | 0 |  |
| October | 18 | 0 | 0 | 0 | 5.8 | 1.2 | 0 | 0.6 | 5.2 | 0 | 11.8 | 22.6 | 0 | 0 | 0 | 0 | 0 | 0 | 4.5 | 0 | 3.75 | 5.75 | 0 | 11.5 | 21.75 | 4.5 | 0 |  |
| October | 19 | 0 | 0 | 0 | 0.2 | 3.2 | 1.6 | 12.8 | 10.2 | 0 | 8.2 | 27.4 | 0 | 0 | 0 | 0 | 19.75 | 0.5 | 0.25 | 0 | 6.5 | 9.75 | 0 | 2.5 | 0 | 0.25 | 0 |  |
| October | 20 | 0 | 0 | 0 | 19.2 | 0 | 1.2 | 85.4 | 3.2 | 0 | 0 | 9.8 | 1.8 | 0 | 0 | 0.25 | 1 | 0.5 | 6.25 | 0.25 | 23 | 2 | 0 | 4.5 | 0 | 0 | 0 |  |
| October | 21 | 0 | 0 | 0 | 0 | 14.4 | 0 | 13.6 | 0 | 0 | 0 | 1.6 | 0 | 4.4 | 0 | 0 | 0.75 | 0 | 4.5 | 0.25 | 11.5 | 2.5 | 0 | 22.25 | 0 | 3 | 3.5 |  |
| October | 22 | 0.2 | 0 | 0 | 0 | 10.4 | 0 | 1.6 | 1.6 | 0 | 0 | 5.6 | 0 | 1.4 | 0.25 | 5.25 | 0 | 0.25 | 2 | 0.25 | 0.25 | 5 | 0 | 0 | 0 | 3.25 | 0 |  |
| October | 23 | 0 | 0 | 0 | 0 | 9.6 | 0 | 1.8 | 0 | 0 | 0 | 0 | 0 | 34.8 | 2.25 | 16.25 | 5.75 | 0.5 | 25.75 | 0.25 | 1.5 | 0 | 0 | 0 | 17.5 | 1.5 | 2.75 |  |
| October | 24 | 0 | 0 | 3.2 | 0 | 0.4 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 17.75 | 0 | 3 | 0 | 8.75 | 0 | 0 | 0 | 0.25 | 0 | 0 | 0.25 | 7.75 |  |
| October | 25 | 0 | 0 | 0.4 | 0 | 21.4 | 0 | 0 | 7.2 | 3 | 0.4 | 0 | 0 | 10 | 22.75 | 0 | 3.75 | 0 | 1.5 | 0.5 | 0.25 | 0 | 1 | 0.5 | 0 | 1.25 | 5.5 |  |
| October | 26 | 6.8 | 0 | 0 | 0 | 9.2 | 0 | 0 | 0 | 0 | 0.2 | 0 | 0 | 7.6 | 3.25 | 0 | 0 | 0.25 | 22.75 | 14.5 | 14.25 | 0 | 4 | 0 | 0.75 | 15.75 | 0.75 |  |
| October | 27 | 0 | 0 | 1.2 | 0 | 1.6 | 1.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.75 | 9.75 | 0 | 9.75 | 2.25 | 4.75 | 0 | 6.25 | 0 | 0 | 4.75 | 21.75 |  |
| October | 28 | 4.2 | 0 | 20 | 4.6 | 0 | 0 | 0 | 1.6 | 11 | 0 | 0 | 0 | 8.4 | 2.75 | 4.25 | 21.25 | 0 | 5.25 | 0 | 0.25 | 0 | 22.5 | 0 | 0.75 | 3.5 | 4.25 |  |
| October | 29 | 21 | 0 | 8 | 0 | 0 | 0 | 0 | 0.2 | 3.2 | 1 | 0 | 0 | 13.4 | 0.5 | 15 | 1.25 | O | 14.75 | 12.5 | 0.5 | 0 | 8.75 | 0 | 0 | 52 | 0.25 |  |
| October | 30 | 47.2 | 0 | 44.4 | 0 | 6.4 | 0 | 0 | 12.6 | 12 | 0 | 0 | 0.6 | 14.4 | 0.5 | 0 | 22.5 | 0 | 2.75 | 0.75 | 6 | 0 | 3.5 | 0 | 8.25 | 0 | 37 |  |
| October | 31 | 7.8 | 1.8 | 27.4 | 1.8 | 13.6 | 0 | 0 | 0 | 17.8 | 0 | 0 | 2.4 | 3 | 4.25 | 7 | 3.75 | 0.25 | 11.25 | 15.75 | 6 | 0 | 6 | 0 | 0 | 0 | 3.75 |  |
| Total Monthly Rainfall Maximum Daily Rainfall |  | 167.4 | 57.6 | 170.4 | 74 | 124.4 | 4.6 | 384.4 | 105.2 | 48.2 | 48 | 100.4 | 56 | 170.8 | 86.25 | 80.25 | 131.25 | 30.25 | 137.25 | 81 | 135 | 45.75 | 89.5 | 93.75 | 139 | 143.5 | 112 |  |
|  |  | 47.2 | 14.2 | 44.4 | 24.8 | 21.4 | 1.8 | 104.4 | 27.2 | 17.8 | 11.8 | 27.4 | 13.4 | 48.6 | 23 | 16.25 | 22.5 | 9.25 | 25.75 | 15.75 | 23 | 13 | 22.5 | 22.25 | 34.25 | 52 | 21.75 |  |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| November | 1 | 0.8 | 1 | 0 | 0 | 0.6 | 0 | 0 | 15.4 | 5.6 | 0 |  | , |  | 2.75 | 0 | 10 | 1.5 | 3 | 48.5 | 14.5 | 0 | 2.25 | 0 | 0 | 0 | 0.25 |  |
| November | 2 | 0 | 19.4 | 0 | 0 | 0 | 0 | 0.6 | 36.4 | 12.6 | 16.6 | 0 | 1.4 | 0.8 | 0.25 | 0 | 1.25 | 4 | 0 | 16 | 14.5 | 22.25 | 20 | 0 | 0 | 8.5 | 13.5 |  |
| November | 3 | 9.4 | 0 | 0.8 | 0 | 0 | 0 | 0 | 0 | 2.2 | 15.6 | 4 | 3.4 | 0 | 0 | 7.75 | 3 | 3 | 1.75 | 0 | 0.25 | 3.5 | 7 | 0 | 27.25 | 14 | 22.75 |  |
| November | 4 | 0 | 7.4 | 1.2 | 9.8 | 5 | 0 | 0 | n/a | 10.2 | 21.4 | 0.2 | 6.4 | 0 | 0 | 1 | 5.75 | 0 | 21.25 | 0 | 2.25 | 4.75 | 21.5 | 0 | 9.75 | 15.75 | 8 |  |
| November | 5 | 0.4 | 13 | 0 | 1.4 | 0.8 | 0 | 0 | n/a | 21.2 | 26.2 | 0 | 0 | 19.4 | 9 | 0 | 11 | 2.25 | 5 | 0.25 | 19 | 0.5 | 9.25 | 0 | 5 | 8.5 | 0 |  |
| November | 6 | 3.2 | 0 | 14 | 7.8 | 0 | 4.2 | 0 | n/a | 2.2 | 88 | 2.6 | 66 | 17 | 2.75 | 0 | 0 | 5.75 | 8.25 | 0.5 | 0.5 | 0 | 0 | 0 | 0 | 1.5 | 7.75 |  |
| November | 7 | 3.6 | 0 | 0 | 0 | 0 | 2.8 | 0 | n/a | 0 | 4.2 | 2 | 36.6 | 18.6 | 0.25 | 0 | 3.75 | 3.25 | 10.25 | 4.25 | 2.75 | 0 | 0 | 0 | 0 | 3.25 | 4.25 |  |
| November | 8 | 0 | 0 | 2.8 | 11.8 | 0 | 4 | 0 | n/a | 2.4 | 17.6 | 7.4 | 2.2 | 0.8 | 0 | 0.75 | 0.25 | 11.75 | 0 | 15 | 1.75 | 3 | 0 | 0.75 | 0 | 12.75 | 3.25 |  |
| November | 9 | 0 | 0 | 1.4 | 4 | 0 | 7 | 0 | n/a | 0.8 | , | 2 | 3 | 6.4 | 1 | 2.75 | 0 | 0 | 0 | 0.5 | 8.25 | 1.75 | 0 | 3.25 | 6 | 0.25 | 0 |  |
| November | 10 | 0 | 0.2 | 6.2 | 0 | 0 | 4 | 0 | n/a | 3.4 | 18.2 | 4.2 | 2.2 | 8.2 | 0 | 0 | 0 | 0 | 3.25 | 2.5 | 0 | 0 | 4.25 | 0 | 0 | 14.5 | 0 |  |
| November | 11 | 0 | 0 | 18.8 | 0 | 0 | 7.6 | 0 | n/a | 27.8 | 1.2 | 1.6 | 12 | 0.6 | 3.25 | 0 | 0 | 0 | 0 | 6.25 | 0 | 3.25 | 0.25 | 0 | 0 | 4 | 0.25 |  |
| November | 12 | 0 | 22.2 | 14.6 | 0 | 4.8 | 11.6 | 0 | n/a | 0 | 42.2 | 22.6 | 22 | 0 | 0 | 9.5 | 24.5 | 0 | 0 | 0.5 | 0.75 | 2.5 | 0 | 0 | 25 | 17.25 | 0 |  |
| November | 13 | 0 | 53.6 | 7.8 | 1.4 | 4.6 | 5.6 | 0 | n/a | 14 | 18 | 1.8 | 0 | 6.4 | 6.25 | 8.25 | 4 | 3.75 | 0 | 44.5 | 7 | 14.5 | 0 | 0 | 13 | 2.75 | 0 |  |
| November | 14 | 0 | 22.2 | 4.4 | 0 | 25.8 | 4 | 0 | n/a | 0 | 0.6 | 2 | 0.8 | 0 | 2.25 | 0.25 | 3 | 0.5 | 0 | 23.25 | 0 | 20.5 | 9 | 0 | 2.5 | 34.5 | 0 |  |
| November | 15 | 0 | 33.6 | 2.6 | 0 | 18.4 | 0 | 2.2 | n/a | 0 | 39.8 | 14.2 | 0.2 | 17 | 1.5 | 0 | 0.25 | 0 | 0 | 13 | 5.25 | 10.5 | 18 | 0 | 6.75 | 71.25 | 0.25 |  |
| November | 16 | 0 | 27 | 10.6 | 0 | 6.4 | 11 | 22.6 | n/a | 0.6 | 0 | 4.6 | 0 | 58.4 | 0 | 0 | 0 | 3 | 0 | 3.75 | 6.5 | 2.5 | 0 | 0.75 | 33 | 81.75 | 0.5 |  |
| November | 17 | 2.4 | 0 | 3.8 | 0 | 0 | 0 | 1.2 | n/a | 0 | 12.8 | 5.4 | 0 | 15 | 12.75 | 11.75 | 0.5 | 0 | 0 | 7 | 0 | 0.25 | 0.25 | 0.25 | 11.75 | 0 | 0 |  |
| November | 18 | 0 | 0 | 0 | 0 | 0 | 15.2 | 68 | 14.4 | 0 | 6.8 | 1 | 0 | 16.4 | 2 | 6.5 | 4.5 | 5 | 0 | 49 | 0 | 0.25 | 0 | 9.75 | 5.5 | 0 | 0 |  |
| November | 19 | 1.2 | 3.6 | 1 | 0 | 15.2 | 21.8 | 27.4 | 0 | 0 | 7.8 | 1.6 | 0 | 42.2 | 0.75 | 0 | 7 | 24 | 0 | 0 | 0 | 18.75 | 0 | 0.25 | 3.5 | 14.5 | 0 |  |
| November | 20 | 8.4 | 35.2 | 9.4 | 0 | 6.8 | 1 | 0 | 0 | 0 | 14.4 | 0 | 3.4 | 7.2 | 0 | 0.25 | 3.75 | 0 | 0 | 0 | 1.25 | 7.75 | 0 | 0 | 1.25 | 0.75 | 0.75 |  |
| November | 21 | 0.2 | 28.8 | 4 | 0 | 7.6 | 0 | 0 | 0 | 0 | 35.2 | 0 | 7.2 | 1 | 0 | 0 | 10.75 | 0 | 2.25 | 0 | 0 | 2.75 | 0 | 0 | 0.25 | 0 | 0 |  |
| November | 22 | 1.4 | 19.6 | 3.4 | 0 | 7.2 | 0 | 0 | 1.2 | 0 | 1.8 | 0 | 0 | 17.6 | 0 | 5.75 | 14 | 0 | 17.75 | 0 | 16 | 7.75 | 9 | 0 | 2.5 | 0 | 17.75 |  |
| November | 23 | 0.8 | 3.8 | 4.8 | 6.6 | 0.8 | 0 | 10.6 | 5.6 | 0 | 10.2 | 0 | 0 | 10.2 | 0 | 58 | 0.25 | 0 | 1.5 | 0 | 0.5 | 8.5 | 6.5 | 2.25 | 6 | 2 | 0 |  |
| November | 24 | 3.2 | 28.2 | 18.4 | 0 | 0 | 0 | 5 | 13.4 | 0.8 | 1.4 | 0 | 0 | 2.6 | 0 | 15 | 10.25 | 0 | 18 | 8 | 11 | 2 | 8.75 | 0.5 | 8.5 | 2.25 | 0 |  |
| November | 25 | 2.2 | 43.2 | 7.2 | 9 | 0.6 | 0 | 8 | 2.8 | 20.4 | 14.4 | 0 | 0 | 17.2 | 4.5 | 9.75 | 0.25 | 0 | 9.75 | 0.75 | 8.25 | 5.25 | 0 | 0 | 0.25 | 0 | 5.75 |  |
| November | 26 | 0 | 10.8 | 0 | 25.6 | 2 | 0 | 0 | 1.6 | 0 | 32 | 10.6 | 0 | 9.4 | 14.25 | , | 0 | 0 | 27.75 | 0 | 6.5 | 6.5 | 3 | 0 | 1 | 43.75 | 10 |  |
| November | 27 | 8 | 0.8 | 0.4 | 3.8 | 6.8 | 0 | 0.2 | 0.8 | 0 | 17.8 | 0 | 0 | 0 | 0.25 | 1.25 | 0 | 0 | 5.75 | 0 | 5.75 | 0.75 | 36 | 0 | 5 | 1.25 | 0 |  |
| November | 28 | 16.8 | 0 | 0.4 | 0 | 33 | 0 | 66.8 | 0 | 5.2 | 0.8 | 10.4 | 2.2 | 8.2 | 0 | 24.25 | 0 | 0 | 7.25 | 0 | 0 | 2.25 | 13.5 | 0 | 0 | 34.25 | 0 |  |
| November | 29 | 1.8 | 0.6 | 0 | 2 | 8.8 | 0 | 2.6 | 1.2 | 12 | 17.4 | 0.2 | 6 | 7 | 8 | 0 | 5.5 | 0 | 3.25 | 0 | 1.5 | 0.5 | 2 | 0 | 0 | 9.25 | 15.25 |  |
| November | 30 | 9.4 | 1.2 | 2.8 | 4.4 | 2.4 | 0 | 0 | 5.6 | 0.2 | 2 | 0 | 0.6 | 1 | 17.75 | 7 | 0.5 | 7.75 | 0.25 | 0.25 | 0.25 | 7.5 | 0.25 | 0 | 11.5 | 1 | 6.5 |  |
| Total Monthly Rainfall |  | 73.2 | 375.4 | 140.8 | 87.6 | 157.6 | 99.8 | 215.2 | 98.4 | 141.6 | 485.4 | 98.4 | 175.6 | 308.6 | 89.5 | 169.75 | 124 | 75.5 | 146.25 | 243.75 | 134.25 | 160.25 | 170.75 | 17.75 | 185.25 | 399.5 | 116.75 |  |
| Maximum Daily Rainfall |  | 16.8 | 53.6 | 18.8 | 25.6 | 33 | 21.8 | 68 | 36.4 | 27.8 | 88 | 22.6 | 66 | 58.4 | 17.75 | 58 | 24.5 | 24 | 27.75 | 49 | 19 | 22.25 | 36 | 9.75 | 33 | 81.75 | 22.75 |  |


| Date |  | Daily Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| December | 1 | 0 | 5.2 | 9.4 | 0 | 8.2 | 0 | 0 | 0 | 2.6 | 0 | 15.2 | 1.8 | 0 | 0.25 | 0.25 | 16 | 6.75 | 0.25 | 0.25 | 1 | 17.25 | 2.5 | 0 | 0 | 8.5 | 0 |  |
| December | 2 | 0 | 19.4 | 16.4 | 12.4 | 12.8 | 0 | 1.4 | 0.2 | 0 | 0 | 46.4 | 0 | 0 | 0.25 | 0 | 23.25 | 5.75 | 0 | 9.5 | 0 | 2.5 | 1 | 0 | 0 | 0 | 3.5 |  |
| December | 3 | 0 | 3.6 | 0 | 0 | 0 | 0 | 9 | 0.2 | 0 | 0 | 100.4 | 0 | 0 | 0.25 | 0 | 10.25 | 0 | 0.25 | 2.25 | 4.5 | 0 | 0 | 0 | 0 | 0 | 10 |  |
| December | 4 | 0 | 0 | 0 | 0 | 6 | 3.2 | 0 | 16.6 | 0 | 3.6 | 18 | 0 | 0 | 0 | 0 | 13 | 0 | 0 | 7.75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| December | 5 | 0 | 2.4 | 4.8 | 0 | 4.6 | 0 | 6.6 | 1.4 | 0 | 0.4 | 0.2 | 0 | 0 | 0 | 0 | 10.75 | 0 | 2 | 2.25 | 5 | 0 | 0 | 0 | 0.75 | 6.5 | 0.25 |  |
| December | 6 | 0 | 0.4 | 16.2 | 0 | 5.2 | 0 | 9.8 | 12.4 | 0 | 0 | 0 | 4.2 | 0 | 0 | 0 | 8.5 | 0 | 1 | 17.5 | 0.25 | 0 | 0 | 0.25 | 0 | 0.25 | 5.25 |  |
| December | 7 | 0.4 | 3 | 0 | 0 | 0 | 0 | 10.8 | 6 | 0 | 0.2 | 0 | 0.4 | 0 | 7 | 0 | 11 | 0 | 5.25 | 9.75 | 0 | 0 | 0 | 0.75 | 10.5 | 17.75 | 0.25 |  |
| December | 8 | 1 | 2.8 | 1 | 0 | 12.4 | 0 | 0 | 16.2 | 0 | 0.2 | 0 | 0 | 0 | 4.25 | 1.75 | 7 | 0 | 0 | 5.75 | 0 | 0 | 0 | 0 | 16 | 2.75 | 7.75 |  |
| December | 9 | 2.6 | 2.8 | 2.6 | 1.6 | 0 | 0 | 0 | 22 | 0 | 0.6 | 3.4 | 9.6 | 0 | 18.25 | 0 | 0 | 0 | 19.75 | 43 | 0 | 0.25 | 0 | 0 | 0 | 0.5 | 2 |  |
| December | 10 | 1.8 | 8.4 | 3.2 | 0.8 | 5.4 | 4.4 | 9.2 | 38.4 | 0 | 1.4 | 0.2 | 3.8 | 0 | 0 | 0 | 9.25 | 0 | 13.75 | 15.25 | 8.25 | 0 | 8 | 0 | 0 | 0.5 | 3 |  |
| December | 11 | 0 | 2.8 | 2.2 | 0 | 2.8 | 6.2 | 2 | 0 | 0 | 19.2 | 0.4 | 0 | 0 | 3.75 | 0 | 0.25 | 0 | 38.25 | 8 | 1.5 | 0 | 4.25 | 0 | 0.5 | 25 | 0 |  |
| December | 12 | 0 | 11.4 | 28.8 | 0 | 8.4 | 20 | 4.4 | 0 | 0.2 | 17 | 0 | 22.2 | 0 | 5 | 0 | 6 | 0 | 6 | 0 | 0.25 | 0 | 36.75 | 4.75 | 0.25 | 27 | 0 |  |
| December | 13 | 0 | 33.2 | 7.2 | 0 | 32.4 | 0.4 | 3.4 | 3 | 0 | 29.4 | 2.2 | 4.2 | 0 | 0.5 | 0 | 0 | 2 | 1.25 | 19 | 0 | 0 | 22.25 | 0.5 | 11.25 | 12 | 0 |  |
| December | 14 | 1.6 | 0 | 22.6 | 1.2 | 15 | 21.4 | 18.6 | 14.4 | 0 | 31.8 | 9 | 1.6 | 10.4 | 3.25 | 0 | 7.75 | 2 | 0 | 10.25 | 0 | 0 | 24 | 0 | 2 | 4.25 | 0 |  |
| December | 15 | 2.8 | 10 | 63 | 4.2 | 3.6 | 5 | 0 | 0.8 | 0 | 6.4 | 16.2 | 0 | 25.75 | 2 | 2.25 | 1.5 | 2.5 | 0 | 0 | 0 | 3 | 2.25 | 0 | 8.25 | 0.5 | 0 |  |
| December | 16 | 47.6 | 7 | 0 | 49.2 | 61.8 | 5.8 | 8.4 | 0.6 | 0 | 0 | 0 | 0 | 14.25 | 0 | 1.25 | 0.75 | 4.5 | 0 | 0.25 | 0 | 4.5 | 0.75 | 0 | 8.5 | 0 | 0 |  |
| December | 17 | 24.6 | 1 | 9.8 | 5.6 | 29.8 | 0.2 | 1.6 | 3.8 | 0 | 0 | 1.8 | 5.8 | 9.5 | 0 | 0.75 | 24.75 | 0 | 0.25 | 0 | 0 | 11 | 7 | 0.25 | 6.5 | 1 | 1.75 |  |
| December | 18 | 2.2 | 0 | 3.8 | 0 | 1.6 | 0.2 | 0 | 1.8 | 0 | 1.8 | 6.2 | 3.4 | 4 | 0.75 | 0.75 | 10.75 | 0 | 2 | 25.75 | 1 | 6.5 | 13.5 | 0 | 16.5 | 9.25 | 0 |  |
| December | 19 | 9.8 | 0 | 2.4 | 1.2 | 6 | 0.2 | 0 | 0 | 2 | 10 | 15.2 | 0 | 4.75 | 0.75 | 0.5 | 1.75 | 0 | 3.25 | 15.25 | 4 | 3 | 20 | 0 | 20.75 | 45.25 | 0 |  |
| December | 20 | 7.8 | 0 | 1 | 3.8 | 0 | 0 | 0.6 | 0 | 15.2 | 6.2 | 0 | 0 | 11 | 1.5 | 0 | 22.75 | 0 | 3.75 | 0 | 1.25 | 0.75 | 5.5 | 0.5 | 0 | 0 | 0.25 |  |
| December | 21 | 0.6 | 6.6 | 0 | 1.2 | 0 | 0 | 0.4 | 0 | 6.4 | 15.8 | 0 | 27.4 | 4.75 | 0 | 0 | 7.25 | 10.75 | 23.5 | 9 | 0 | 2.25 | 11.25 | 0.25 | 35.5 | 0 | 0 |  |
| December | 22 | 0 | 0 | 0 | 4.8 | 0 | 0.4 | 0 | 0 | 10.6 | 1 | 19.4 | 13.2 | 0 | 0 | 0 | 0 | 2.5 | 3 | 9.25 | 0.25 | 0.75 | 0 | 0 | 12.5 | 5.25 | 0 |  |
| December | 23 | 7 | 0.6 | 0 | 3.2 | 0 | 0 | 0 | 0 | 11.8 | 17.2 | 12.6 | 0.4 | 0 | 7 | 0.25 | 2.5 | 7.25 | 0 | 2.5 | 1.5 | 0 | 10.75 | 0 | 0.25 | 21.75 | 8.25 |  |
| December | 24 | 0.2 | 18.2 | 0 | 4.6 | 0 | 1.2 | 4.8 | 0.8 | 15.4 | 23 | 0.2 | 21.8 | 0 | 5.25 | 1.25 | 4.25 | 6.75 | 11.75 | 8.25 | 0.25 | 0 | 12.25 | 0 | 0 | 1.25 | 44.75 |  |
| December | 25 | 0 | 17.6 | 0 | 7.6 | 0 | 14.6 | 5.2 | 9.4 | 12.8 | 3.4 | 0 | 1.8 | 0 | 1 | 5.75 | 2 | 0 | 0.75 | 0 | 0 | 2.75 | 0 | 0 | 7 | 4.5 | 9.25 |  |
| December | 26 | 7.4 | 21.8 | 0 | 2.4 | 0 | 12.2 | 0 | 4.4 | 0.4 | 4.8 | 0 | 7.8 | 0 | 8 | 3.5 | 22.25 | 0 | 0 | 0 | 0 | 0 | 0.5 | 0 | 4.75 | 3.25 | 26.75 |  |
| December | 27 | 4.6 | 17 | 0 | 4.8 | 0 | 22.4 | 3.6 | 0 | 6.2 | 0 | 4.2 | 5 | 0 | 10.75 | 5.25 | 10 | 0 | 1.5 | 1 | 1 | 0.25 | 5.5 | 0 | 6.5 | 0 | 10.5 |  |
| December | 28 | 10.6 | 15.4 | 0 | 0 | 2 | 1 | 1.8 | 0 | 13.8 | 0 | 7.6 | 0.6 | 0.25 | 0.5 | 30.75 | 0.75 | 2.5 | 6.25 | 12.25 | 0.25 | 19 | 1 | 0 | 0 | 0 | 0.75 |  |
| December | 29 | 0.2 | 46.4 | 0.2 | 0 | 0 | 0 | 0 | 5.6 | 0.2 | 0.6 | 4.2 | 12.6 | 0 | 0 | 19.75 | , | 0 | 0.25 | 0 | 2.75 | 2.75 | 24.25 | 0 | 12.5 | 0 | 5 |  |
| December | 30 | 0 | 0 | 0.2 | 3 | 0 | 4.2 | 0.2 | 0.6 | 5.6 | 0 | 1.8 | 1.6 | 3.5 | 0 | 5.25 | 3 | 0 | 0 | 0 | 0 | 4 | 19.5 | 0 | 29.25 | 0 | 13 |  |
| December | 31 | 0 | 0.8 | 11.6 | 3.6 | 6.4 | 0 | 4.6 | 3.2 | 16.2 | 0 | 0 | 3 | 16.25 | 0 | 3.75 | 0 | 0.25 | 0 | 0 | 0.25 | 0 | 0 | 0 | 9.75 | 0 | 2.75 |  |
| Total Monthly Rainfall Maximum Daily Rainfall |  | 132.8 | 257.8 | 206.4 | 115.2 | 224.4 | 123 | 106.4 | 161.8 | 119.4 | 194 | 284.8 | 152.2 | 104.4 | 80.25 | 83 | 240.25 | 53.5 | 144 | 234 | 33.25 | 80.5 | 232.75 | 7.25 | 219.75 | 197 | 155 |  |
|  |  | 47.6 | 46.4 | 63 | 49.2 | 61.8 | 22.4 | 18.6 | 38.4 | 16.2 | 31.8 | 100.4 | 27.4 | 25.75 | 18.25 | 30.75 | 24.75 | 10.75 | 38.25 | 43 | 8.25 | 19 | 36.75 | 4.75 | 35.5 | 45.25 | 44.75 |  |

## C2. Monthly Rainfall Data Hartland Landfill Weather Station

| Month | Monthly Rainfall (mm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 | Avg |
| January | 205.8 | 119.8 | 210.2 | 277.4 | 312.4 | 106.8 | 121.6 | 180.5 | 161.8 | 150.75 | 53.75 | 70.25 | 160.8 | 114.5 | 26 | 153.5 | 158.8 | 166.5 | 194.5 | 149.5 | 45.3 | 150.3 |
| February | 29 | 53 | 39.2 | 79.2 | 94 | 63 | 43.8 | 50.25 | 72.5 | 122.5 | 67.3 | 130 | 133 | 172.8 | 50 | 70 | 39.3 | 129.0 | 104.8 | 56.0 | 99.0 | 86.7 |
| March | 149.6 | 97 | 86 | 36.6 | 162.2 | 71.4 | 99.2 | 92 | 108 | 129.5 | 91 | 113 | 123 | 137 | 158 | 132 | 15.0 | 40.0 | 41.3 | 117.0 | 70.0 | 94.7 |
| April | 46.2 | 4.8 | 43.6 | 55.6 | 40.8 | 70.4 | 94.6 | 67.75 | 109.75 | 53.5 | 70.5 | 66 | 35 | 14 | 90.75 | 108 | 54.3 | 32.8 | 37.3 | 118.0 |  | 54.8 |
| May | 25.4 | 38.4 | 9.2 | 34.2 | 18.4 | 15.8 | 62.6 | 76.5 | 61.75 | 32 | 53.5 | 40.25 | 4 | 11.8 | 41.25 | 10.5 | 31.3 | 47.5 | 22.0 | 50.3 |  | 35.4 |
| June | 2 | 20.2 | 1.6 | 30.4 | 34.4 | 24.2 | 7.2 | 1.8 | 7.5 | 44.75 | 30.25 | 12 | 3.5 | 41.5 | 34 | 19.3 | 19.0 | 45.3 | 26.3 | 46.5 |  | 25.4 |
| July | 11.8 | 31 | 1.2 | 2 | 41.7 | 7 | 10.6 | 2.65 | 26.25 | 21 | 0 | 6.5 | 17.5 | 18.5 | 0 | 3.3 | 22.3 | 10.3 | 0.0 | 21.5 |  | 14.2 |
| August | 0.6 | 106.2 | 2.8 | 3.4 | 23.2 | 43.6 | 22.2 | 26.25 | 16.5 | 2.25 | 3 | 12.75 | 15.5 | 2.8 | 3.25 | 0.5 | 12.8 | 18.8 | 7.8 | 1.9 |  | 16.8 |
| September | 34.6 | 83.2 | 0.4 | 31.8 | 27.2 | 21.4 | 54.8 | 133.75 | 43.75 | 0.5 | 60.25 | 51.25 | 76.5 | 42.8 | 19.25 | 40.5 | 80.3 | 94.3 | 123.8 | 1.3 |  | 45.4 |
| October | 384.4 | 105.2 | 48.2 | 48 | 100.4 | 56 | 170.8 | 86.25 | 80.25 | 131.25 | 30.25 | 137.25 | 81 | 135 | 45.75 | 89.5 | 93.8 | 139.0 | 143.5 | 112.0 |  | 108.3 |
| November | 215.2 | 98.4 | 141.6 | 485 | 98.4 | 175.6 | 308.6 | 89.5 | 169.75 | 124 | 75.5 | 146.25 | 243.8 | 134.3 | 160.25 | 170.8 | 17.8 | 185.3 | 399.5 | 116.8 |  | 172.7 |
| December | 106.4 | 161.8 | 119.4 | 194 | 284.8 | 152.2 | 104.4 | 80.25 | 83 | 240.25 | 53.5 | 144 | 234 | 33.3 | 80.5 | 232.8 | 7.3 | 219.8 | 197.0 | 155.0 |  | 151.7 |
| Total Yearly Rainfall | 1211 | 919 | 703 | 1278 | 1238 | 807 | 1100 | 887 | 941 | 1052 | 589 | 929 | 1127 | 858 | 709 | 1031 | 552 | 1128 | 1298 | 946 |  | 956 |

No weater was colletert Nov 4-17, 200
Weather data collected in 2009 is unreliable due to equipment failure

## Appendix D

Leachate Pipeline Flow Data

## Appendix D. Leachate Pipeline Flow Data

| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in $M^{\wedge} 3$ | Minimum Daily <br> Flow in $M^{\wedge 3}$ | Average Daily <br> Flow in $M^{\wedge 3}$ | Monthly Total <br> Flow in $M^{\wedge 3}$ |
| Jan-1997 | 13 | 2148 | 0 | 2148 | 40816 |
| Feb-1997 | 0 | 2700 | 732 | 2226 | 62334 |
| Mar-1997 | 0 | 2759 | 2412 | 2696 | 83557 |
| Apr-1997 | 0 | 2655 | 214 | 2545 | 78899 |
| May-1997 | 4 | 2654 | 0 | 956 | 29561 |
| Jun-1997 | 11 | 1563 | 0 | 546 | 17503 |
| Jul-1997 | 17 | 2452 | 0 | 525 | 16090 |
| Aug-1997 | 15 | 1608 | 0 | 386 | 11606 |
| Sep-1997 | 21 | 2662 | 0 | 456 | 14159 |
| Oct-1997 | 6 | 2782 | 0 | 1233 | 36996 |
| Nov-1997 | 4 | 2533 | 0 | 1437 | 44565 |
| Dec-1997 | 2 | 2408 | 0 | 1695 | 52547 |
| TOTAL | 93 | 2782 | 0 | 1404 | 488633 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Days Without Flow | Maximum Daily Flow in $M^{\wedge} 3$ | Minimum Daily Flow in $M^{\wedge} 3$ | Average Daily Flow in $M^{\wedge} 3$ | Monthly Total Flow in M^3 |
| Jan-1998 | 0 | 2603 | 879 | 2187 | 65977 |
| Feb-1998 | 3 | 2603 | 0 | 1624 | 43531 |
| Mar-1998 | 3 | 2419 | 0 | n/a | 35275 |
| Apr-1998 | 10 | 1542 | 0 | 542 | 16675 |
| May-1998 | 13 | 1890 | 0 | 423 | 12574 |
| Jun-1998 | 17 | 1241 | 0 | 239 | 7239 |
| Jul-1998 | 21 | 1919 | 0 | 240 | 7645 |
| Aug-1998 | 31 | 0 | 0 | 0 | 0 |
| Sep-1998 | 27 | 1020 | 0 | 58 | 1821 |
| Oct-1998 | 19 | 2379 | 0 | 511 | 14962 |
| Nov-1998 | 7 | 2462 | 0 | 1538 | 45248 |
| Dec-1998 | 0 | 2619 | n/a | n/a | 76518 |
| TOTAL | 151 | 2619 | 0 | 736 | 327465 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Days Without Flow | Maximum Daily Flow in $M^{\wedge} 3$ | Minimum Daily Flow in $M^{\wedge} 3$ | Average Daily Flow in $M^{\wedge} 3$ | Monthly Total Flow in $M^{\wedge} 3$ |
| Jan-1999 | 0 | n/a | n/a | n/a | 69000 |
| Feb-1999 | 0 | 2970 | 1966 | 2681 | 74930 |
| Mar-1999 | 0 | 2970 | 2539 | 2875 | 92028 |
| Apr-1999 | 5 | 2894 | 0 | 1812 | 57986 |
| May-1999 | 11 | 2889 | 0 | 660 | 20656 |
| Jun-1999 | 18 | 1215 | 0 | 257 | 7480 |
| Jul-1999 | 18 | 2772 | 0 | 365 | 11687 |
| Aug-1999 | 30 | n/a | n/a | n/a | 1000 |
| Sep-1999 | 4 | 1193 | 0 | 148 | 4597 |
| Oct-1999 | 19 | 1473 | 0 | 163 | 4719 |
| Nov-1999 | 4 | 2840 | 0 | 1497 | 43972 |
| Dec-1999 | 4 | 3480 |  | n/a | 67907 |
| TOTAL | 113 | 3480 | 0 | 1162 | 455962 |

## Appendix D. Leachate Pipeline Flow Data

| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Days Without Flow | Maximum Daily Flow in $M^{\wedge} 3$ | Minimum Daily Flow in $M^{\wedge} 3$ | Average Daily Flow in M^3 | Monthly Total Flow in M^3 |
| Jan-2000 | 3 | 2791 | 0 | n/a | 55088 |
| Feb-2000 | 7 | 2235 | 0 | n/a | 29318 |
| Mar-2000 | 7 | 2211 | 0 | n/a | 37258 |
| Apr-2000 | 8 | 1737 | 0 | n/a | 21289 |
| May-2000 | 19 | 2772 | 0 | 501 | 15547 |
| Jun-2000 | 19 | 2685 | 0 | 396 | 11902 |
| Jul-2000 | 22 | 1086 | 0 | 155 | 4831 |
| Aug-2000 | 26 | 907 | 0 | 116 | 3619 |
| Sep-2000 | 21 | 2640 | 0 | 354 | 10623 |
| Oct-2000 | n/a | n/a | n/a | n/a | 8572 |
| Nov-2000 | n/a | n/a | n/a | n/a | 11737 |
| Dec-2000 | 12 | 2926 | 0 | 943 | 29235 |
| TOTAL | 144 | 2926 | 0 | 411 | 239019 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 | Minimum Daily <br> Flow in M^3 | Average Daily <br> Flow in M^3 | Monthly Total <br> Flow in M^3 $^{\prime}$ |
| Jan-2001 |  |  |  |  | 38089 |
| Feb-2001 |  |  |  |  | 30154 |
| Mar-2001 |  |  |  |  | 24565 |
| Apr-2001 | 15 | 2622 | 0 | 646 | 18746 |
| May-2001 | 16 | 1647 | 2026 | 0 | 400 |
| Jun-2001 | 22 | 1521 | 2873 | 0 | 295 |
| Jul-2001 | 22 | 1633 | 0 | 244 | 12407 |
| Aug-2001 | 24 | 2631 | 0 | 291 | 8862 |
| Sep-2001 | 23 | 2899 | 0 | 157 | 7576 |
| Oct-2001 | 15 | 3397 | 3397 | 58 | 5047 |
| Nov-2001 | 7 |  | 0 | 993 | 4723 |
| Dec-2001 | 0 | 144 |  |  | 2241 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in $\mathbf{M}^{\wedge} \mathbf{3}$ | Minimum Daily <br> Flow in $\mathbf{M}^{\wedge}$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge} 3$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge 3}$ |
| Jan-2002 | 1 | 2892 | 0 | 1946 | 60342 |
| Feb-2002 | 1 | 3395 | 0 | 1946 | 54497 |
| Mar-2002 | 2 | 3412 | 0 | 2200 | 68224 |
| Apr-2002 | 6 | 2444 | 0 | 1022 | 29651 |
| May-2002 | 13 | 2118 | 0 | 578 | 17928 |
| Jun-2002 | 16 | 1683 | 0 | 381 | 11435 |
| Jul-2002 | 20 | 2330 | 0 | 330 | 10231 |
| Aug-2002 | 23 | 1521 | 0 | 253 | 7848 |
| Sep-2002 | 28 | 1154 | 0 | 72 | 2160 |
| Oct-2002 | 26 | 1356 | 0 | 143 | 4450 |
| Nov-2002 | 14 | 2477 | 0 | 618 | 18555 |
| Dec-2002 | 9 | 2896 | 3412 | 0 | 987 |
| TOTAL | 159 |  |  | 873 | 30615 |

## Appendix D. Leachate Pipeline Flow Data

| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Days Without Flow | Maximum Daily Flow in $M^{\wedge} 3$ | Minimum Daily Flow in $M^{\wedge} 3$ | Average Daily Flow in $M^{\wedge} 3$ | Monthly Total Flow in $M^{\wedge} 3$ |
| Jan-2003 | 5 | 4880 | 0 | 1569 | 48664 |
| Feb-2003 | 4 | 2819 | 0 | 1354 | 37921 |
| Mar-2003 | 3 | 2849 | 0 | 1533 | 47546 |
| Apr-2003 | 6 | 2185 | 0 | 846 | 24562 |
| May-2003 | 18 | 1892 | 0 | 514 | 15962 |
| Jun-2003 | 15 | 1397 | 0 | 340 | 10224 |
| Jul-2003 | 20 | 1126 | 0 | 282 | 8768 |
| Aug-2003 | 21 | 1140 | 0 | 214 | 6658 |
| Sep-2003 | 23 | 1147 | 0 | 222 | 6671 |
| Oct-2003 | 15 | 3869 | 0 | 1621 | 50260 |
| Nov-2003 | 4 | 3681 | 0 | 2546 | 76397 |
| Dec-2003 | 0 | 3587 | 96 | 2218 | 68773 |
| TOTAL | 134 | 4880 | 0 | 1105 | 402406 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 | Minimum Daily <br> Flow in M^3 | Average Daily <br> Flow in M^3 $^{\prime}$ | Monthly Total <br> Flow in M^3 $^{\prime}$ |
| Jan-2004 | 0 | 2859 | 99 | 1597 | 49509 |
| Feb-2004 | 0 | 2833 | 2 | 1536 | 44551 |
| Mar-2004 | 2 | 2792 | 0 | 1170 | 36288 |
| Apr-2004 | 22 | 1675 | 0 | 319 | 9272 |
| May-2004 | 31 | 0 | 0 | 0 | 0 |
| Jun-2004 | 13 | 3565 | 0 | 1104 | 33135 |
| Jul-2004 | 30 | 63 | 0 | 2 | 63 |
| Aug-2004 | 23 | 3007 | 2998 | 0 | 601 |
| Sep-2004 | 24 | 3668 | 0 | 454 | 18636 |
| Oct-2004 | 14 | 2889 | 0 | 1138 | 13640 |
| Nov-2004 | 2 | 3659 | 0 | 1631 | 35284 |
| Dec-2004 | 4 | 3668 | 0 | 2191 | 68938 |
| TOTAL | 165 |  | 979 | 357263 |  |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Days Without Flow | Maximum Daily Flow in $M^{\wedge} 3$ | Minimum Daily Flow in $M^{\wedge} 3$ | Average Daily Flow in $M^{\wedge} 3$ | Monthly Total Flow in M^3 |
| Jan-2005 | 7 | 3671 | 0 | 2247 | 69680 |
| Feb-2005 | 2 | 3662 | 0 | 2571 | 72002 |
| Mar-2005 | 11 | 2484 | 0 | 824 | 25551 |
| Apr-2005 | 4 | 2800 | 0 | 1041 | 30206 |
| May-2005 | 18 | 2840 | 0 | 691 | 21437 |
| Jun-2005 | 16 | 1800 | 0 | 431 | 12946 |
| Jul-2005 | 18 | 1301 | 0 | 274 | 8512 |
| Aug-2005 | 22 | 2771 | 0 | 417 | 12927 |
| Sep-2005 | 19 | 1301 | 0 | 274 | 8512 |
| Oct-2005 | 24 | 3025 | 0 | 432 | 13393 |
| Nov-2005 | 5 | 2999 | 0 | 1347 | 40421 |
| Dec-2005 | 5 | 2916 | 0 | 1180 | 36601 |
| TOTAL | 151 | 3671 | 0 | 977 | 352188 |

## Appendix D. Leachate Pipeline Flow Data

| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in $\mathbf{M}^{\wedge}$ | Minimum Daily <br> Flow in $\mathbf{M}^{\wedge}$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge}$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge} 3$ |
| Jan-2006 | 0 | 3415 | 1123 | 3075 | 95329 |
| Feb-2006 | 2 | 3435 | 0 | 2206 | 61795 |
| Mar-2006 | 8 | 2926 | 0 | 846 | 26247 |
| Apr-2006 | 14 | 2141 | 0 | 558 | 16762 |
| May-2006 | 19 | 2058 | 0 | 464 | 14402 |
| Jun-2006 | 21 | 2175 | 0 | 359 | 10795 |
| Jul-2006 | 28 | 1996 | 0 | 105 | 3277 |
| Aug-2006 | 23 | 1991 | 0 | 327 | 10148 |
| Sep-2006 | 22 | 2043 | 0 | 255 | 7654 |
| Oct-2006 | 23 | 3677 | 0 | 304 | 9427 |
| Nov-2006 | 2 | 3563 | 3130 | 2872 | 86167 |
| Dec-2006 | 0 | 3677 | 0 | 3425 | 106180 |
| TOTAL | 162 |  |  | 1233 | 448183 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 | Minimum Daily <br> Flow in M^3 | Average Daily <br> Flow in M^3 $^{\prime}$ | Monthly Total <br> Flow in M^3 $^{\prime}$ |
| Jan-2007 | 0 | 3657 | 3272 | 3464 | 107411 |
| Feb-2007 | 0 | 3619 | 815 | 2398 | 67168 |
| Mar-2007 | 0 | 3442 | 38 | 2130 | 66034 |
| Apr-2007 | 6 | 1929 | 0 | 872 | 26177 |
| May-2007 | 14 | 2120 | 2127 | 0 | 572 |
| Jun-2007 | 18 | 2157 | 0 | 477 | 17740 |
| Jul-2007 | 22 | 1730 | 0 | 387 | 14331 |
| Aug-2007 | 28 | 76 | 0 | 92 | 12019 |
| Sep-2007 | 29 | 3288 | 0 | 2 | 2859 |
| Oct-2007 | 14 | 2306 | 0 | 945 | 76 |
| Nov-2007 | 11 | 3715 | 0 | 793 | 29316 |
| Dec-2007 | 1 | 143 | 3715 | 0 | 2549 |
| TOTAL |  |  | 1223 | 79041 |  |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in $M^{\wedge} 3$ | Minimum Daily <br> Flow in M^3 | Average Daily <br> Flow in $M^{\wedge 3}$ | Monthly Total <br> Flow in $M^{\wedge 3}$ |
| Jan-2008 | 0 | 2949 | 260 | 1644 | 50970 |
| Feb-2008 | 2 | 2927 | 0 | 1322 | 38343 |
| Mar-2008 | 0 | 1874 | 0 | 901 | 27953 |
| Apr-2008 | 7 | 1849 | 0 | 845 | 25364 |
| May-2008 | 18 | 2946 | 0 | 537 | 16670 |
| Jun-2008 | 25 | 2695 | 0 | 286 | 8594 |
| Jul-2008 |  |  |  |  | 0 |
| Aug-2008 |  |  |  |  | 0 |
| Sep-2008 |  |  |  |  | 28892 |
| Oct-2008 |  |  |  |  | 29892 |
| Nov-2008 |  |  |  |  | 17433 |
| Dec-2008 |  |  |  |  | 37620 |
| TOTAL | 52 |  |  |  |  |

## Appendix D. Leachate Pipeline Flow Data

| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 $^{2}$ | Minimum Daily <br> Flow in $M^{\wedge 3}$ | Average Daily <br> Flow in $M^{\wedge 3}$ | Monthly Total <br> Flow in $M^{\wedge 3}$ |
| Jan-2009 | 0 | 3522 | 799 | 2400.13 | 74404 |
| Feb-2009 | 2 | 1551 | 0 | 866.54 | 24263 |
| Mar-2009 | 0 | 2388 | 42 | 1040.23 | 32247 |
| Apr-2009 | 5 | 2838 | 0 | 1056.40 | 31692 |
| May-2009 | 11 | 2292 | 0 | 760.94 | 23589 |
| Jun-2009 | 18 | 1839 | 2880 | 0 | 426.58 |
| Jul-2009 | 9 | 2871 | 2847 | 402 | 804.66 |
| Aug-2009 | 0 | 2838 | 0 | 942.19 | 25224 |
| Sep-2009 | 18 | 3345 | 210 | 313.20 | 29208 |
| Oct-2009 | 7 | 3246 | 1125 | 814.52 | 9396 |
| Nov-2009 | 0 | 3522 | 0 | 2426.10 | 25250 |
| Dec-2009 | 0 |  |  | 2169.39 | 672783 |
| TOTAL | 70 |  |  | 1168 | 429056 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 | Minimum Daily <br> Flow in M^3 | Average Daily <br> Flow in M^3 $^{\prime}$ | Monthly Total <br> Flow in M^3 $^{\prime}$ |
| Jan-2010 | 0 | 3438 | 1080 | 2460.58 | 76278 |
| Feb-2010 | 0 | 1824 | 387 | 1013.46 | 28377 |
| Mar-2010 | 10 | 1617 | 0 | 629.32 | 19509 |
| Apr-2010 | 10 | 2853 | 0 | 1280.63 | 38419 |
| May-2010 | 11 | 1989 | 1776 | 0 | 704.32 |
| Jun-2010 | 16 | 1815 | 0 | 498.70 | 21834 |
| Jul-2010 | 17 | 2481 | 0 | 399.87 | 14961 |
| Aug-2010 | 18 | 1752 | 0 | 391.84 | 12396 |
| Sep-2010 | 13 | 1719 | 0 | 647.50 | 12147 |
| Oct-2010 | 7 | 2034 | 0 | 791.61 | 24925 |
| Nov-2010 | 1 | 3462 | 0 | 1107.90 | 33237 |
| Dec-2010 | 1 | 3462 | 0 | 2673.97 | 82893 |
| TOTAL | 104 |  |  | 1050 | 384016 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in $\mathbf{M}^{\wedge} \mathbf{3}$ | Minimum Daily <br> Flow in $\mathbf{M}^{\wedge}$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge} 3$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge 3}$ |
| Jan-2011 | 1 | 3324 | 0 | 2521.00 | 75630 |
| Feb-2011 | 5 | 2823 | 0 | 1544.10 | 46323 |
| Mar-2011 | 2 | 2832 | 0 | 1762.00 | 52860 |
| Apr-2011 | 5 | 2715 | 0 | 1429.70 | 42891 |
| May-2011 | 7 | 2847 | 0 | 894.68 | 27735 |
| Jun-2011 | 13 | 1563 | 0 | 534.80 | 16044 |
| Jul-2011 | 15 | 1428 | 0 | 441.19 | 13677 |
| Aug-2011 | 13 | 1434 | 0 | 338.06 | 10818 |
| Sep-2011 | 8 | 2604 | 0 | 454.94 | 14103 |
| Oct-2011 | 20 | 1932 | 0 | 359.61 | 11148 |
| Nov-2011 | 8 | 3288 | 0 | 1020.30 | 30609 |
| Dec-2011 | 6 | 2589 | 0 | 1011.77 | 31365 |
| TOTAL | 103 | 3324 | 0 | 1026 | 373203 |

## Appendix D. Leachate Pipeline Flow Data

| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 $^{2}$ | Minimum Daily <br> Flow in $M^{\wedge 3}$ | Average Daily <br> Flow in $M^{\wedge 3}$ | Monthly Total <br> Flow in $M^{\wedge 3}$ |
| Jan-2012 | 0 | 2556 | 156 | 1693.94 | 52512 |
| Feb-2012 | 1 | 2742 | 0 | 1634.28 | 47394 |
| Mar-2012 | 1 | 2472 | 0 | 1518.68 | 47079 |
| Apr-2012 | 3 | 2121 | 0 | 914.60 | 27438 |
| May-2012 | 11 | 1671 | 1848 | 0 | 563.16 |
| Jun-2012 | 16 | 2979 | 0 | 449.77 | 17458 |
| Jul-2012 | 10 | 1266 | 0 | 606.97 | 13493 |
| Aug-2012 | 20 | 1410 | 0 | 205.55 | 18816 |
| Sep-2012 | 9 | 2688 | 0 | 460.90 | 6372 |
| Oct-2012 | 15 | 2847 | 0 | 641.03 | 13827 |
| Nov-2012 | 3 | 2610 | 779 | 1601.70 | 49872 |
| Dec-2012 | 0 | 2979 | 0 | 2346.68 | 72747 |
| TOTAL | 89 |  |  | 1053 | 385059 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 | Minimum Daily <br> Flow in M^3 | Average Daily <br> Flow in M^3 $^{\prime}$ | Monthly Total <br> Flow in M^3 $^{\prime}$ |
| Jan-2013 | 0 | 2526 | 43 | 1615.94 | 50094 |
| Feb-2013 | 2 | 1986 | 0 | 1017.00 | 28476 |
| Mar-2013 | 3 | 2694 | 0 | 1750.26 | 54258 |
| Apr-2013 | 5 | 2136 | 0 | 997.30 | 29919 |
| May-2013 | 10 | 1840 | 2515 | 0 | 627.00 |
| Jun-2013 | 12 | 1390 | 0 | 594.30 | 19437 |
| Jul-2013 | 14 | 1590 | 0 | 379.26 | 17829 |
| Aug-2013 | 16 | 2199 | 0 | 355.55 | 11757 |
| Sep-2013 | 9 | 2448 | 0 | 523.50 | 11022 |
| Oct-2013 | 6 | 2418 | 0 | 772.81 | 15705 |
| Nov-2013 | 6 | 1848 | 2694 | 0 | 844.87 |
| Dec-2013 | 7 |  |  | 0 | 23957 |
| TOTAL | 90 |  |  | 708.74 | 21974 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 $^{\prime}$ | Minimum Daily <br> Flow in $M^{\wedge 3}$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge 3}$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge 3}$ |
| Jan-2014 | 1 | 2316 | 0 | 1348.52 | 41804 |
| Feb-2014 | 2 | 3366 | 0 | 1951.21 | 54634 |
| Mar-2014 | 0 | 3387 | 150 | 2252.39 | 69824 |
| Apr-2014 | 2 | 2535 | 0 | 1004.27 | 30128 |
| May-2014 | 7 | 2562 | 0 | 796.90 | 24704 |
| Jun-2014 | 5 | 1677 | 0 | 473.77 | 14213 |
| Jul-2014 | 18 | 2517 | 0 | 485.68 | 15056 |
| Aug-2014 | 14 | 911 | 0 | 302.55 | 9379 |
| Sep-2014 | 16 | 959 | 0 | 305.43 | 9163 |
| Oct-2014 | 16 | 2163 | 0 | 447.48 | 13872 |
| Nov-2014 | 8 | 2808 | 0 | 1420.00 | 42601 |
| Dec-2014 | 0 | 2589 | 600 | 1814.42 | 56247 |
| TOTAL | 89 | 3387 | 0 | 1050 | 381625 |

## Appendix D. Leachate Pipeline Flow Data

| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 $^{\prime}$ | Minimum Daily <br> Flow in $M^{\wedge 3}$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge 3}$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge 3}$ |
| Jan-2015 | 0 | 2472 | 460 | 1976.13 | 61260 |
| Feb-2015 | 0 | 2508 | 312 | 1750.36 | 49008 |
| Mar-2015 | 0 | 2481 | 45 | 1493.71 | 46305 |
| Apr-2015 | 1 | 2763 | 0 | 1118.30 | 33549 |
| May-2015 | 15 | 2078 | 0 | 519.23 | 16096 |
| Jun-2015 | 17 | 1744 | 0 | 449.57 | 13487 |
| Jul-2015 | 23 | 1758 | 0 | 267.19 | 8283 |
| Aug-2015 | 19 | 2595 | 0 | 376.71 | 11678 |
| Sep-2015 | 15 | 1679 | 0 | 454.97 | 13649 |
| Oct-2015 | 15 | 2251 | 0 | 460.35 | 14271 |
| Nov-2015 | 3 | 3186 | 0 | 2154.30 | 64629 |
| Dec-2015 | 0 | 3350 | 386 | 2693.29 | 83492 |
| TOTAL | 108 | 3350 | 0 | 1143 | 415707 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in $M^{\wedge} 3$ | Minimum Daily <br> Flow in $\boldsymbol{M}^{\wedge 3}$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge 3}$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge} 3$ |
| Jan-2016 | 1 | 3240 | 0 | 1836.23 | 56923 |
| Feb-2016 | 1 | 3267 | 0 | 2639.38 | 76542 |
| Mar-2016 | 1 | 3099 | 0 | 1818.87 | 56385 |
| Apr-2016 | 8 | 3096 | 0 | 784.40 | 23532 |
| May-2016 | 16 | 3098 | 0 | 540.00 | 16740 |
| Jun-2016 | 10 | 1251 | 0 | 454.73 | 13642 |
| Jul-2016 | 16 | 1222 | 0 | 340.00 | 10540 |
| Aug-2016 | 16 | 1553 | 0 | 330.42 | 10243 |
| Sep-2016 | 19 | 1707 | 0 | 324.00 | 9720 |
| Oct-2016 | 7 | 2610 | 0 | 1328.81 | 41193 |
| Nov-2016 | 0 | 3228 | 603 | 2038.40 | 61152 |
| Dec-2016 | 0 | 3378 | 774 | 2344.35 | 72675 |
| TOTAL | 95 | 3378 | 0 | 1232 | 449287 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in $\boldsymbol{M}^{\wedge} 3$ | Minimum Daily <br> Flow in $M^{\wedge} 3$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge} 3$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge 3}$ |
| Jan-2017 | 1 | 2436 | 0 | 1390.29 | 43099 |
| Feb-2017 | 2 | 3234 | 0 | 2049.18 | 57377 |
| Mar-2017 | 0 | 3192 | 384 | 2050.16 | 63555 |
| Apr-2017 | 0 | 2328 | 111 | 1547.40 | 46422 |
| May-2017 | 11 | 3312 | 0 | 758.90 | 23526 |
| Jun-2017 | 21 | 1410 | 0 | 291.60 | 8748 |
| Jul-2017 | 7 | 882 | 0 | 356.03 | 11037 |
| Aug-2017 | 15 | 2790 | 0 | 494.32 | 15324 |
| Sep-2017 | 17 | 3375 | 0 | 494.00 | 14820 |
| Oct-2017 | 23 | 2028 | 0 | 290.32 | 9000 |
| Nov-2017 | 3 | 3324 | 0 | 1850.90 | 55527 |
| Dec-2017 | 5 | 3276 | 0 | 1763.90 | 54681 |
| TOTAL | 105 | 3375 | 0 | 1111 | 403116 |

## Appendix D. Leachate Pipeline Flow Data

| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 $^{\prime}$ | Minimum Daily <br> Flow in $M^{\wedge 3}$ | Average Daily <br> Flow in $M^{\wedge 3}$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge 3}$ |
| Jan-2018 | 2 | 3213 | 0 | 2401.16 | 74436 |
| Feb-2018 | 1 | 3438 | 0 | 2097.71 | 58736 |
| Mar-2018 | 9 | 27775 | 0 | 1048.97 | 32518 |
| Apr-2018 | 11 | 3111 | 0 | 1023.00 | 30690 |
| May-2018 | 17 | 2892 | 0 | 629.00 | 19497 |
| Jun-2018 | 17 | 2184 | 0 | 482.00 | 14473 |
| Jul-2018 | 11 | 1608 | 0 | 385.00 | 11942 |
| Aug-2018 | 21 | 1593 | 0 | 158.00 | 4908 |
| Sep-2018 | 18 | 1286 | 0 | 234.00 | 7026 |
| Oct-2018 | 8 | 1692 | 0 | 520.00 | 16127 |
| Nov-2018 | 1 | 2676 | 0 | 1157.00 | 35865 |
| Dec-2018 | 1 | 3234 | 0 | 2428.00 | 75255 |
| TOTAL | 117 | 27775 | 0 | 1047 | 381473 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in $\mathbf{M} \wedge^{\wedge}$ | Minimum Daily <br> Flow in $M^{\wedge 3}$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge 3}$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge} 3$ |
| Jan-2019 | 2 | 3300 | 0 | 2607.00 | 80823 |
| Feb-2019 | 3 | 2787 | 0 | 1302.00 | 36442 |
| Mar-2019 | 5 | 2439 | 0 | 1013.00 | 31393 |
| Apr-2019 | 8 | 2526 | 0 | 769.00 | 23078 |
| May-2019 | 11 | 2172 | 0 | 595.00 | 18437 |
| Jun-2019 | 21 | 802 | 0 | 152.00 | 4546 |
| Jul-2019 | 29 | 14 | 0 | 1.00 | 19 |
| Aug-2019 | 25 | 3320 | 0 | 158.00 | 4888 |
| Sep-2019 | 12 | 3298 | 0 | 858.00 | 25725 |
| Oct-2019 | 14 | 3010 | 0 | 897.00 | 27803 |
| Nov-2019 | 7 | 2114 | 0 | 785.00 | 23560 |
| Dec-2019 | 2 | 2196 | 0 | 1196.00 | 37091 |
| TOTAL | 139 | 3320 | 0 | 861 | 313805 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in $\mathbf{M}^{\wedge}$ | Minimum Daily <br> Flow in $\mathbf{M}^{\wedge}$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge} 3$ | Monthly Total <br> Flow in $\mathbf{M}^{\wedge} 3$ |
| Jan-2020 | 0 | 3108 | 1092 | 2793.00 | 86573 |
| Feb-2020 | 0 | 2990 | 2608 | 2910.00 | 84399 |
| Mar-2020 | 3 | 2892 | 0 | 1117.00 | 34616 |
| Apr-2020 | 6 | 2772 | 0 | 711.00 | 21319 |
| May-2020 | 13 | 1342 | 0 | 483.00 | 14962 |
| Jun-2020 | 26 | 1502 | 0 | 126.00 | 3791 |
| Jul-2020 | 21 | 692 | 0 | 68.00 | 2107 |
| Aug-2020 | 15 | 2840 | 0 | 356.00 | 11031 |
| Sep-2020 | 17 | 4804 | 0 | 710.00 | 21298 |
| Oct-2020 | 4 | 1956 | 0 | 976.00 | 30254 |
| Nov-2020 | 4 | 2474 | 0 | 1749.00 | 52828 |
| Dec-2020 | 0 | 3396 | 274 | 1767.00 | 54774 |
| TOTAL | 109 | 4804 | 0 | 1147 | 417952 |

## Appendix D. Leachate Pipeline Flow Data

| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 $^{\prime}$ | Minimum Daily <br> Flow in $M^{\wedge 3}$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge 3}$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge 3}$ |
| Jan-2021 | 0 | 6558 | 2122 | 3284.00 | 101810 |
| Feb-2021 | 15 | 18835 | 0 | 1722.00 | 48208 |
| Mar-2021 | 20 | 12720 | 0 | 1444.00 | 44771 |
| Apr-2021 | 29 | 3446 | 0 | 114.00 | 3446 |
| May-2021 | 25 | 32661 | 0 | 1297.00 | 40133 |
| Jun-2021 | 30 | 0 | 0 | 0.00 | 0 |
| Jul-2021 | 21 | 17081 | 0 | 1158.00 | 37060 |
| Aug-2021 | 25 | 2067 | 0 | 318.00 | 9846 |
| Sep-2021 | 7 | 6273 | 0 | 954.00 | 28633 |
| Oct-2021 | 9 | 3532 | 0 | 894.00 | 27700 |
| Nov-2021 | 0 | 6133 | 1496 | 3850.00 | 115487 |
| Dec-2021 | 0 | 5258 | 1377 | 3079.00 | 95452 |
| TOTAL | 181 | 32661 | 0 | 1510 | 552546 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in $M^{\wedge} 3$ | Minimum Daily <br> Flow in $\mathbf{M}^{\wedge} 3$ | Average Daily <br> Flow in $\boldsymbol{M}^{\wedge} 3$ | Monthly Total <br> Flow in $\boldsymbol{M}^{\wedge} \mathbf{3}$ |
| Jan-2022 | 0 | 6256 | 1303 | 3202.00 | 99263 |
| Feb-2022 | 2 | 4019 | 0 | 1443.00 | 40395 |
| Mar-2022 | 3 | 6506 | 0 | 2118.00 | 65646 |
| Apr-2022 | 4 | 3539 | 0 | 1880.00 | 56406 |
| May-2022 | 10 | 6914 | 0 | 1538.00 | 47688 |
| Jun-2022 | 20 | 5661 | 0 | 854.00 | 25631 |
| Jul-2022 | 14 | 5058 | 0 | 1293.00 | 41368 |
| Aug-2022 | 8 | 4974 | 0 | 716.00 | 22212 |
| Sep-2022 | 0 | 6619 | 3 | 1266.00 | 38006 |
| Oct-2022 | 1 | 2454 | 0 | 578.00 | 17933 |
| Nov-2022 | 5 | 4306 | 0 | 1052.00 | 31578 |
| Dec-2022 | 2 | 6489 | 0 | 1588.00 | 49249 |
| TOTAL | 69 | 6914 | 0 | 1461 | 535375 |


| Hartland Leachate System Monthly Flow Data |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Days Without <br> Flow | Maximum Daily <br> Flow in M^3 | Minimum Daily <br> Flow in M^3 | Average Daily <br> Flow in $M^{\wedge 3}$ | Monthly Total <br> Flow in $M^{\wedge 3}$ |
| Jan-2023 | 0 | 7783 | 538 | 2912.00 | 90297 |
| Feb-2023 | 0 | 4881 | 503 | 1762.00 | 49340 |
| Mar-2023 | 5 | 5463 | 0 | 1701.00 | 51032 |
| Apr-2023 |  |  |  |  |  |
| May-2023 |  |  |  |  |  |
| Jun-2023 |  |  |  |  |  |
| Jul-2023 |  |  |  |  |  |
| Aug-2023 |  |  |  |  |  |
| Sep-2023 |  |  |  |  |  |
| Oct-2023 |  |  |  |  |  |
| Nov-2023 |  |  |  |  |  |
| Dec-2023 |  |  |  |  |  |
| TOTAL | 5 |  |  |  |  |

## Appendix E

## Hartland Landfill Site Plan and Sampling Locations

- E1. Hartland Landfill Site Plan
- E2. Groundwater Level Monitoring Locations
- E3. Groundwater Quality Monitoring Locations
- E4. Surface Water Quality Monitoring Locations
- E5. Leachate Quality Monitoring Locations

E1. Hartland Landfill Site Plan

## H A R T L A N D L A N D F I L L S ITEPLAN

## Mount Work <br> Regional Park



Mount Work
Regional Park

| СГ】】 |  |
| :---: | :---: |
| mimamamam |  |


| - Observation Chamber | Contours (5m interval) | Cell Phases |
| :---: | :---: | :---: |
| [Ps Pump Station | Index | Lot Lines |
| Residual Solids \& Centrate Return Line Hartland Leachate Line | Index-Depression | Regional Park |

E2. Groundwater Level
Monitoring Locations

E3. Groundwater Quality Monitoring Locations


E4. Surface Water Quality Monitoring Locations

E5. Leachate Quality Monitoring Locations

## LEACHATE QUALITY MONITORING LOCATIONS



## Appendix F

Hartland Landfill Leachate Pipeline Plan

## HARTLAND LANDFILL - LEACHATE PIPELINE PLAN



## Appendix G

## Results of Statistical Analysis

- G1. Groundwater
- G2. Surface Water
- G3. Leachate


## G1. Groundwater

Appendix G-1. Results of Statistical Analysis - Groundwater

|  |  | Parameter |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Station | Time Period | Conductivity |  | Ammonia |  | Chloride |  | Sulphate |  | Nitrate |  |
|  |  | Increasing | Decreasing | Increasing | Decreasing | Increasing | Decreasing | Increasing | Decreasing | Increasing | Decreasing |
|  |  | Y/N | Y/N | Y/N | Y/N | Y/N | Y/N | Y/N | Y/N | Y/N | Y/N |
| Boundary Compliance Monitoring Wells |  |  |  |  |  |  |  |  |  |  |  |
| Gw-04-3-1 | 2018-2023 | Y | N | N | N | Y | N | Y | N | N | Y |
| Gw-04-4-1 | 2018-2023 | N | N | N | N | Y | N | N | N | N | N |
| Gw-17-1-1 | 2018-2023 | N | N | N | N | N | Y | N | N | N | N |
| Gw-17-1-2 | 2018-2023 | N | N | N | N | N | Y | N | N | N | N |
| Gw-17-1-3 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| Gw-18-1-1 | 2018-2023 | N | N | N | N | N | N | Y | N | N | N |
| Gw-18-2-1 | 2018-2023 | Y | N | N | N | Y | N | Y | N | Y | N |
| Gw-18-2-2 | 2018-2023 | N | N | N | Y | N | N | Y | N | N | N |
| Gw-20-1-1 | 2018-2023 | N | Y | N | N | N | Y | N | N | N | Y |
| Gw-20-1-2 | 2018-2023 | N | N | N | Y | N | Y | N | N | N | N |
| Gw-21-1-1 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| Gw-21-1-2 | 2018-2023 | N | N | N | N | N | N | Y | N | N | N |
| Gw-21-2-1 | 2018-2023 | Y | N | N | Y | N | N | Y | N | N | N |
| Gw-28-1-0 | 2018-2023 | Y | N | N | N | N | N | N | N | Y | N |
| Gw-29-1-1 | 2018-2023 | N | N | N | N | N | N | Y | N | Y | N |
| Gw-29-1-2 | 2018-2023 | Y | N | N | N | N | N | Y | N | Y | N |
| Gw-30-1-1 | 2018-2023 | Y | N | N | N | N | Y | N | N | N | Y |
| Gw-30-1-2 | 2018-2023 | N | N | N | N | N | N | Y | N | N | N |
| Gw-31-1-1 | 2018-2023 | Y | N | N | N | N | Y | Y | N | N | N |
| Gw-31-1-2 | 2018-2023 | Y | N | N | N | N | N | Y | N | N | N |
| Gw-39-1-1 | 2018-2023 | N | N | N | Y | N | N | Y | N | Y | N |
| Gw-39-2-1 | 2018-2023 | N | N | N | N | N | N | Y | N | Y | N |
| Gw-41-1-1 | 2018-2023 | N | N | N | N | Y | N | N | Y | N | N |
| Gw-42-1-1 | 2018-2023 | Y | N | N | N | Y | N | Y | N | N | N |
| Gw-53-1-1 | 2018-2023 | Y | N | N | N | N | Y | Y | N | N | N |
| Gw-55-1-1 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| Gw-56-1-1 | 2018-2023 | Y | N | N | N | N | N | N | N | N | N |
| Gw-57-1-1 | 2018-2023 | Y | N | N | N | N | N | N | N | N | N |
| Gw-71-1-1 | 2018-2023 | Y | N | N | N | N | N | Y | N | N | N |
| Gw-71-2-1 | 2018-2023 | Y | N | N | N | N | N | Y | N | N | N |
| Gw-71-3-1 | 2018-2023 | Y | N | N | N | N | N | Y | N | Y | N |
| Gw-72-1-1 | 2018-2023 | Y | N | N | Y | N | N | N | N | N | N |
| Gw-72-3-1 | 2018-2023 | N | Y | N | N | N | Y | N | N | N | N |
| Gw-73-1-1 | 2018-2023 | Y | N | N | N | N | N | Y | N | Y | N |
| Gw-73-2-1 | 2018-2023 | Y | N | N | Y | N | N | N | N | Y | N |
| Gw-73-3-1 | 2018-2023 | N | N | N | N | N | N | N | N | Y | N |

Notes
NA indicates the analysis is unable to run due to insufficient dataset or over $80 \%$ of values were under detection limits
N-No Trend
Green highlights indicate decreasing trends

Appendix G-1. Results of Statistical Analysis - Groundwater

| Additional Monitorin |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gw-07-1-0 | 2018-2023 | Y | N | N | N | Y | N | N | N | N | Y |
| Gw-16-1-1 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| Gw-16-1-2 | 2018-2023 | N | N | N | N | Y | N | N | N | N | N |
| Gw-16-2-1 | 2018-2023 | N | N | N | N | Y | N | N | Y | N | N |
| Gw-16-2-2 | 2018-2023 | Y | N | N | N | Y | N | N | N | N | N |
| Gw-19-1-1 | 2018-2023 | Y | N | Y | N | N | N | N | N | N | N |
| Gw-19-1-2 | 2018-2023 | N | N | N | Y | N | N | N | N | N | N |
| Gw-19-2-1 | 2018-2023 | Y | N | N | N | Y | N | N | N | N | N |
| Gw-19-2-2 | 2018-2023 | N | N | N | N | Y | N | N | Y | N | Y |
| Gw-25-1-1 | 2018-2023 | Y | N | N | Y | N | Y | Y | N | Y | N |
| Gw-25-1-2 | 2018-2023 | Y | N | N | N | N | N | Y | N | N | N |
| Gw-27-1-1 | 2018-2023 | Y | N | N | N | N | N | Y | N | N | N |
| Gw-27-1-2 | 2018-2023 | Y | N | N | N | N | N | Y | N | N | N |
| Gw-36-2-1 | 2018-2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-36-3-1 | 2018-2023 | Y | N | N | N | Y | N | N | N | N | Y |
| Gw-37-2-1 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| Gw-37-3-1 | 2018-2023 | Y | N | N | N | Y | N | N | N | N | N |
| Gw-38-1-1 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| Gw-40-1-1 | 2018-2023 | N | N | N | N | N | Y | N | N | N | N |
| Gw-43-1-1 | 2018-2023 | Y | N | N | N | N | Y | Y | N | N | N |
| Gw-44-1-1 | 2018-2023 | N | N | N | N | N | N | N | N | Y | N |
| Gw-51-1-1 | 2018-2023 | N | N | N | Y | N | Y | N | Y | N | N |
| Gw-51-2-1 | 2018-2023 | N | N | N | Y | N | Y | Y | N | N | N |
| Gw-51-3-1 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| Gw-52-1-1 | 2018-2023 | N | N | N | Y | N | N | N | N | N | N |
| Gw-52-4-0 (P7) | 2018-2023 | Y | N | N | N | N | N | N | N | N | N |
| Gw-58-1-0 | 2018-2023 | N | N | N | N | N | N | Y | N | Y | N |
| Gw-60-1-1 | 2018-2023 | Y | N | N | Y | Y | N | N | Y | N | N |
| Gw-60-2-1 | 2018-2023 | Y | N | N | N | Y | N | N | N | N | Y |
| Gw-60-3-1 | 2018-2023 | Y | N | N | N | Y | N | Y | N | N | N |
| Gw-62-1-1 | 2018-2023 | N | Y | N | N | N | Y | N | N | Y | N |
| Gw-62-2-1 | 2018-2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-63-1-1 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| Gw-63-2-1 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| Gw-77-1-1 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| Gw-77-2-1 | 2018-2023 | N | N | N | N | N | N | Y | N | N | N |
| Gw-78-1-1 | 2018-2023 | N | N | N | N | N | N | Y | N | N | N |
| Gw-78-2-1 | 2018-2023 | N | N | N | Y | N | Y | N | N | N | N |
| Gw-85-1-1 | 2018-2023 | N | N | N | Y | N | N | N | N | N | N |
| Gw-87-1-1 | 2018-2023 | N | N | N | N | N | N | Y | N | N | N |
| Gw-87-2-1 | 2018-2023 | Y | N | N | N | N | N | N | N | Y | N |
| Gw-88-1-1 | 2018-2023 | Y | N | N | N | N | Y | Y | N | Y | N |
| Gw-88-2-1 | 2018-2023 | N | N | N | Y | N | Y | N | N | N | N |
| Gw-91-1-1 | 2018-2023 | Y | N | N | N | N | N | Y | N | N | N |
| Gw-92-1-1 | 2018-2023 | Y | N | N | N | N | Y | Y | N | Y | N |
| Gw-93-1-1 | 2018-2023 | N | N | N | N | N | N | Y | N | N | N |
| Gw-94-1-1 | 2018-2023 | N | N | N | N | N | Y | N | Y | N | N |
| Gw-95-1-1 | 2022 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-96-1-1 | 2022 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-97-1-1 | 2022 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-98-1-1 | 2022-2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-100-1-1 | 2022-2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-103-1-1 | 2022 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-104-1-1 | 2022-2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-105-1-1 | 2022-2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-106-1-1 | 2022-2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-107-1-1 | 2022 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-108-1-1 | 2022 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-109-1-1 | 2022 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-110-1-1 | 2022 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-P1 | 2018-2023 | Y | N | N | N | Y | N | N | N | N | N |
| Gw-P2 | 2018-2023 | Y | N | N | N | N | N | N | N | N | N |
| Gw-P3 | 2018-2023 | N | N | N | Y | N | N | N | N | N | N |
| Gw-P4 | 2018-2023 | Y | N | N | Y | N | Y | N | N | N | N |
| Gw-80-1-0 (P8) | 2018-2023 | Y | N | Y | N | Y | N | N | N | N | Y |
| Gw-81-1-0 (P9) | 2018-2023 | N | N | N | N | N | Y | N | N | N | Y |
| Gw-P10 | 2018-2023 | Y | N | N | N | Y | N | N | N | N | N |
| Gw-P11 | 2022-2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Gw-P12 | 2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |

NA indicates the analysis is unable to run due to insufficient dataset or over $80 \%$ of values were under detection limits
N - No Trend

[^3]
## G2. Surface Water

| Station | Time Period | Parameter |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Conductivity |  | Ammonia |  | Chloride |  | Sulphate |  | Nitrate |  |
|  |  | Increasing | Decreasing | Increasing | Decreasing | Increasing | Decreasing | Increasing | Decreasing | Increasing | Decreasing |
|  |  | Y/N | Y/N | Y/N | Y/N | Y/N | Y/N | Y/N | Y/N | Y/N | Y/N |
| Boundary Surface Water Quality Location |  |  |  |  |  |  |  |  |  |  |  |
| SW-S-04 | 2018-2023 | N | N | N | N | Y | N | N | N | N | N |
| SW-N-05 | 2018-2023 | Y | N | N | N | Y | N | Y | N | Y | N |
| SW-N-16 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| SW-N-41s1 | 2018-2023 | N | N | N | Y | Y | N | N | Y | N | N |
| SW-N-42s1 | 2018-2023 | Y | N | N | N | Y | N | N | N | N | N |
| Routine Surface Water Quality Location |  |  |  |  |  |  |  |  |  |  |  |
| SW-S-03 | 2018-2023 | N | N | N | N | Y | N | N | N | N | N |
| SW-N-CSs2 | 2018-2023 | Y | N | N | N | N | Y | N | N | N | N |
| SW-S-12 | 2018-2023 | N | N | N | N | Y | N | N | N | N | N |
| SW-N-14 | 2018-2023 | N | N | N | N | Y | N | N | N | N | N |
| SW-N-15 | 2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-17 | 2018-2023 | Y | N | N | N | Y | N | N | N | Y | N |
| SW-N-18 | 2018-2023 | Y | N | Y | N | Y | N | N | N | Y | N |
| SW-N-19 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| SW-S-20 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| SW-S-21 | 2018-2023 | Y | N | N | N | N | N | N | N | Y | N |
| SW-S-24 | 2018-2023 | N | N | N | N | Y | N | N | N | N | N |
| SW-S-27 | 2018-2023 | N | N | N | N | N | N | N | N | N | Y |
| SW-S-52 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| SW-N-41s3 | 2018-2023 | N | N | N | N | N | N | N | Y | N | N |
| SW-N-41s4 | 2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-45 | 2017-2022 | Y | N | N | N | Y | N | N | N | Y | N |
| SW-N-50 | 2018-2021 | N | N | N | N | N | N | N | N | N | N |
| SW-N-51 | 2018, 2020 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-53 | 2018, 2021, 2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-54 | 2018-2023 | N | N | N | N | N | N | N | N | N | N |
| SW-N-58 | 2023 (dry) | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-59 | 2023 (dry) | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-60 | 2023 (dry) | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-61 | 2023 (dry) | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-62 | 2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-63 | 2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-64 | 2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| SW-N-65 | 2023 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |

Notes:
NA indicates the analysis is unable to run due to insufficient dataset or over $80 \%$ of values were under detection limits
Green highlights indicate decreasing trends
nighlights indicate decreasing tren

## G3. Leachate

## AECOM

Appendix G-3. Results of Statistical Analysis - Leachate

| Station | Time Period | Parameter |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Conductivity |  | Ammonia |  | Chloride |  | Sulphate |  | Nitrate |  |
|  |  | Increasing Y/N | Decreasing Y/N | Increasing Y/N | Decreasing Y/N | Increasing Y/N | Decreasing Y/N | Increasing Y/N | Decreasing Y/N | Increasing Y/N | Decreasing Y/N |
| Hartland Valve Chamber | 2018-2023 | N | N | N | Y | N | N | N | N | Y | N |

Notes:
Green highlights indicate decreasing trends
aecom.com


[^0]:    frn

[^1]:    Footvalves are $16 \mathrm{~mm}(5 / 5)^{\circ}$ unless ol
    $N A$ - Not Available or Not Applicable.

[^2]:    
    

[^3]:    Green highlights indicate decreasing trends

